

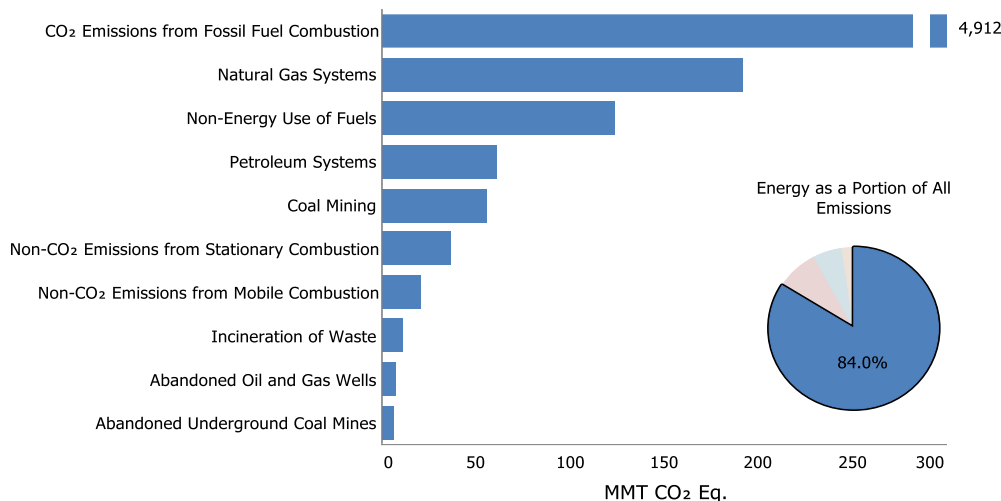
### 3. Energy

Energy-related activities were the primary sources of U.S. anthropogenic greenhouse gas emissions, accounting for 84.0 percent of total greenhouse gas emissions on a carbon dioxide (CO<sub>2</sub>) equivalent basis in 2017.<sup>1</sup> This included 97, 43, and 13 percent of the nation’s CO<sub>2</sub>, methane (CH<sub>4</sub>), and nitrous oxide (N<sub>2</sub>O) emissions, respectively. Energy-related CO<sub>2</sub> emissions alone constituted 81.6 percent of national emissions from all sources on a CO<sub>2</sub> equivalent basis, while the non-CO<sub>2</sub> emissions from energy-related activities represented a much smaller portion of total national emissions (5.1 percent collectively).

Emissions from fossil fuel combustion comprise the vast majority of energy-related emissions, with CO<sub>2</sub> being the primary gas emitted (see Figure 3-1). Globally, approximately 32,310 million metric tons (MMT) of CO<sub>2</sub> were added to the atmosphere through the combustion of fossil fuels in 2016, of which the United States accounted for approximately 15 percent.<sup>2</sup> Due to their relative importance, fossil fuel combustion-related CO<sub>2</sub> emissions are considered separately and in more detail than other energy-related emissions (see Figure 3-2).

Fossil fuel combustion also emits CH<sub>4</sub> and N<sub>2</sub>O. Stationary combustion of fossil fuels was the second largest source of N<sub>2</sub>O emissions in the United States and mobile fossil fuel combustion was the fourth largest source. Energy-related activities other than fuel combustion, such as the production, transmission, storage, and distribution of fossil fuels, also emit greenhouse gases. These emissions consist primarily of fugitive CH<sub>4</sub> from natural gas systems, coal mining, and petroleum systems.

**Figure 3-1: 2017 Energy Chapter Greenhouse Gas Sources (MMT CO<sub>2</sub> Eq.)**



<sup>1</sup> Estimates are presented in units of million metric tons of carbon dioxide equivalent (MMT CO<sub>2</sub> Eq.), which weight each gas by its global warming potential, or GWP, value. See section on global warming potentials in the Executive Summary.

<sup>2</sup> Global CO<sub>2</sub> emissions from fossil fuel combustion were taken from International Energy Agency *CO<sub>2</sub> Emissions from Fossil Fuels Combustion Overview* <<https://webstore.iea.org/co2-emissions-from-fuel-combustion-2018>> IEA (2018).

**Figure 3-2: 2017 U.S. Fossil Carbon Flows (MMT CO<sub>2</sub> Eq.)**

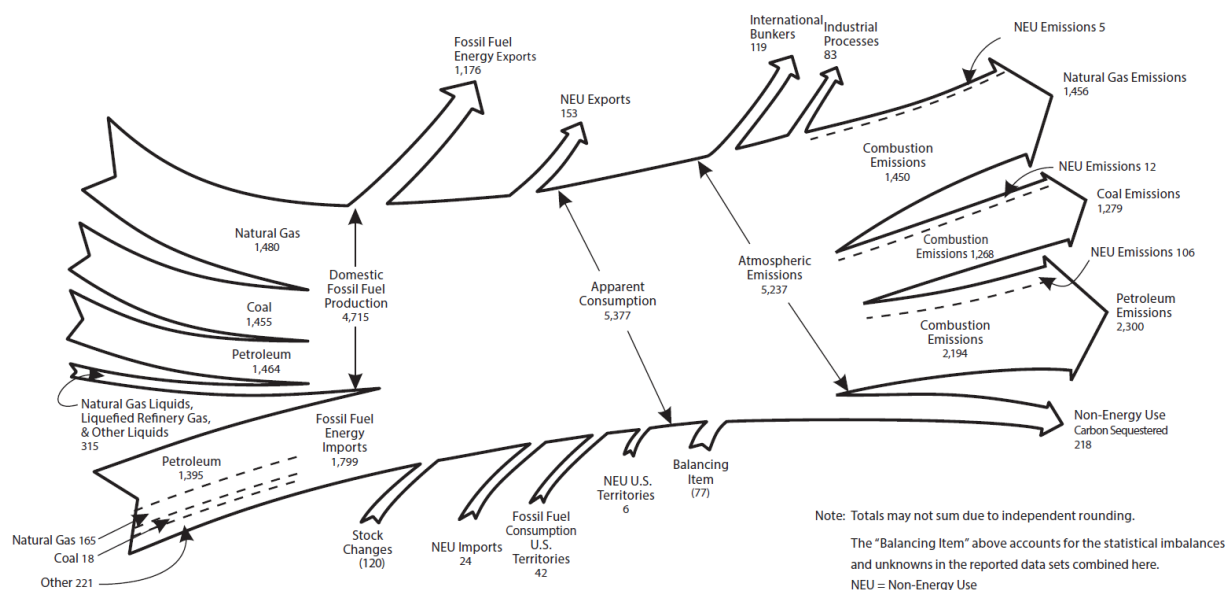


Table 3-1 summarizes emissions from the Energy sector in units of MMT CO<sub>2</sub> Eq., while unweighted gas emissions in kilotons (kt) are provided in Table 3-2. Overall, emissions due to energy-related activities were 5,424.8 MMT CO<sub>2</sub> Eq. in 2017,<sup>3</sup> an increase of 1.6 percent since 1990 and a decrease of 0.7 percent since 2016.

**Table 3-1: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from Energy (MMT CO<sub>2</sub> Eq.)**

Gas/Source	1990	2005	2013	2014	2015	2016	2017
<b>CO<sub>2</sub></b>	<b>4,905.3</b>	<b>5,931.0</b>	<b>5,341.5</b>	<b>5,384.8</b>	<b>5,241.5</b>	<b>5,134.1</b>	<b>5,095.6</b>
Fossil Fuel Combustion	4,738.8	5,744.8	5,157.4	5,199.3	5,047.1	4,961.9	4,912.0
<i>Transportation</i>	1,469.1	1,857.0	1,682.7	1,721.6	1,734.0	1,779.0	1,800.6
<i>Electric Power</i>	1,820.0	2,400.0	2,038.3	2,037.1	1,900.6	1,808.9	1,732.0
<i>Industrial</i>	857.5	853.4	840.0	819.6	807.9	807.6	810.7
<i>Residential</i>	338.2	357.9	329.3	346.8	317.8	292.9	294.5
<i>Commercial</i>	226.5	226.8	224.6	232.9	245.5	232.1	232.9
<i>U.S. Territories</i>	27.6	49.7	42.5	41.4	41.4	41.4	41.4
Non-Energy Use of Fuels	119.6	139.6	123.5	119.9	126.9	113.7	123.2
Natural Gas Systems	30.0	22.6	25.1	25.5	25.1	25.5	26.3
Petroleum Systems	9.0	11.6	25.1	29.6	31.7	22.2	23.3
Incineration of Waste	8.0	12.5	10.3	10.4	10.7	10.8	10.8
Abandoned Oil and Gas Wells	+	+	+	+	+	+	+
Biomass-Wood <sup>a</sup>	215.2	206.9	227.3	233.8	224.7	216.3	221.4
International Bunker Fuels <sup>b</sup>	103.5	113.1	99.8	103.4	110.9	116.6	120.1
Biofuels-Ethanol <sup>a</sup>	4.2	22.9	74.7	76.1	78.9	81.2	82.1
Biofuels-Biodiesel <sup>a</sup>	0.0	0.9	13.5	13.3	14.1	19.6	18.7
<b>CH<sub>4</sub></b>	<b>366.9</b>	<b>303.2</b>	<b>298.4</b>	<b>298.1</b>	<b>293.5</b>	<b>283.0</b>	<b>283.3</b>
Natural Gas Systems	193.1	171.4	165.6	165.1	167.2	165.7	165.6
Coal Mining	96.5	64.1	64.6	64.6	61.2	53.8	55.7
Petroleum Systems	42.1	36.7	41.6	42.1	39.5	38.2	37.7
Stationary Combustion	8.6	7.8	8.7	8.9	8.5	7.9	7.8

<sup>3</sup> Following the current reporting requirements under the UNFCCC, this Inventory report presents CO<sub>2</sub> equivalent values based on the IPCC Fourth Assessment Report (AR4) GWP values. See the Introduction chapter for more information.

Abandoned Oil and Gas Wells	6.6	6.9	7.0	7.1	7.1	7.2	6.9
Abandoned Underground Coal Mines	7.2	6.6	6.2	6.3	6.4	6.7	6.4
Mobile Combustion	12.9	9.6	4.5	4.1	3.6	3.4	3.2
Incineration of Waste	+	+	+	+	+	+	+
<i>International Bunker Fuels<sup>b</sup></i>	<i>0.2</i>	<i>0.1</i>	<i>0.1</i>	<i>0.1</i>	<i>0.1</i>	<i>0.1</i>	<i>0.1</i>
<b>N<sub>2</sub>O</b>	<b>67.6</b>	<b>73.7</b>	<b>55.2</b>	<b>53.5</b>	<b>49.7</b>	<b>48.3</b>	<b>45.9</b>
Stationary Combustion	25.1	34.3	32.7	33.0	30.6	30.1	28.6
Mobile Combustion	42.0	39.0	22.1	20.2	18.8	17.9	16.9
Incineration of Waste	0.5	0.4	0.3	0.3	0.3	0.3	0.3
Petroleum Systems	+	+	+	+	+	+	+
Natural Gas Systems	+	+	+	+	+	+	+
<i>International Bunker Fuels<sup>b</sup></i>	<i>0.9</i>	<i>1.0</i>	<i>0.9</i>	<i>0.9</i>	<i>0.9</i>	<i>1.0</i>	<i>1.0</i>
<b>Total</b>	<b>5,339.8</b>	<b>6,308.0</b>	<b>5,695.0</b>	<b>5,736.4</b>	<b>5,584.7</b>	<b>5,465.3</b>	<b>5,424.8</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub> Eq.

<sup>a</sup> Emissions from Wood Biomass, Ethanol, and Biodiesel Consumption are not included specifically in summing Energy sector totals. Net carbon fluxes from changes in biogenic carbon reservoirs are accounted for in the estimates for LULUCF.

<sup>b</sup> Emissions from International Bunker Fuels are not included in totals. These values are presented for informational purposes only, in line with the 2006 IPCC Guidelines and UNFCCC reporting obligations.

Note: Totals may not sum due to independent rounding.

**Table 3-2: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from Energy (kt)**

Gas/Source	1990	2005	2013	2014	2015	2016	2017
<b>CO<sub>2</sub></b>	<b>4,905,262</b>	<b>5,931,045</b>	<b>5,341,485</b>	<b>5,384,792</b>	<b>5,241,538</b>	<b>5,134,056</b>	<b>5,095,643</b>
Fossil Fuel Combustion	4,738,756	5,744,754	5,157,391	5,199,345	5,047,107	4,961,876	4,911,962
Non-Energy Use of Fuels	119,551	139,625	123,476	119,895	126,939	113,719	123,221
Natural Gas Systems	30,048	22,638	25,148	25,518	25,071	25,488	26,327
Petroleum Systems	8,950	11,552	25,130	29,597	31,672	22,200	23,336
Incineration of Waste	7,950	12,469	10,333	10,429	10,742	10,765	10,790
Abandoned Oil and Gas Wells	6	7	7	7	7	7	7
<i>Biomass-Wood<sup>a</sup></i>	<i>215,186</i>	<i>206,901</i>	<i>227,340</i>	<i>233,762</i>	<i>224,730</i>	<i>216,293</i>	<i>221,432</i>
<i>International Bunker Fuels<sup>b</sup></i>	<i>103,463</i>	<i>113,139</i>	<i>99,763</i>	<i>103,400</i>	<i>110,887</i>	<i>116,594</i>	<i>120,107</i>
<i>Biofuels-Ethanol<sup>a</sup></i>	<i>4,227</i>	<i>22,943</i>	<i>74,743</i>	<i>76,075</i>	<i>78,934</i>	<i>81,250</i>	<i>82,088</i>
<i>Biofuels-Biodiesel<sup>a</sup></i>	<i>0</i>	<i>856</i>	<i>13,462</i>	<i>13,349</i>	<i>14,077</i>	<i>19,648</i>	<i>18,705</i>
<b>CH<sub>4</sub></b>	<b>14,677</b>	<b>12,127</b>	<b>11,935</b>	<b>11,922</b>	<b>11,738</b>	<b>11,320</b>	<b>11,332</b>
Natural Gas Systems	7,723	6,856	6,624	6,603	6,686	6,629	6,624
Coal Mining	3,860	2,565	2,584	2,583	2,449	2,154	2,227
Petroleum Systems	1,682	1,469	1,665	1,682	1,579	1,528	1,506
Stationary Combustion	344	313	350	355	340	318	312
Abandoned Oil and Gas Wells	262	277	282	283	285	289	277
Abandoned Underground Coal Mines	288	264	249	253	256	268	257
Mobile Combustion	518	384	181	163	143	135	128
Incineration of Waste	+	+	+	+	+	+	+
<i>International Bunker Fuels<sup>b</sup></i>	<i>7</i>	<i>5</i>	<i>3</i>	<i>3</i>	<i>3</i>	<i>4</i>	<i>4</i>
<b>N<sub>2</sub>O</b>	<b>227</b>	<b>247</b>	<b>185</b>	<b>180</b>	<b>167</b>	<b>162</b>	<b>154</b>
Stationary Combustion	84	115	110	111	103	101	96
Mobile Combustion	141	131	74	68	63	60	57
Incineration of Waste	2	1	1	1	1	1	1
Petroleum Systems	+	+	+	+	+	+	+

Natural Gas Systems	+		+		+		+		+
International Bunker Fuels <sup>b</sup>	3		3		3		3		3

+ Does not exceed 0.5 kt.

<sup>a</sup> Emissions from Wood Biomass, Ethanol, and Biodiesel Consumption are not included specifically in summing Energy sector totals. Net carbon fluxes from changes in biogenic carbon reservoirs are accounted for in the estimates for LULUCF.

<sup>b</sup> Emissions from International Bunker Fuels are not included in totals. These values are presented for informational purposes only, in line with the 2006 IPCC Guidelines and UNFCCC reporting obligations.

Note: Totals may not sum due to independent rounding.

Each year, some emission and sink estimates in the Inventory are recalculated and revised with improved methods and/or data. In general, recalculations are made to the U.S. greenhouse gas emission estimates either to incorporate new methodologies or, most commonly, to update recent historical data. These improvements are implemented consistently across the previous Inventory’s time series (i.e., 1990 to 2016) to ensure that the trend is accurate. Updates to N<sub>2</sub>O emissions from Stationary Combustion in the Energy sector resulted in an average change over the time series of greater than 10 MMT CO<sub>2</sub> Eq. For more information on specific methodological updates, please see the Recalculations Discussion for each category, in this chapter.

### Box 3-1: Methodological Approach for Estimating and Reporting U.S. Emissions and Removals

In following the United Nations Framework Convention on Climate Change (UNFCCC) requirement under Article 4.1 to develop and submit national greenhouse gas emission inventories, the emissions and removals presented in this report and this chapter, are organized by source and sink categories and calculated using internationally-accepted methods provided by the Intergovernmental Panel on Climate Change (IPCC) in the 2006 IPCC Guidelines for National Greenhouse Gas Inventories (2006 IPCC Guidelines). Additionally, the calculated emissions and removals in a given year for the United States are presented in a common manner in line with the UNFCCC reporting guidelines for the reporting of inventories under this international agreement. The use of consistent methods to calculate emissions and removals by all nations providing their inventories to the UNFCCC ensures that these reports are comparable. The presentation of emissions and removals provided in this Inventory do not preclude alternative examinations, but rather, this Inventory presents emissions and removals in a common format consistent with how countries are to report Inventories under the UNFCCC. The report itself, and this chapter, follows this standardized format, and provides an explanation of the application of methods used to calculate emissions and removals.

### Box 3-2: Energy Data from EPA’s Greenhouse Gas Reporting Program

On October 30, 2009, the U.S. Environmental Protection Agency (EPA) published a rule requiring annual reporting of greenhouse gas data from large greenhouse gas emission sources in the United States. Implementation of the rule, codified at 40 CFR Part 98, is referred to as EPA’s Greenhouse Gas Reporting Program (GHGRP). The rule applies to direct greenhouse gas emitters, fossil fuel suppliers, industrial gas suppliers, and facilities that inject CO<sub>2</sub> underground for sequestration or other reasons and requires reporting by sources or suppliers in 41 industrial categories. Annual reporting is at the facility level, except for certain suppliers of fossil fuels and industrial greenhouse gases. Data reporting by affected facilities includes the reporting of emissions from fuel combustion at that affected facility. In general, the threshold for reporting is 25,000 metric tons or more of CO<sub>2</sub> Eq. per year.

EPA’s GHGRP dataset and the data presented in this Inventory are complementary. The GHGRP dataset continues to be an important resource for the Inventory, providing not only annual emissions information, but also other annual information, such as activity data and emission factors that can improve and refine national emission estimates and trends over time. GHGRP data also allow EPA to disaggregate national inventory estimates in new ways that can highlight differences across regions and sub-categories of emissions, along with enhancing application of QA/QC procedures and assessment of uncertainties.

EPA uses annual GHGRP data in a number of categories to improve the national estimates presented in this Inventory consistent with IPCC guidelines (see also Box 3-4).<sup>4</sup> As indicated in the respective Planned Improvements

<sup>4</sup> See <[http://www.ipcc-nggip.iges.or.jp/public/tb/TFI\\_Technical\\_Bulletin\\_1.pdf](http://www.ipcc-nggip.iges.or.jp/public/tb/TFI_Technical_Bulletin_1.pdf)>.

sections for source categories in this chapter, EPA continues to examine the uses of facility-level GHGRP data to improve the national estimates presented in this Inventory. Most methodologies used in EPA’s GHGRP are consistent with IPCC, though for EPA’s GHGRP, facilities collect detailed information specific to their operations according to detailed measurement standards, which may differ with the more aggregated data collected for the Inventory to estimate total national U.S. emissions. It should be noted that the definitions and provisions for reporting fuel types in EPA’s GHGRP may differ from those used in the Inventory in meeting the UNFCCC reporting guidelines. In line with the UNFCCC reporting guidelines, the Inventory report is a comprehensive accounting of all emissions from fuel types identified in the IPCC guidelines and provides a separate reporting of emissions from biomass. Further information on the reporting categories in EPA’s GHGRP and specific data caveats associated with monitoring methods in EPA’s GHGRP is provided on the GHGRP website.<sup>5</sup>

EPA presents the data collected by its GHGRP through a data publication tool that allows data to be viewed in several formats including maps, tables, charts and graphs for individual facilities or groups of facilities.<sup>6</sup>

In addition to using GHGRP data to estimate emissions, EPA also uses the GHGRP fuel consumption activity data in the Energy sector to disaggregate industrial end-use sector emissions in the category of CO<sub>2</sub> Emissions from Fossil Fuel Combustion, for use in reporting emissions in Common Reporting Format (CRF) tables. The industrial end-use sector activity data collected for the Inventory (EIA 2019) represent aggregated data for the industrial end-use sector. EPA’s GHGRP collects industrial fuel consumption activity data by individual categories within the industrial end-use sector. Therefore, the GHGRP data are used to provide a more detailed breakout of total emissions in the industrial end-use sector within that source category.

## 3.1 Fossil Fuel Combustion (CRF Source Category 1A)

Emissions from the combustion of fossil fuels for energy include the gases CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O. Given that CO<sub>2</sub> is the primary gas emitted from fossil fuel combustion and represents the largest share of U.S. total emissions, CO<sub>2</sub> emissions from fossil fuel combustion are discussed at the beginning of this section. Following that is a discussion of emissions of all three gases from fossil fuel combustion presented by sectoral breakdowns. Methodologies for estimating CO<sub>2</sub> from fossil fuel combustion also differ from the estimation of CH<sub>4</sub> and N<sub>2</sub>O emissions from stationary combustion and mobile combustion. Thus, three separate descriptions of methodologies, uncertainties, recalculations, and planned improvements are provided at the end of this section. Total CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O emissions from fossil fuel combustion are presented in Table 3-3 and Table 3-4.

**Table 3-3: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from Fossil Fuel Combustion (MMT CO<sub>2</sub> Eq.)**

Gas	1990	2005	2013	2014	2015	2016	2017
CO <sub>2</sub>	4,738.8	5,744.8	5,157.4	5,199.3	5,047.1	4,961.9	4,912.0
CH <sub>4</sub>	21.5	17.4	13.3	13.0	12.1	11.3	11.0
N <sub>2</sub> O	67.1	73.3	54.8	53.2	49.4	47.9	45.5
<b>Total</b>	<b>4,827.4</b>	<b>5,835.5</b>	<b>5,225.5</b>	<b>5,265.5</b>	<b>5,108.6</b>	<b>5,021.1</b>	<b>4,968.5</b>

Note: Totals may not sum due to independent rounding.

<sup>5</sup> See

<<http://www.ccdsupport.com/confluence/display/ghgp/Detailed+Description+of+Data+for+Certain+Sources+and+Processes>>.

<sup>6</sup> See <<http://ghgdata.epa.gov>>.

**Table 3-4: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from Fossil Fuel Combustion (kt)**

Gas	1990	2005	2013	2014	2015	2016	2017
CO <sub>2</sub>	4,738,756	5,744,754	5,157,391	5,199,345	5,047,107	4,961,876	4,911,962
CH <sub>4</sub>	862	696	531	518	483	453	440
N <sub>2</sub> O	225	246	184	178	166	161	153

## CO<sub>2</sub> from Fossil Fuel Combustion

Carbon dioxide is the primary gas emitted from fossil fuel combustion and represents the largest share of U.S. total greenhouse gas emissions. Carbon dioxide emissions from fossil fuel combustion are presented in Table 3-5. In 2017, CO<sub>2</sub> emissions from fossil fuel combustion decreased by 1.0 percent relative to the previous year. The decrease in CO<sub>2</sub> emissions from fossil fuel combustion was a result of multiple factors, primarily a continued shift from coal to natural gas and substitution from fossil to non-fossil energy sources in the electric power sector. In 2017, CO<sub>2</sub> emissions from fossil fuel combustion were 4,912.0 MMT CO<sub>2</sub> Eq., or 3.7 percent above emissions in 1990 (see Table 3-5).<sup>7</sup>

**Table 3-5: CO<sub>2</sub> Emissions from Fossil Fuel Combustion by Fuel Type and Sector (MMT CO<sub>2</sub> Eq.)**

Fuel/Sector	1990	2005	2013	2014	2015	2016	2017
<b>Coal</b>	<b>1,717.3</b>	<b>2,111.2</b>	<b>1,654.1</b>	<b>1,652.4</b>	<b>1,424.7</b>	<b>1,307.5</b>	<b>1,267.5</b>
Residential	3.0	0.8	0.0	0.0	0.0	0.0	0.0
Commercial	12.0	9.3	3.9	3.8	3.0	2.3	2.0
Industrial	155.2	115.3	76.0	76.0	66.3	59.2	54.4
Transportation	NE	NE	NE	NE	NE	NE	NE
Electric Power	1,546.5	1,982.8	1,571.3	1,568.6	1,351.4	1,242.0	1,207.1
U.S. Territories	0.6	3.0	2.8	4.0	4.0	4.0	4.0
<b>Natural Gas</b>	<b>999.7</b>	<b>1,167.0</b>	<b>1,391.9</b>	<b>1,420.0</b>	<b>1,460.2</b>	<b>1,471.8</b>	<b>1,450.3</b>
Residential	237.8	262.2	266.4	277.7	252.7	238.4	241.5
Commercial	142.0	162.9	179.2	189.2	175.4	170.5	173.2
Industrial	408.5	388.6	452.1	467.1	464.4	474.8	484.7
Transportation	36.0	33.1	47.0	40.2	39.4	40.1	42.3
Electric Power	175.4	318.9	444.2	442.9	525.2	545.0	505.6
U.S. Territories	NO	1.3	3.0	3.0	3.0	3.0	3.0
<b>Petroleum</b>	<b>2,021.2</b>	<b>2,466.2</b>	<b>2,111.0</b>	<b>2,126.5</b>	<b>2,161.8</b>	<b>2,182.1</b>	<b>2,193.7</b>
Residential	97.4	94.9	63.0	69.2	65.1	54.5	53.0
Commercial	72.6	54.6	41.5	39.9	67.1	59.3	57.7
Industrial	293.7	349.5	311.9	276.5	277.1	273.6	271.5
Transportation	1,433.1	1,823.9	1,635.6	1,681.3	1,694.6	1,739.0	1,758.3
Electric Power	97.5	97.9	22.4	25.3	23.7	21.4	18.9
U.S. Territories	26.9	45.4	36.6	34.3	34.3	34.3	34.3
<b>Geothermal<sup>a</sup></b>	<b>0.5</b>	<b>0.5</b>	<b>0.4</b>	<b>0.4</b>	<b>0.4</b>	<b>0.4</b>	<b>0.4</b>
<b>Total</b>	<b>4,738.8</b>	<b>5,744.8</b>	<b>5,157.4</b>	<b>5,199.3</b>	<b>5,047.1</b>	<b>4,961.9</b>	<b>4,912.0</b>

NE (Not Estimated)

NO (Not Occurring)

<sup>a</sup> Although not technically a fossil fuel, geothermal energy-related CO<sub>2</sub> emissions are included for reporting purposes.

Note: Totals may not sum due to independent rounding.

Trends in CO<sub>2</sub> emissions from fossil fuel combustion are influenced by many long-term and short-term factors. On a year-to-year basis, the overall demand for fossil fuels in the United States and other countries generally fluctuates in response to changes in general economic conditions, energy prices, weather, and the availability of non-fossil alternatives. For example, in a year with increased consumption of goods and services, low fuel prices, severe

<sup>7</sup> An additional discussion of fossil fuel emission trends is presented in the Trends in U.S. Greenhouse Gas Emissions chapter.

summer and winter weather conditions, nuclear plant closures, and lower precipitation feeding hydroelectric dams, there would likely be proportionally greater fossil fuel consumption than a year with poor economic performance, high fuel prices, mild temperatures, and increased output from nuclear and hydroelectric plants.

Longer-term changes in energy usage patterns, however, tend to be more a function of aggregate societal trends that affect the scale of energy use (e.g., population, number of cars, size of houses, and number of houses), the efficiency with which energy is used in equipment (e.g., cars, power plants, steel mills, and light bulbs), and social planning and consumer behavior (e.g., walking, bicycling, or telecommuting to work instead of driving).

Carbon dioxide emissions also depend on the source of energy and its carbon (C) intensity. The amount of C in fuels varies significantly by fuel type. For example, coal contains the highest amount of C per unit of useful energy.

Petroleum has roughly 75 percent of the C per unit of energy as coal, and natural gas has only about 55 percent.<sup>8</sup>

Table 3-6 shows annual changes in emissions during the last five years for coal, petroleum, and natural gas in selected sectors.

**Table 3-6: Annual Change in CO<sub>2</sub> Emissions and Total 2017 CO<sub>2</sub> Emissions from Fossil Fuel Combustion for Selected Fuels and Sectors (MMT CO<sub>2</sub> Eq. and Percent)**

Sector	Fuel Type	2013 to 2014		2014 to 2015		2015 to 2016		2016 to 2017		Total 2017
Electric Power	Coal	-2.7	-0.2%	-217.2	-13.8%	-109.4	-8.1%	-34.9	-2.8%	1,207.1
Electric Power	Natural Gas	-1.3	-0.3%	82.3	18.6%	19.8	3.8%	-39.4	-7.2%	505.6
Transportation	Petroleum	45.7	2.8%	13.2	0.8%	44.4	2.6%	19.3	1.1%	1,758.3
Residential	Natural Gas	11.3	4.3%	-24.9	-9.0%	-14.3	-5.7%	3.1	1.3%	241.5
Commercial	Natural Gas	10.0	5.6%	-13.8	-7.3%	-4.9	-2.8%	2.6	1.5%	173.2
Industrial	Natural Gas	15.0	3.3%	-2.6	-0.6%	10.4	2.2%	9.9	2.1%	484.7
<b>All Sectors<sup>a</sup></b>	<b>All Fuels<sup>a</sup></b>	<b>42.0</b>	<b>0.8%</b>	<b>-152.2</b>	<b>-2.9%</b>	<b>-85.2</b>	<b>-1.7%</b>	<b>-49.9</b>	<b>-1.0%</b>	<b>4,912.0</b>

<sup>a</sup> Includes sector and fuel combinations not shown in this table.

As shown in Table 3-6, recent trends in CO<sub>2</sub> emissions from fossil fuel combustion show a 0.8 percent increase from 2013 to 2014, then a 2.9 percent decrease from 2014 to 2015, then a 1.7 percent decrease from 2015 to 2016, and a 1.0 percent decrease from 2016 to 2017. These changes contributed to a 4.8 percent decrease in CO<sub>2</sub> emissions from fossil fuel combustion from 2013 to 2017.

Trends in CO<sub>2</sub> emissions from fossil fuel combustion over the past five years have been in large part driven by the electric power sector, which historically has accounted for the largest portion of these emissions. The types of fuels consumed to produce electricity have changed in recent years. Total electric power generation remained relatively flat over the past five years, but emissions have decreased due to a decreasing reliance on coal used to generate electricity. Carbon dioxide emissions from coal consumption for electric power generation decreased by 23.2 percent since 2013, which can be largely attributed to a shift to the use of less-CO<sub>2</sub>-intensive natural gas to generate electricity and a rapid increase in renewable energy capacity additions in the electric power sector in recent years.

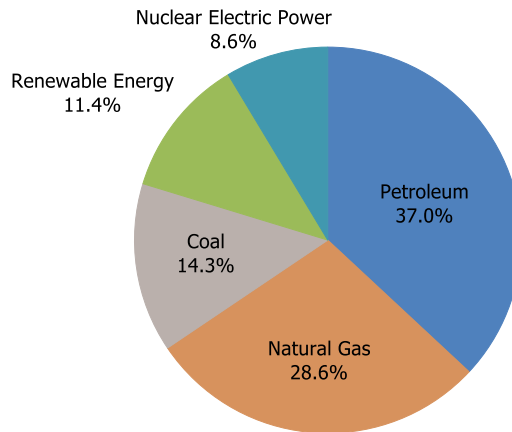
The trends in CO<sub>2</sub> emissions from fossil fuel combustion over the past five years also follow changes in heating degree days. Emissions from natural gas consumption in the residential and commercial sectors decreased by 9.3 percent and 3.4 percent from 2013 to 2017, respectively. This trend can be largely attributed to a 14 percent decrease in heating degree days, which led to a decreased demand for heating fuel and electricity for heat in these sectors. In addition, an increase in energy efficiency standards and the use of energy-efficient products in residential and commercial buildings has resulted in an overall reduction in energy use, contributing to a decrease in CO<sub>2</sub> emissions in both of these sectors (EIA 2018e).

Petroleum use in the transportation sector is another major driver of emissions, representing the largest source of CO<sub>2</sub> emissions from fossil fuel combustion in 2017. Despite the overall decreasing trend in CO<sub>2</sub> emissions from fossil fuel combustion over the past five years, emissions from petroleum consumption for transportation have increased by 7.5 percent since 2013; this trend can be primarily attributed to a 7.5 percent increase in vehicle miles traveled (VMT) over the same time period.

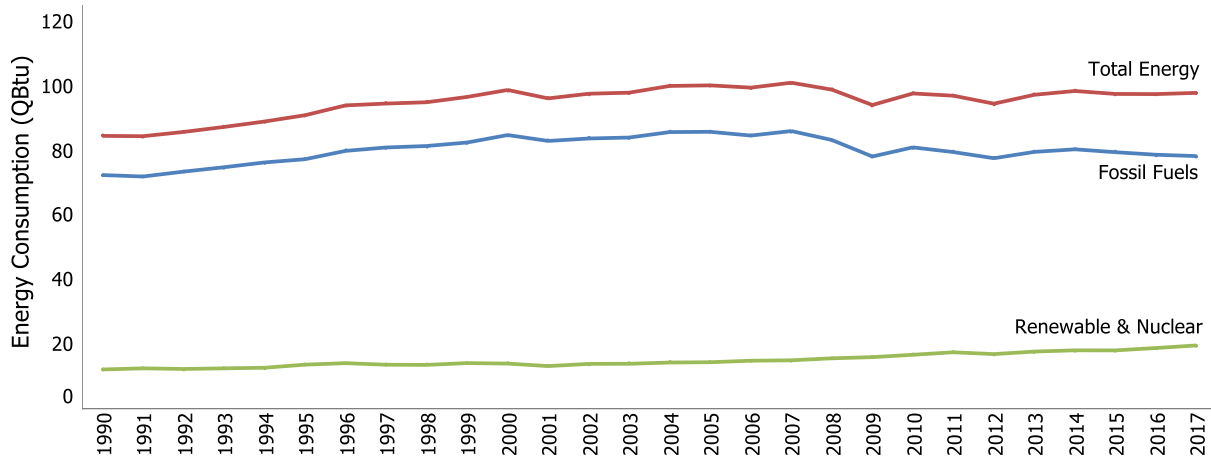
<sup>8</sup> Based on national aggregate carbon content of all coal, natural gas, and petroleum fuels combusted in the United States.

In the United States, 80 percent of the energy used in 2017 was produced through the combustion of fossil fuels such as coal, natural gas, and petroleum (see Figure 3-3 and Figure 3-4). The remaining portion was supplied by nuclear electric power (9 percent) and by a variety of renewable energy sources (11 percent), primarily hydroelectric power, wind energy and biofuels (EIA 2019a).<sup>9</sup> Specifically, petroleum supplied the largest share of domestic energy demands, accounting for 37 percent of total U.S. energy used in 2017. Natural gas and coal followed in order of energy demand importance, accounting for approximately 29 percent and 14 percent of total U.S. energy used, respectively. Petroleum was consumed primarily in the transportation end-use sector and the vast majority of coal was used in the electric power sector. Natural gas was broadly consumed in all end-use sectors except transportation (see Figure 3-5) (EIA 2019a).

**Figure 3-3: 2017 U.S. Energy Use by Energy Source (Percent)**



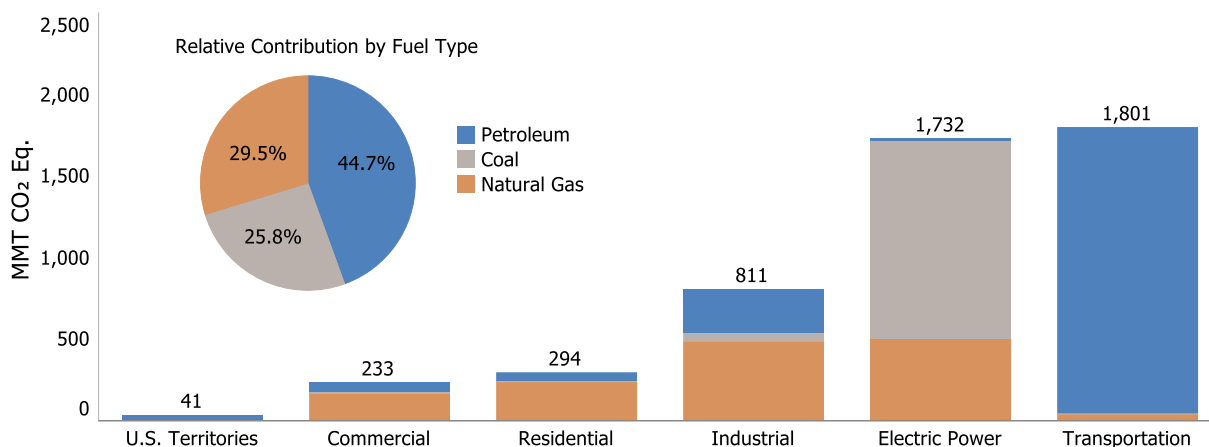
**Figure 3-4: U.S. Energy Use (Quadrillion Btu)**



<sup>9</sup> Renewable energy, as defined in EIA’s energy statistics, includes the following energy sources: hydroelectric power, geothermal energy, biofuels, solar energy, and wind energy.



**Figure 3-5: 2017 CO<sub>2</sub> Emissions from Fossil Fuel Combustion by Sector and Fuel Type (MMT CO<sub>2</sub> Eq.)**



Fossil fuels are generally combusted for the purpose of producing energy for useful heat and work. During the combustion process, the C stored in the fuels is oxidized and emitted as CO<sub>2</sub> and smaller amounts of other gases, including CH<sub>4</sub>, CO, and NMVOCs.<sup>10</sup> These other C-containing non-CO<sub>2</sub> gases are emitted as a byproduct of incomplete fuel combustion, but are, for the most part, eventually oxidized to CO<sub>2</sub> in the atmosphere. Therefore, it is assumed all of the C in fossil fuels used to produce energy is eventually converted to atmospheric CO<sub>2</sub>.

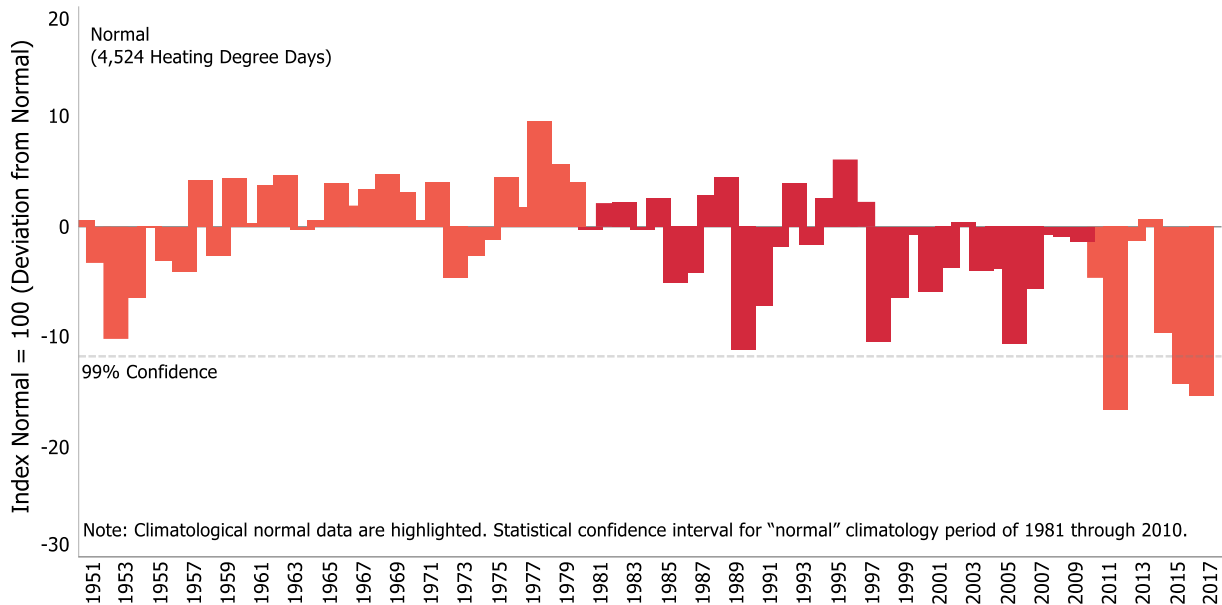
**Box 3-3: Weather and Non-Fossil Energy Effects on CO<sub>2</sub> from Fossil Fuel Combustion Trends**

The United States in 2017 experienced a warmer winter overall compared to 2016, as heating degree days decreased (1.3 percent). Warmer winter conditions compared to 2016 impacted the amount of energy required for heating, and heating degree days in the United States were 15.4 percent below normal (see Figure 3-6). Cooling degree days decreased by 8.4 percent compared to 2016, which decreased demand for air conditioning in the residential and commercial sector. This led in part to an overall residential electricity demand decrease of 2.3 percent. Summer conditions in 2017 were still warmer than normal, with cooling degree days 17.4 percent above normal (see Figure 3-7) (EIA 2019a).<sup>11</sup>

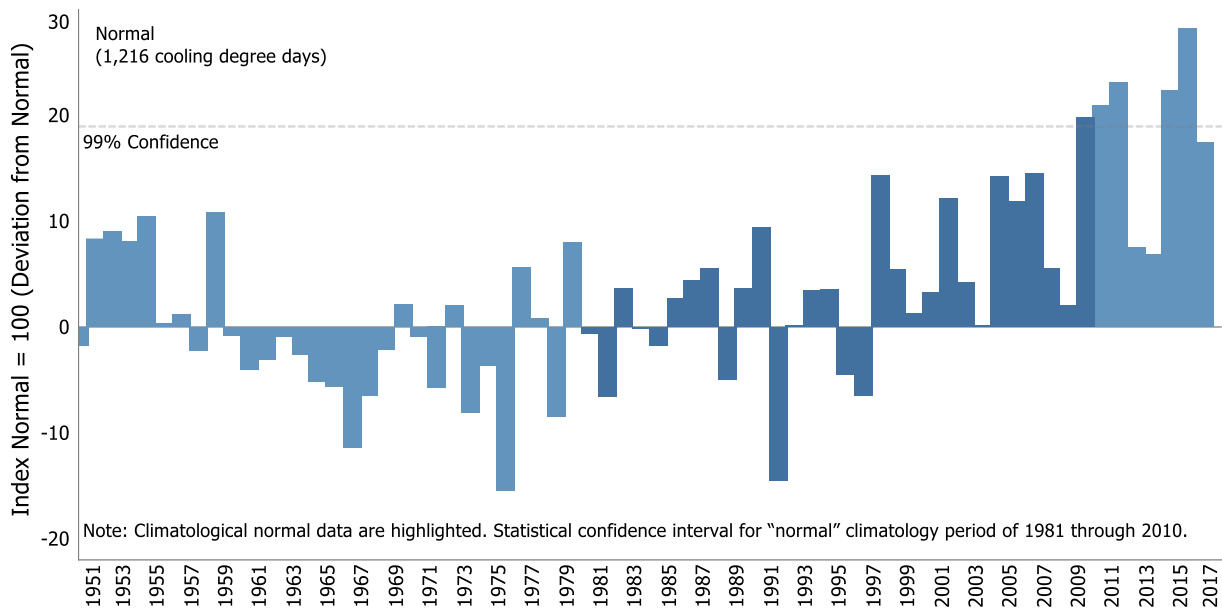
<sup>10</sup> See the sections entitled Stationary Combustion and Mobile Combustion in this chapter for information on non-CO<sub>2</sub> gas emissions from fossil fuel combustion.

<sup>11</sup> Degree days are relative measurements of outdoor air temperature. Heating degree days are deviations of the mean daily temperature below 65 degrees Fahrenheit, while cooling degree days are deviations of the mean daily temperature above 65 degrees Fahrenheit. Heating degree days have a considerably greater effect on energy demand and related emissions than do cooling degree days. Excludes Alaska and Hawaii. Normals are based on data from 1981 through 2010. The variation in these normals during this time period was ±12 percent and ±19 percent for heating and cooling degree days, respectively (99 percent confidence interval).

**Figure 3-6: Annual Deviations from Normal Heating Degree Days for the United States (1950–2017, Index Normal = 100)**



**Figure 3-7: Annual Deviations from Normal Cooling Degree Days for the United States (1950–2017, Index Normal = 100)**



The carbon intensity of the electric power sector is impacted by the amount of non-fossil energy sources of electricity. The utilization (i.e., capacity factors)<sup>12</sup> of nuclear power plants in 2017 remained high at 92 percent. In

<sup>12</sup> The capacity factor equals generation divided by net summer capacity. Summer capacity is defined as “The maximum output that generating equipment can supply to system load, as demonstrated by a multi-hour test, at the time of summer peak demand (period of June 1 through September 30).” Data for both the generation and net summer capacity are from EIA (2018d).

2017, nuclear power represented 21 percent of total electricity generation. In recent years, the wind and solar power sectors have shown strong growth, such that, on the margin, they are becoming relatively important electricity sources. Between 1990 and 2017, renewable energy generation (in kWh) from solar and wind energy have increased from 0.1 percent in 1990 to 8 percent of total electricity generation in 2017, which helped drive the decrease in the carbon intensity of the electricity supply in the United States.

## Fossil Fuel Combustion Emissions by Sector

In addition to the CO<sub>2</sub> emitted from fossil fuel combustion, CH<sub>4</sub> and N<sub>2</sub>O are emitted from stationary and mobile combustion as well. Table 3-7 provides an overview of the CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O emissions from fossil fuel combustion by sector.

**Table 3-7: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from Fossil Fuel Combustion by Sector (MMT CO<sub>2</sub> Eq.)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
<b>Transportation</b>	<b>1,524.1</b>	<b>1,905.6</b>	<b>1,709.3</b>	<b>1,745.9</b>	<b>1,756.4</b>	<b>1,800.3</b>	<b>1,820.7</b>
CO <sub>2</sub>	1,469.1	1,857.0	1,682.7	1,721.6	1,734.0	1,779.0	1,800.6
CH <sub>4</sub>	12.9	9.6	4.5	4.1	3.6	3.4	3.2
N <sub>2</sub> O	42.0	39.0	22.1	20.2	18.8	17.9	16.9
<b>Electric Power</b>	<b>1,840.9</b>	<b>2,430.9</b>	<b>2,067.9</b>	<b>2,067.1</b>	<b>1,928.3</b>	<b>1,836.2</b>	<b>1,757.9</b>
CO <sub>2</sub>	1,820.0	2,400.0	2,038.3	2,037.1	1,900.6	1,808.9	1,732.0
CH <sub>4</sub>	0.4	0.9	1.0	1.1	1.2	1.2	1.1
N <sub>2</sub> O	20.5	30.1	28.6	28.9	26.5	26.2	24.8
<b>Industrial</b>	<b>862.4</b>	<b>858.1</b>	<b>844.5</b>	<b>823.9</b>	<b>812.2</b>	<b>811.8</b>	<b>814.9</b>
CO <sub>2</sub>	857.5	853.4	840.0	819.6	807.9	807.6	810.7
CH <sub>4</sub>	1.8	1.7	1.7	1.6	1.6	1.6	1.6
N <sub>2</sub> O	3.1	2.9	2.8	2.7	2.7	2.6	2.7
<b>Residential</b>	<b>344.5</b>	<b>362.8</b>	<b>335.2</b>	<b>352.8</b>	<b>323.2</b>	<b>297.6</b>	<b>299.0</b>
CO <sub>2</sub>	338.2	357.9	329.3	346.8	317.8	292.9	294.5
CH <sub>4</sub>	5.2	4.1	4.9	5.0	4.5	3.9	3.8
N <sub>2</sub> O	1.0	0.9	1.0	1.0	0.9	0.8	0.8
<b>Commercial</b>	<b>228.0</b>	<b>228.2</b>	<b>226.0</b>	<b>234.3</b>	<b>247.0</b>	<b>233.7</b>	<b>234.4</b>
CO <sub>2</sub>	226.5	226.8	224.6	232.9	245.5	232.1	232.9
CH <sub>4</sub>	1.1	1.1	1.1	1.1	1.2	1.2	1.2
N <sub>2</sub> O	0.4	0.3	0.3	0.3	0.4	0.3	0.3
<b>U.S. Territories<sup>a</sup></b>	<b>27.7</b>	<b>49.9</b>	<b>42.6</b>	<b>41.5</b>	<b>41.5</b>	<b>41.5</b>	<b>41.5</b>
<b>Total</b>	<b>4,827.4</b>	<b>5,835.5</b>	<b>5,225.5</b>	<b>5,265.5</b>	<b>5,108.6</b>	<b>5,021.1</b>	<b>4,968.5</b>

<sup>a</sup> U.S. Territories are not apportioned by sector, and emissions are total greenhouse gas emissions from all fuel combustion sources.

Notes: Totals may not sum due to independent rounding.

Other than CO<sub>2</sub>, gases emitted from stationary combustion include the greenhouse gases CH<sub>4</sub> and N<sub>2</sub>O and greenhouse gas precursors NO<sub>x</sub>, CO, and NMVOCs.<sup>13</sup> Methane and N<sub>2</sub>O emissions from stationary combustion sources depend upon fuel characteristics, size and vintage, along with combustion technology, pollution control equipment, ambient environmental conditions, and operation and maintenance practices. Nitrous oxide emissions from stationary combustion are closely related to air-fuel mixes and combustion temperatures, as well as the characteristics of any pollution control equipment that is employed. Methane emissions from stationary combustion are primarily a function of the CH<sub>4</sub> content of the fuel and combustion efficiency.

Mobile combustion produces greenhouse gases other than CO<sub>2</sub>, including CH<sub>4</sub>, N<sub>2</sub>O, and greenhouse gas precursors including NO<sub>x</sub>, CO, and NMVOCs. As with stationary combustion, N<sub>2</sub>O and NO<sub>x</sub> emissions from mobile combustion are closely related to fuel characteristics, air-fuel mixes, combustion temperatures, and the use of pollution control equipment. Nitrous oxide from mobile sources, in particular, can be formed by the catalytic

<sup>13</sup> Sulfur dioxide (SO<sub>2</sub>) emissions from stationary combustion are addressed in Annex 6.3.

processes used to control NO<sub>x</sub>, CO, and hydrocarbon emissions. Carbon monoxide emissions from mobile combustion are significantly affected by combustion efficiency and the presence of post-combustion emission controls. Carbon monoxide emissions are highest when air-fuel mixtures have less oxygen than required for complete combustion. These emissions occur especially in idle, low speed, and cold start conditions. Methane and NMVOC emissions from motor vehicles are a function of the CH<sub>4</sub> content of the motor fuel, the amount of hydrocarbons passing uncombusted through the engine, and any post-combustion control of hydrocarbon emissions (such as catalytic converters).

An alternative method of presenting combustion emissions is to allocate emissions associated with electric power to the sectors in which it is used. Four end-use sectors were defined: industrial, transportation, residential, and commercial. In the table below, electric power emissions have been distributed to each end-use sector based upon the sector's share of national electricity use, with the exception of CH<sub>4</sub> and N<sub>2</sub>O from transportation.<sup>14,15</sup> Emissions from U.S. Territories are also calculated separately due to a lack of end-use-specific consumption data.<sup>16</sup> This method assumes that emissions from combustion sources are distributed across the four end-use sectors based on the ratio of electricity use in that sector. The results of this alternative method are presented in Table 3-8.

**Table 3-8: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from Fossil Fuel Combustion by End-Use Sector (MMT CO<sub>2</sub> Eq.)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
<b>Transportation</b>	<b>1,527.1</b>	<b>1,910.3</b>	<b>1,713.6</b>	<b>1,750.3</b>	<b>1,760.6</b>	<b>1,804.5</b>	<b>1,825.0</b>
CO <sub>2</sub>	1,472.1	1,861.7	1,686.9	1,726.0	1,738.2	1,783.2	1,804.9
CH <sub>4</sub>	12.9	9.6	4.5	4.1	3.6	3.4	3.2
N <sub>2</sub> O	42.0	39.0	22.1	20.2	18.8	17.9	16.9
<b>Industrial</b>	<b>1,556.7</b>	<b>1,603.9</b>	<b>1,447.9</b>	<b>1,425.6</b>	<b>1,369.7</b>	<b>1,337.2</b>	<b>1,326.9</b>
CO <sub>2</sub>	1,543.9	1,589.7	1,434.8	1,412.5	1,357.4	1,325.2	1,315.1
CH <sub>4</sub>	2.0	2.0	2.0	2.0	2.0	1.9	1.9
N <sub>2</sub> O	10.8	12.2	11.1	11.1	10.4	10.1	9.9
<b>Residential</b>	<b>944.0</b>	<b>1,229.9</b>	<b>1,080.6</b>	<b>1,097.7</b>	<b>1,016.9</b>	<b>961.0</b>	<b>925.4</b>
CO <sub>2</sub>	930.9	1,213.9	1,064.1	1,080.9	1,001.6	946.3	911.5
CH <sub>4</sub>	5.4	4.4	5.3	5.4	4.9	4.3	4.2
N <sub>2</sub> O	7.7	11.6	11.3	11.4	10.5	10.3	9.6
<b>Commercial</b>	<b>771.9</b>	<b>1,041.5</b>	<b>940.8</b>	<b>950.3</b>	<b>919.8</b>	<b>876.9</b>	<b>849.7</b>
CO <sub>2</sub>	764.3	1,029.7	929.1	938.5	908.5	865.8	839.1
CH <sub>4</sub>	1.2	1.4	1.4	1.5	1.6	1.6	1.6
N <sub>2</sub> O	6.5	10.4	10.2	10.3	9.6	9.5	9.0
<b>U.S. Territories<sup>a</sup></b>	<b>27.7</b>	<b>49.9</b>	<b>42.6</b>	<b>41.5</b>	<b>41.5</b>	<b>41.5</b>	<b>41.5</b>
<b>Total</b>	<b>4,827.4</b>	<b>5,835.5</b>	<b>5,225.5</b>	<b>5,265.5</b>	<b>5,108.6</b>	<b>5,021.1</b>	<b>4,968.5</b>

<sup>a</sup> U.S. Territories are not apportioned by sector, and emissions are total greenhouse gas emissions from all fuel combustion sources.

Notes: Totals may not sum due to independent rounding. Emissions from fossil fuel combustion by electric power are allocated based on aggregate national electricity use by each end-use sector.

## Stationary Combustion

The direct combustion of fuels by stationary sources in the electric power, industrial, commercial, and residential

<sup>14</sup> Separate calculations were performed for transportation-related CH<sub>4</sub> and N<sub>2</sub>O. The methodology used to calculate these emissions are discussed in the Mobile Combustion section.

<sup>15</sup> In this year's Inventory, electricity use from electric vehicle charging in commercial and residential locations was re-allocated from the residential and commercial sectors to the transportation sector. These changes apply to the time period from 2010 through 2017. Electricity consumption by passenger cars, light-duty trucks (SUVs), and buses is based on plug-in electric vehicle sales and engine efficiency data, as outlined in Browning (2018a).

<sup>16</sup> U.S. Territories consumption data that are obtained from EIA are only available at the aggregate level and cannot be broken out by end-use sector. The distribution of emissions to each end-use sector for the 50 states does not apply to territories data.

sectors represent the greatest share of U.S. greenhouse gas emissions. Table 3-9 presents CO<sub>2</sub> emissions from fossil fuel combustion by stationary sources. The CO<sub>2</sub> emitted is closely linked to the type of fuel being combusted in each sector (see Methodology section of CO<sub>2</sub> from Fossil Fuel Combustion). Other than CO<sub>2</sub>, gases emitted from stationary combustion include the greenhouse gases CH<sub>4</sub> and N<sub>2</sub>O. Table 3-10 and Table 3-11 present CH<sub>4</sub> and N<sub>2</sub>O emissions from the combustion of fuels in stationary sources. The CH<sub>4</sub> and N<sub>2</sub>O emission estimation methodology utilizes facility-specific technology and fuel use data reported to EPA's Acid Rain Program (EPA 2018a) (see Methodology section for CH<sub>4</sub> and N<sub>2</sub>O from Stationary Combustion). Table 3-7 presents the corresponding direct CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O emissions from all sources of fuel combustion, without allocating emissions from electricity use to the end-use sectors.

**Table 3-9: CO<sub>2</sub> Emissions from Stationary Fossil Fuel Combustion (MMT CO<sub>2</sub> Eq.)**

Sector/Fuel Type	1990	2005	2013	2014	2015	2016	2017
<b>Electric Power</b>	<b>1,820.0</b>	<b>2,400.0</b>	<b>2,038.3</b>	<b>2,037.1</b>	<b>1,900.6</b>	<b>1,808.9</b>	<b>1,732.0</b>
Coal	1,546.5	1,982.8	1,571.3	1,568.6	1,351.4	1,242.0	1,207.1
Natural Gas	175.4	318.9	444.2	442.9	525.2	545.0	505.6
Fuel Oil	97.5	97.9	22.4	25.3	23.7	21.4	18.9
Geothermal	0.5	0.5	0.4	0.4	0.4	0.4	0.4
<b>Industrial</b>	<b>857.5</b>	<b>853.4</b>	<b>840.0</b>	<b>819.6</b>	<b>807.9</b>	<b>807.6</b>	<b>810.7</b>
Coal	155.2	115.3	76.0	76.0	66.3	59.2	54.4
Natural Gas	408.5	388.6	452.1	467.1	464.4	474.8	484.7
Fuel Oil	293.7	349.5	311.9	276.5	277.1	273.6	271.5
<b>Commercial</b>	<b>226.5</b>	<b>226.8</b>	<b>224.6</b>	<b>232.9</b>	<b>245.5</b>	<b>232.1</b>	<b>232.9</b>
Coal	12.0	9.3	3.9	3.8	3.0	2.3	2.0
Natural Gas	142.0	162.9	179.2	189.2	175.4	170.5	173.2
Fuel Oil	72.6	54.6	41.5	39.9	67.1	59.3	57.7
<b>Residential</b>	<b>338.2</b>	<b>357.9</b>	<b>329.3</b>	<b>346.8</b>	<b>317.8</b>	<b>292.9</b>	<b>294.5</b>
Coal	3.0	0.8	0.0	0.0	0.0	0.0	0.0
Natural Gas	237.8	262.2	266.4	277.7	252.7	238.4	241.5
Fuel Oil	97.4	94.9	63.0	69.2	65.1	54.5	53.0
<b>U.S. Territories</b>	<b>27.6</b>	<b>49.7</b>	<b>42.5</b>	<b>41.4</b>	<b>41.4</b>	<b>41.4</b>	<b>41.4</b>
Coal	0.6	3.0	2.8	4.0	4.0	4.0	4.0
Natural Gas	NO	1.3	3.0	3.0	3.0	3.0	3.0
Fuel Oil	26.9	45.4	36.6	34.3	34.3	34.3	34.3
<b>Total</b>	<b>3,269.7</b>	<b>3,887.8</b>	<b>3,474.7</b>	<b>3,477.8</b>	<b>3,313.1</b>	<b>3,182.8</b>	<b>3,111.4</b>

NO (Not Occurring)

Note: Totals may not sum due to independent rounding.

**Table 3-10: CH<sub>4</sub> Emissions from Stationary Combustion (MMT CO<sub>2</sub> Eq.)**

Sector/Fuel Type	1990	2005	2013	2014	2015	2016	2017
<b>Electric Power</b>	<b>0.4</b>	<b>0.9</b>	<b>1.0</b>	<b>1.1</b>	<b>1.2</b>	<b>1.2</b>	<b>1.1</b>
Coal	0.3	0.4	0.3	0.3	0.3	0.2	0.2
Fuel Oil	+	+	+	+	+	+	+
Natural gas	0.1	0.5	0.7	0.8	0.9	0.9	0.9
Wood	+	+	+	+	+	+	+
<b>Industrial</b>	<b>1.8</b>	<b>1.7</b>	<b>1.7</b>	<b>1.6</b>	<b>1.6</b>	<b>1.6</b>	<b>1.6</b>
Coal	0.4	0.3	0.2	0.2	0.2	0.2	0.1
Fuel Oil	0.2	0.2	0.2	0.2	0.2	0.2	0.2
Natural gas	0.2	0.2	0.2	0.2	0.2	0.2	0.2
Wood	1.0	1.0	1.1	1.1	1.1	1.0	1.1
<b>Commercial</b>	<b>1.1</b>	<b>1.1</b>	<b>1.1</b>	<b>1.1</b>	<b>1.2</b>	<b>1.2</b>	<b>1.2</b>
Coal	+	+	+	+	+	+	+
Fuel Oil	0.3	0.2	0.1	0.1	0.2	0.2	0.2
Natural gas	0.3	0.4	0.4	0.4	0.4	0.4	0.4
Wood	0.5	0.5	0.5	0.5	0.6	0.6	0.6

<b>Residential</b>	<b>5.2</b>	<b>4.1</b>	<b>4.9</b>	<b>5.0</b>	<b>4.5</b>	<b>3.9</b>	<b>3.8</b>
Coal	0.2	0.1	0.0	0.0	0.0	0.0	0.0
Fuel Oil	0.3	0.3	0.2	0.3	0.2	0.2	0.2
Natural Gas	0.5	0.6	0.6	0.6	0.6	0.5	0.5
Wood	4.1	3.1	4.1	4.1	3.7	3.2	3.1
<b>U.S. Territories</b>	<b>+</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>
Coal	+	+	+	+	+	+	+
Fuel Oil	+	0.1	0.1	0.1	0.1	0.1	0.1
Natural Gas	NO	+	+	+	+	+	+
Wood	NO	NO	NO	NO	NO	NO	NO
<b>Total</b>	<b>8.6</b>	<b>7.8</b>	<b>8.7</b>	<b>8.9</b>	<b>8.5</b>	<b>7.9</b>	<b>7.8</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub> Eq.

NO (Not Occurring)

Note: Totals may not sum due to independent rounding.

**Table 3-11: N<sub>2</sub>O Emissions from Stationary Combustion (MMT CO<sub>2</sub> Eq.)**

Sector/Fuel Type	1990	2005	2013	2014	2015	2016	2017
<b>Electric Power</b>	<b>20.5</b>	<b>30.1</b>	<b>28.6</b>	<b>28.9</b>	<b>26.5</b>	<b>26.2</b>	<b>24.8</b>
Coal	20.1	28.0	25.5	25.7	22.8	22.4	21.2
Fuel Oil	0.1	0.1	+	+	+	+	+
Natural Gas	0.3	1.9	3.1	3.1	3.7	3.8	3.6
Wood	+	+	+	+	+	+	+
<b>Industrial</b>	<b>3.1</b>	<b>2.9</b>	<b>2.8</b>	<b>2.7</b>	<b>2.7</b>	<b>2.6</b>	<b>2.7</b>
Coal	0.7	0.5	0.4	0.4	0.3	0.3	0.3
Fuel Oil	0.5	0.5	0.5	0.4	0.4	0.4	0.4
Natural Gas	0.2	0.2	0.2	0.2	0.2	0.3	0.3
Wood	1.6	1.6	1.7	1.7	1.7	1.7	1.7
<b>Commercial</b>	<b>0.4</b>	<b>0.3</b>	<b>0.3</b>	<b>0.3</b>	<b>0.4</b>	<b>0.3</b>	<b>0.3</b>
Coal	0.1	+	+	+	+	+	+
Fuel Oil	0.2	0.1	0.1	0.1	0.2	0.2	0.1
Natural Gas	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Wood	0.1	0.1	0.1	0.1	0.1	0.1	0.1
<b>Residential</b>	<b>1.0</b>	<b>0.9</b>	<b>1.0</b>	<b>1.0</b>	<b>0.9</b>	<b>0.8</b>	<b>0.8</b>
Coal	+	+	0.0	0.0	0.0	0.0	0.0
Fuel Oil	0.2	0.2	0.2	0.2	0.2	0.1	0.1
Natural Gas	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Wood	0.7	0.5	0.6	0.7	0.6	0.5	0.5
<b>U.S. Territories</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>
Coal	+	+	+	+	+	+	+
Fuel Oil	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Natural Gas	NO	+	+	+	+	+	+
Wood	NO	NO	NO	NO	NO	NO	NO
<b>Total</b>	<b>25.1</b>	<b>34.3</b>	<b>32.7</b>	<b>33.0</b>	<b>30.6</b>	<b>30.1</b>	<b>28.6</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub> Eq.

NO (Not Occurring)

Note: Totals may not sum due to independent rounding.

## Electric Power Sector

The process of generating electricity is the largest stationary source of CO<sub>2</sub> emissions in the United States, representing 33 percent of total CO<sub>2</sub> emissions from all CO<sub>2</sub> emissions sources across the United States. Methane and N<sub>2</sub>O accounted for a small portion of total greenhouse gas emissions from electric power, representing 0.1 percent and 1.4 percent, respectively. Electric power also accounted for 35.3 percent of CO<sub>2</sub> emissions from fossil fuel combustion in 2017. Methane and N<sub>2</sub>O from electric power represented 10.0 and 54.4 percent of total CH<sub>4</sub> and N<sub>2</sub>O emissions from fossil fuel combustion in 2017, respectively.

For the underlying energy data used in this chapter, the Energy Information Administration (EIA) places electric power generation into three functional categories: the electric power sector, the commercial sector, and the industrial sector. The electric power sector consists of electric utilities and independent power producers whose primary business is the production of electricity. This includes both regulated utilities and non-utilities (e.g., independent power producers, qualifying co-generators, and other small power producers). Electric generation is reported as occurring in other sectors where the producer of the power indicates that its primary business is something other than the production of electricity.<sup>17</sup>

Total emissions from the electric power sector have decreased by 4.5 percent since 1990. The carbon intensity of the electric power sector, in terms of CO<sub>2</sub> Eq. per QBTu, has decreased by 11 percent during that same timeframe with the majority of the emissions and carbon intensity decreases occurring in the past decade as shown below in Figure 3-8. This recent decarbonization of the electric power sector is a result of several key drivers. Coal-fired electric generation (in kilowatt-hours [kWh]) decreased from 54 percent of generation in 1990 to 31 percent in 2017.<sup>18</sup> This corresponded with an increase in natural gas generation and renewable energy generation, largely from wind and solar energy. Natural gas generation (in kWh) represented 11 percent of electric power generation in 1990, and increased over the 28-year period to represent 31 percent of electric power sector generation in 2017 (see Table 3-12).

**Table 3-12: Electric Power Generation by Fuel Type (Percent)**

Fuel Type	1990	2005	2013	2014	2015	2016	2017
Coal	54.1%	51.1%	40.2%	39.9%	34.2%	31.4%	30.9%
Natural Gas	10.7%	17.5%	26.4%	26.3%	31.6%	32.7%	30.9%
Nuclear	19.9%	20.0%	20.2%	20.3%	20.4%	20.6%	20.8%
Renewables	11.3%	8.3%	12.5%	12.8%	13.0%	14.7%	16.8%
Petroleum	4.1%	3.0%	0.6%	0.7%	0.7%	0.6%	0.5%
Other Gases <sup>a</sup>	+	0.1%	0.1%	0.1%	0.1%	0.1%	0.1%
<i>Net Electricity Generation (Billion kWh)<sup>b</sup></i>	2,905	3,902	3,901	3,936	3,917	3,917	3,877

+ Does not exceed 0.05 percent

<sup>a</sup> Other gases include blast furnace gas, propane gas, and other manufactured and waste gases derived from fossil fuels.

<sup>b</sup> Represents net electricity generation from the electric power sector. Excludes net electricity generation from commercial and industrial combined-heat-and-power and electricity-only plants.

In 2017, CO<sub>2</sub> emissions from the electric power sector decreased by 4.2 percent relative to 2016. This decrease in CO<sub>2</sub> emissions was a result of a decrease in fossil fuels consumed to produce electricity in the electric power sector. Consumption of coal and natural gas for electric power decreased by 2.9 percent and 7.2 percent, respectively, from 2016 to 2017. There has also been a rapid increase in renewable energy electricity generation in the electric power sector in recent years. Electricity generation from renewable sources increased by 14 percent from 2016 to 2017 (see Table 3-12). The decrease in coal-powered electricity generation and increase in renewable energy electricity generation contributed to a decrease in emissions from electric power generation over the time series (see Figure 3-8).

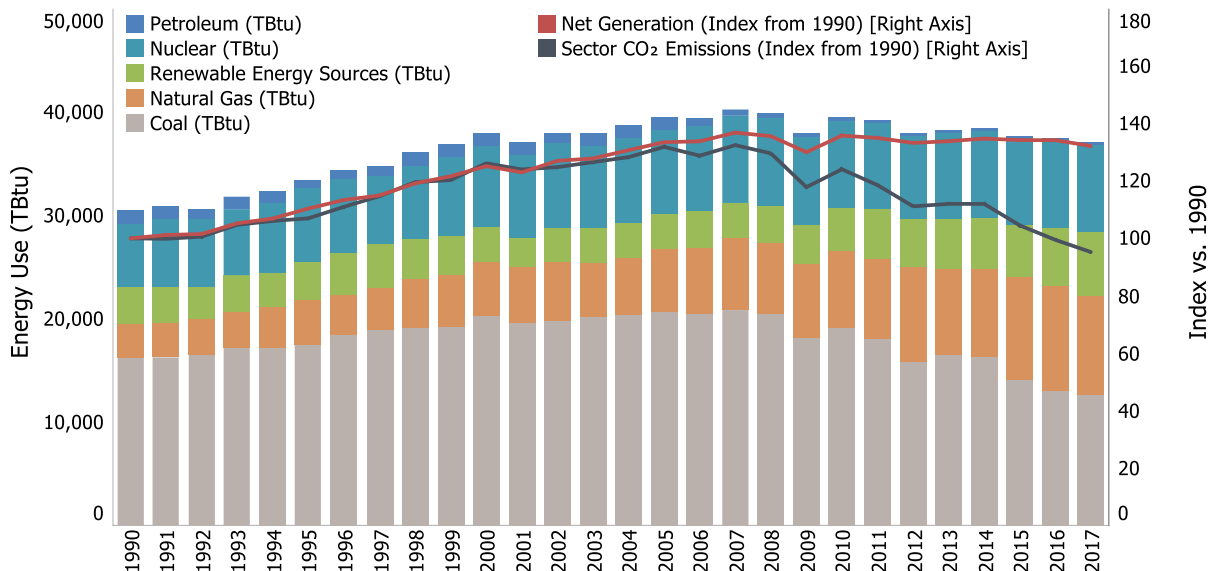
Decreases in natural gas costs and the associated increase in natural gas generation, particularly between 2005 and 2017, was one of the main drivers of the recent fuel switching and decrease in electric power sector carbon intensity. During this time period, the cost of natural gas (in \$/MMBtu) decreased by 49 percent while the cost of coal (in \$/MMBtu) increased by 78 percent (EIA 2019a). Also, between 1990 and 2017, renewable energy generation (in kWh) from wind and solar energy have increased from 0.1 percent of total generation in 1990 to 8 percent in 2017, which also helped drive the decrease in electric power sector carbon intensity. This decrease in carbon intensity

<sup>17</sup> Utilities primarily generate power for the U.S. electric grid for sale to retail customers. Non-utilities typically generate electricity for sale on the wholesale electricity market (e.g., to utilities for distribution and resale to retail customers). Where electricity generation occurs outside the EIA-defined electric power sector, it is typically for the entity's own use.

<sup>18</sup> Values represent electricity *net* generation from the electric power sector (EIA 2019a).

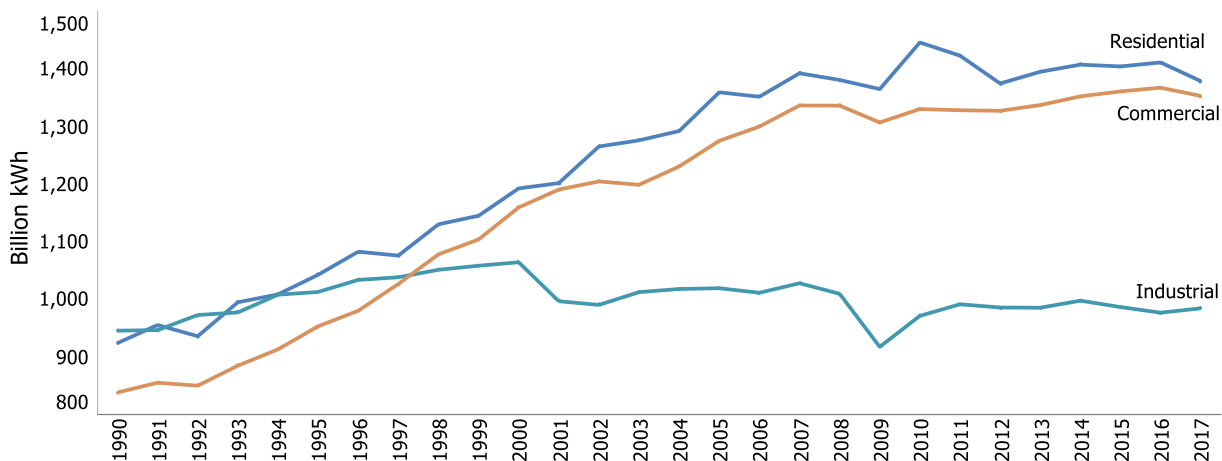
occurred even as total electricity retail sales increased 37 percent, from 2,713 billion kWh in 1990 to 3,723 billion kWh in 2017.

**Figure 3-8: Fuels Used in Electric Power Generation (TBtu) and Total Electric Power Sector CO<sub>2</sub> Emissions**



Electricity was used primarily in the residential, commercial, and industrial end-use sectors for lighting, heating, electric motors, appliances, electronics, and air conditioning (see Figure 3-9).

**Figure 3-9: Electric Power Retail Sales by End-Use Sector (Billion kWh)**



In 2017, electricity sales to the residential and commercial end-use sectors, as presented in Figure 3-9, decreased by 2.3 percent and 1.0 percent relative to 2016, respectively. Electricity sales to the industrial sector in 2017 increased approximately 0.8 percent relative to 2016. Overall, in 2017, the amount of electricity retail sales (in kWh) decreased by 1.0 percent relative to 2016.



## Industrial Sector

Industrial sector CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O, emissions accounted for 17, 15, and 6 percent of CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O, emissions from fossil fuel combustion, respectively. Carbon dioxide, CH<sub>4</sub>, and N<sub>2</sub>O emissions resulted from the direct consumption of fossil fuels for steam and process heat production.

The industrial end-use sector, per the underlying energy use data from EIA, includes activities such as manufacturing, construction, mining, and agriculture. The largest of these activities in terms of energy use is manufacturing, of which six industries—Petroleum Refineries, Chemicals, Paper, Primary Metals, Food, and Nonmetallic Mineral Products—represent the vast majority of the energy use (EIA 2019a, 2009b).

There are many dynamics that impact emissions from the industrial sector including economic activity, changes in the make-up of the industrial sector, changes in the emissions intensity of industrial processes, and weather impacts on heating of industrial buildings.<sup>19</sup> Structural changes within the U.S. economy that lead to shifts in industrial output away from energy-intensive manufacturing products to less energy-intensive products (e.g., from steel to computer equipment) have had a significant effect on industrial emissions.

From 2016 to 2017, total industrial production and manufacturing output increased by 1.6 percent (FRB 2019). Over this period, output increased across production indices for Food, Petroleum Refineries, Chemicals, and Nonmetallic Mineral Products, and decreased slightly for Primary Metals and Paper (see Figure 3-10). Through EPA's Greenhouse Gas Reporting Program (GHGRP), specific industrial sector trends can be discerned from the overall total EIA industrial fuel consumption data used for these calculations.

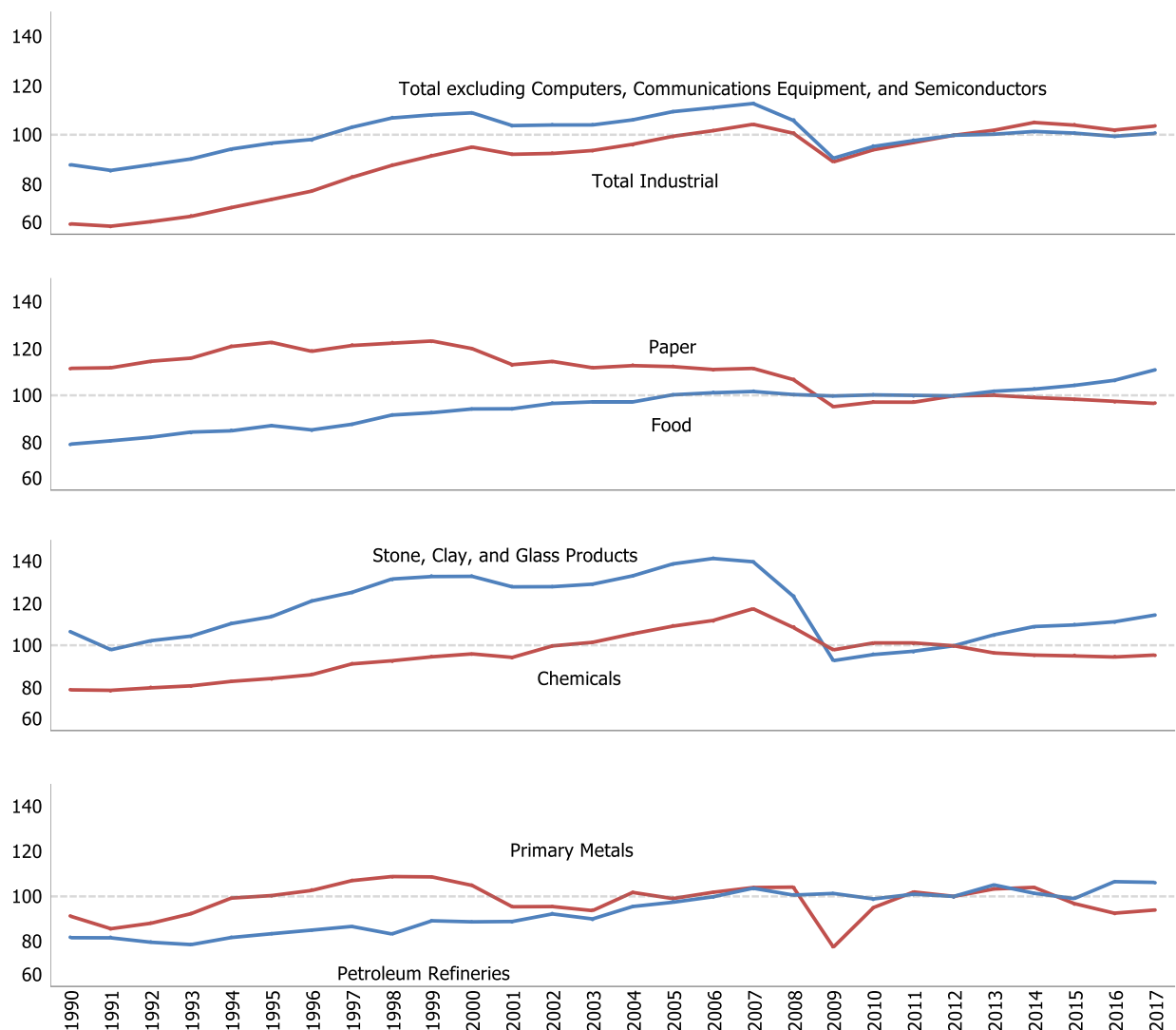
For example, from 2016 to 2017, the underlying EIA data showed decreased consumption of coal, and increase of natural gas in the industrial sector. The GHGRP data highlights that several industries contributed to these trends, including chemical manufacturing; pulp, paper and print; food processing, beverages and tobacco; minerals manufacturing; and agriculture-forest-fisheries.<sup>20</sup>

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<sup>19</sup> Some commercial customers are large enough to obtain an industrial price for natural gas and/or electricity and are consequently grouped with the industrial end-use sector in U.S. energy statistics. These misclassifications of large commercial customers likely cause the industrial end-use sector to appear to be more sensitive to weather conditions.

<sup>20</sup> Further details on industrial sector combustion emissions are provided by EPA's GHGRP. See <<http://ghgdata.epa.gov/ghgp/main.do>>.

**Figure 3-10: Industrial Production Indices (Index 2012=100)**

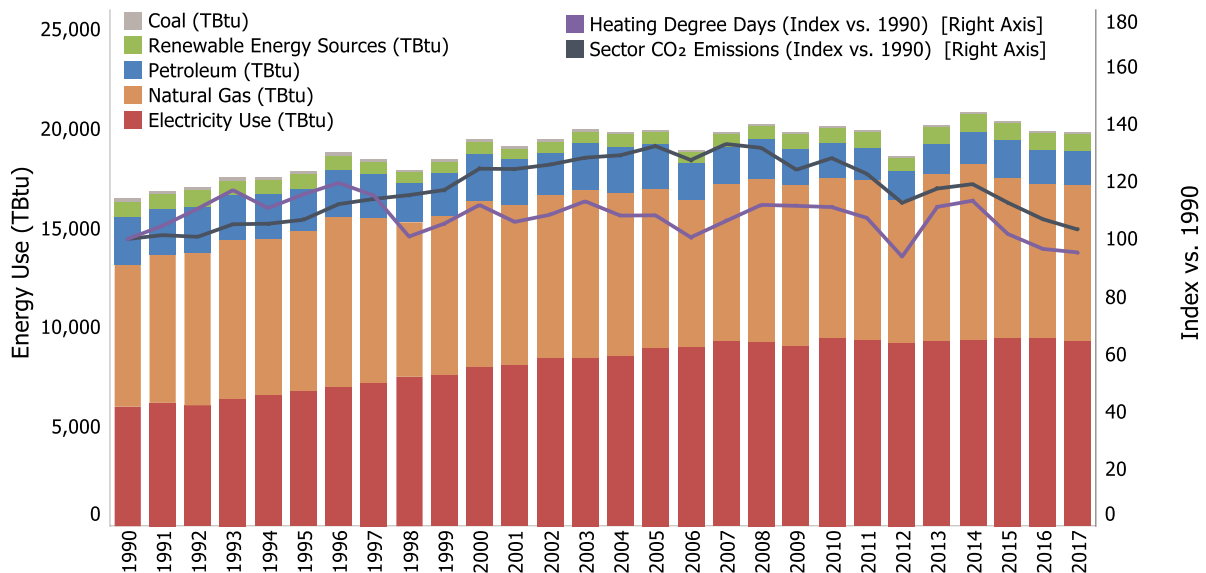


Despite the growth in industrial output (62 percent) and the overall U.S. economy (93 percent) from 1990 to 2017, CO<sub>2</sub> emissions from fossil fuel combustion in the industrial sector decreased by 5.5 percent over the same time series. A number of factors are believed to have caused this disparity between growth in industrial output and decrease in industrial emissions, including: (1) more rapid growth in output from less energy-intensive industries relative to traditional manufacturing industries, and (2) energy-intensive industries such as steel are employing new methods, such as electric arc furnaces, that are less carbon intensive than the older methods. In 2017, CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O emissions from fossil fuel combustion and electricity use within the industrial end-use sector totaled 1,326.9 MMT CO<sub>2</sub> Eq., a 0.8 percent decrease from 2016 emissions.

## Residential and Commercial Sectors

Emissions from the residential and commercial sectors have increased since 1990, and are often correlated with short-term fluctuations in energy use caused by weather conditions, rather than prevailing economic conditions. More significant changes in emissions from the residential and commercial sectors in recent years can be largely attributed to an overall reduction in energy use, a reduction in heating degree days, and increases in energy efficiency (see Figure 3-11).

**Figure 3-11: Fuels Used in Residential and Commercial Sectors (TBtu), Heating Degree Days, and Total Sector CO<sub>2</sub> Emissions**



In 2017 the residential and commercial sectors accounted for 6 and 5 percent of CO<sub>2</sub> emissions from fossil fuel combustion, respectively, 35 and 11 percent of CH<sub>4</sub> emissions from fossil fuel combustion, respectively, and 2 and 1 percent of N<sub>2</sub>O emissions from fossil fuel combustion, respectively. Emissions from these sectors were largely due to the direct consumption of natural gas and petroleum products, primarily for heating and cooking needs. Coal consumption was a minor component of energy use in both of these end-use sectors. In 2017, total emissions (CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O) from fossil fuel combustion and electricity use within the residential and commercial end-use sectors were 925.4 MMT CO<sub>2</sub> Eq. and 849.7 MMT CO<sub>2</sub> Eq., respectively. Total CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O emissions from combined fossil fuel combustion and electricity use within the residential and commercial end-use sectors decreased by 3.7 and 3.1 percent from 2016 to 2017, respectively, and heating degree days decreased by 1 percent over the same time period. A decrease in heating degree days impacted demand for heating fuel and electricity for heat in the residential and commercial sectors. In addition, a shift toward energy efficient products and more stringent energy efficiency standards for household equipment has also contributed to a decrease in energy demand in households (EIA 2018d), resulting in a decrease in energy-related emissions. In the long term, the residential sector is also affected by population growth, migration trends toward warmer areas, and changes in total housing units and building attributes (e.g., larger sizes and improved insulation).

In 2017, combustion emissions from natural gas consumption represented 82 and 74 percent of the direct fossil fuel CO<sub>2</sub> emissions from the residential and commercial sectors, respectively. Natural gas combustion CO<sub>2</sub> emissions from the residential and commercial sectors in 2017 increased by 1.3 percent and 1.5 percent from 2016 levels, respectively.

## U.S. Territories

Emissions from U.S. Territories are based on the fuel consumption in American Samoa, Guam, Puerto Rico, U.S. Virgin Islands, Wake Island, and other U.S. Pacific Islands. As described in the Methodology section of CO<sub>2</sub> from Fossil Fuel Combustion, this data is collected separately from the sectoral-level data available for the general calculations. As sectoral information is not available for U.S. Territories, CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O emissions are not presented for U.S. Territories in the tables above by sector, though the emissions will include some transportation and mobile combustion sources.

## Transportation Sector and Mobile Combustion

This discussion of transportation emissions follows the alternative method of presenting combustion emissions by allocating emissions associated with electricity generation to the transportation end-use sector, as presented in

Table 3-8. Table 3-7 presents direct CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O emissions from all transportation sources (i.e., excluding emissions allocated to electricity consumption in the transportation end-use sector).

The transportation end-use sector and other mobile combustion accounted for 1,825.1 MMT CO<sub>2</sub> Eq. in 2017, which represented 37 percent of CO<sub>2</sub> emissions, 29 percent of CH<sub>4</sub> emissions, and 37 percent of N<sub>2</sub>O emissions from fossil fuel combustion, respectively.<sup>21</sup> Fuel purchased in the United States for international aircraft and marine travel accounted for an additional 121.2 MMT CO<sub>2</sub> Eq. in 2017; these emissions are recorded as international bunkers and are not included in U.S. totals according to UNFCCC reporting protocols.

### *Transportation End-Use Sector*

From 1990 to 2017, transportation emissions from fossil fuel combustion rose by 20 percent due, in large part, to increased demand for travel (see Figure 3-12). The number of vehicle miles traveled (VMT) by light-duty motor vehicles (passenger cars and light-duty trucks) increased 45 percent from 1990 to 2017,<sup>22</sup> as a result of a confluence of factors including population growth, economic growth, urban sprawl, and periods of low fuel prices.

From 2016 to 2017, CO<sub>2</sub> emissions from the transportation end-use sector increased by 1.21 percent. The small increase in emissions is attributed to an increase in diesel fuel consumption by medium- and heavy-duty trucks and jet fuel consumption by commercial aircraft.

Commercial aircraft emissions increased between 2016 and 2017, but have decreased 8 percent since 2007 (FAA 2019).<sup>23</sup> Decreases in jet fuel emissions (excluding bunkers) since 2007 are due in part to improved operational efficiency that results in more direct flight routing, improvements in aircraft and engine technologies to reduce fuel burn and emissions, and the accelerated retirement of older, less fuel-efficient aircraft.

Almost all of the energy consumed for transportation was supplied by petroleum-based products, with more than half being related to gasoline consumption in automobiles and other highway vehicles. Other fuel uses, especially diesel fuel for freight trucks and jet fuel for aircraft, accounted for the remainder. The primary driver of transportation-related emissions was CO<sub>2</sub> from fossil fuel combustion, which increased by 23 percent from 1990 to 2017. Annex 3.2 presents the total emissions from all transportation and mobile sources, including CO<sub>2</sub>, N<sub>2</sub>O, CH<sub>4</sub>, and HFCs.

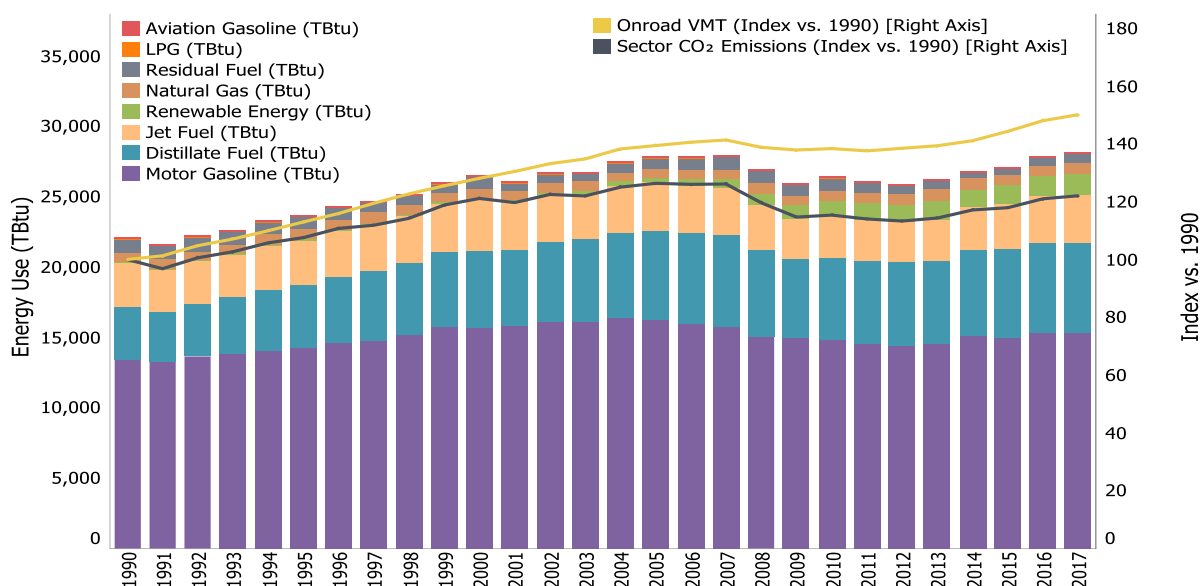
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<sup>21</sup> Note that these totals include CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O emissions from some sources in the U.S. Territories (ships and boats, recreational boats, non-transportation mobile sources) and CH<sub>4</sub> and N<sub>2</sub>O emissions from transportation rail electricity.

<sup>22</sup> VMT estimates are based on data from FHWA Highway Statistics Table VM-1 (FHWA 1996 through 2017). In 2011, FHWA changed its methods for estimating VMT by vehicle class, which led to a shift in VMT and emissions among on-road vehicle classes in the 2007 to 2017 time period. In absence of these method changes, light-duty VMT growth between 1990 and 2017 would likely have been even higher.

<sup>23</sup> Commercial aircraft, as modeled in FAA's AEDT (FAA 2019), consists of passenger aircraft, cargo, and other chartered flights.

**Figure 3-12: Fuels Used in Transportation Sector (TBtu), Onroad VMT, and Total Sector CO<sub>2</sub> Emissions**



Notes: Distillate fuel, residual fuel, and jet fuel include adjustments for international bunker fuels. Distillate fuel and motor gasoline include adjustments for the sectoral allocation of these fuels.

Source: Information on fuel consumption was obtained from EIA (2019a).

### *Transportation Fossil Fuel Combustion CO<sub>2</sub> Emissions*

Domestic transportation CO<sub>2</sub> emissions increased by 23 percent (332.8 MMT CO<sub>2</sub>) between 1990 and 2017, an annualized increase of 0.8 percent. Among domestic transportation sources in 2017, light-duty vehicles (including passenger cars and light-duty trucks) represented 58 percent of CO<sub>2</sub> emissions from fossil fuel combustion, medium- and heavy-duty trucks and buses 25 percent, commercial aircraft 7 percent, and other sources 10 percent. See Table 3-13 for a detailed breakdown of transportation CO<sub>2</sub> emissions by mode and fuel type.

Almost all of the energy consumed by the transportation sector is petroleum-based, including motor gasoline, diesel fuel, jet fuel, and residual oil. Carbon dioxide emissions from the combustion of ethanol and biodiesel for transportation purposes, along with the emissions associated with the agricultural and industrial processes involved in the production of biofuel, are captured in other Inventory sectors.<sup>24</sup> Ethanol consumption from the transportation sector has increased from 0.7 billion gallons in 1990 to 13.4 billion gallons in 2017, while biodiesel consumption has increased from 0.01 billion gallons in 2001 to 2.0 billion gallons in 2017. For further information, see Section 3.11 on biofuel consumption at the end of this chapter and Table A-98 in Annex 3.2.

Carbon dioxide emissions from passenger cars and light-duty trucks totaled 1,055.4 MMT CO<sub>2</sub> in 2017. This is an increase of 14 percent (130.9 MMT CO<sub>2</sub>) from 1990 due, in large part, to increased demand for travel as fleet-wide light-duty vehicle fuel economy was relatively stable (average new vehicle fuel economy declined slowly from 1990 through 2004 and then increased more rapidly from 2005 through 2017). Carbon dioxide emissions from passenger cars and light-duty trucks peaked at 1,151.5 MMT CO<sub>2</sub> in 2004, and since then have declined about 8 percent. The decline in new light-duty vehicle fuel economy between 1990 and 2004 (Figure 3-13) reflected the increasing

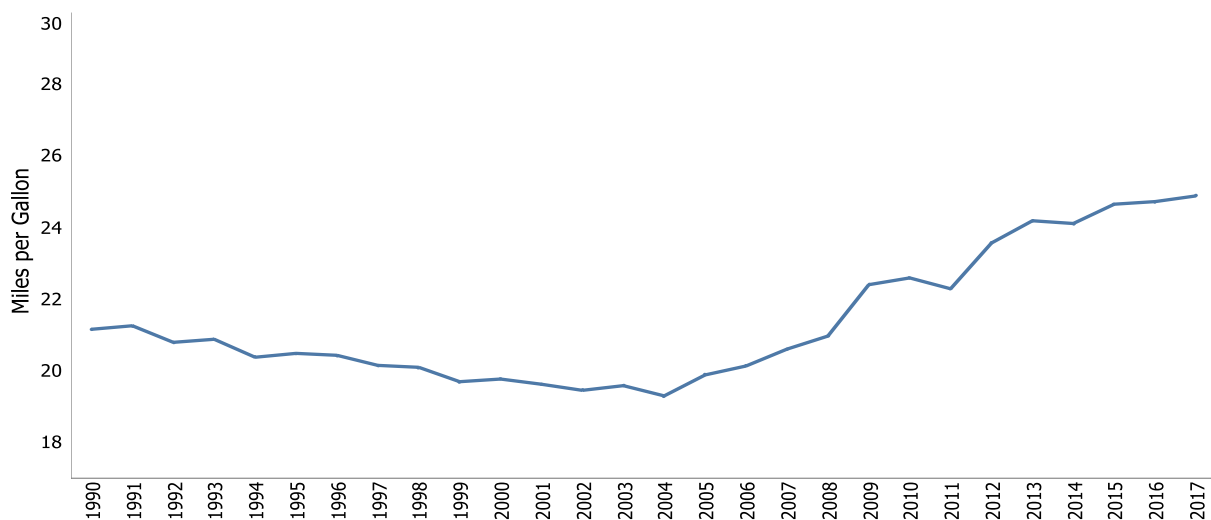
<sup>24</sup> Biofuel estimates are presented in the Energy chapter for informational purposes only, in line with IPCC methodological guidance and UNFCCC reporting obligations. Net carbon fluxes from changes in biogenic carbon reservoirs in croplands are accounted for in the estimates for Land Use, Land-Use Change, and Forestry (see Chapter 6). More information and additional analyses on biofuels are available at EPA's Renewable Fuels Standards website. See <<https://www.epa.gov/renewable-fuel-standard-program>>.

market share of light-duty trucks, which grew from about 30 percent of new vehicle sales in 1990 to 48 percent in 2004. Starting in 2005, average new vehicle fuel economy began to increase while light-duty VMT grew only modestly for much of the period. Light-duty VMT grew by less than one percent or declined each year between 2005 and 2013<sup>25</sup> and has since grown at a faster rate (2.5 percent from 2015 to 2016, and 1.0 percent from 2016 to 2017). Average new vehicle fuel economy has increased almost every year since 2005, while the light-duty truck share decreased to about 33 percent in 2009 and has since varied from year to year between 36 and 45 percent. Light-duty truck share is about 42 percent of new vehicles in model year 2017 (EPA 2018a). See also Annex 3.2 for data by vehicle mode and information on VMT and the share of new vehicles (in VMT).

Medium- and heavy-duty truck CO<sub>2</sub> emissions increased by 88 percent from 1990 to 2017. This increase was largely due to a substantial growth in medium- and heavy-duty truck VMT, which increased by 107 percent between 1990 and 2017.<sup>26</sup> Carbon dioxide from the domestic operation of commercial aircraft increased by 17 percent (18.1 MMT CO<sub>2</sub>) from 1990 to 2017.<sup>27</sup> Across all categories of aviation, excluding international bunkers, CO<sub>2</sub> emissions decreased by 8 percent (14.2 MMT CO<sub>2</sub>) between 1990 and 2017.<sup>28</sup> This includes a 65 percent (22.8 MMT CO<sub>2</sub>) decrease in CO<sub>2</sub> emissions from domestic military operations.

Transportation sources also produce CH<sub>4</sub> and N<sub>2</sub>O; these emissions are included in Table 3-14 and Table 3-15 and in the CH<sub>4</sub> and N<sub>2</sub>O from Mobile Combustion section. Annex 3.2 presents total emissions from all transportation and mobile sources, including CO<sub>2</sub>, CH<sub>4</sub>, N<sub>2</sub>O, and HFCs.

**Figure 3-13: Sales-Weighted Fuel Economy of New Passenger Cars and Light-Duty Trucks, 1990–2017 (miles/gallon)**



Source: EPA (2018a).

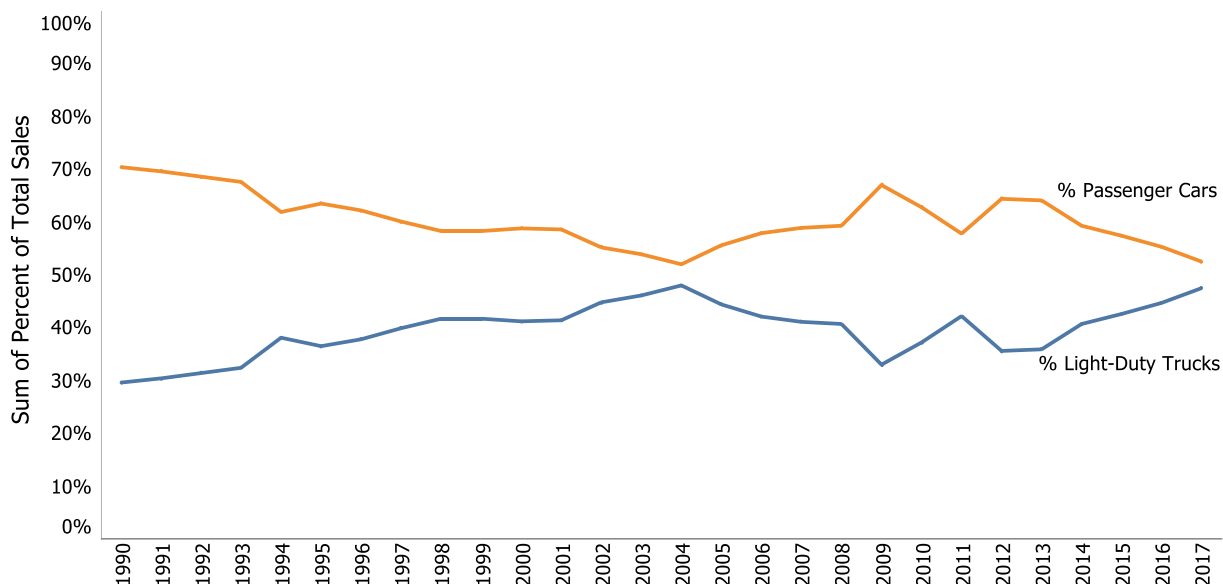
<sup>25</sup> VMT estimates are based on data from FHWA Highway Statistics Table VM-1 (FHWA 1996 through 2017). In 2007 and 2008 light-duty VMT decreased 3.0 percent and 2.3 percent, respectively. Note that the decline in light-duty VMT from 2006 to 2007 is due at least in part to a change in FHWA's methods for estimating VMT. In 2011, FHWA changed its methods for estimating VMT by vehicle class, which led to a shift in VMT and emissions among on-road vehicle classes in the 2007 to 2017 time period. In absence of these method changes, light-duty VMT growth between 2006 and 2007 would likely have been higher.

<sup>26</sup> While FHWA data shows consistent growth in medium- and heavy-duty truck VMT over the 1990 to 2017 time period, part of the growth reflects a method change for estimating VMT starting in 2007. This change in methodology in FHWA's VM-1 table resulted in large changes in VMT by vehicle class, thus leading to a shift in VMT and emissions among on-road vehicle classes in the 2007 to 2017 time period. During the time period prior to the method change (1990 to 2006), VMT for medium- and heavy-duty trucks increased by 51 percent.

<sup>27</sup> Commercial aircraft, as modeled in FAA's AEDT, consists of passenger aircraft, cargo, and other chartered flights.

<sup>28</sup> Includes consumption of jet fuel and aviation gasoline. Does not include aircraft bunkers, which are not included in national emission totals, in line with IPCC methodological guidance and UNFCCC reporting obligations.

**Figure 3-14: Sales of New Passenger Cars and Light-Duty Trucks, 1990–2017 (Percent)**



Source: EPA (2018a).

**Table 3-13: CO<sub>2</sub> Emissions from Fossil Fuel Combustion in Transportation End-Use Sector (MMT CO<sub>2</sub> Eq.)**

Fuel/Vehicle Type	1990	2005	2013 <sup>a</sup>	2014 <sup>a</sup>	2015 <sup>a</sup>	2016 <sup>a</sup>	2017 <sup>a</sup>
<b>Gasoline<sup>b</sup></b>	<b>958.9</b>	<b>1,153.6</b>	<b>1,037.4</b>	<b>1,077.4</b>	<b>1,070.0</b>	<b>1,095.3</b>	<b>1,092.3</b>
Passenger Cars	604.3	639.1	712.9	730.2	732.0	744.9	745.3
Light-Duty Trucks	300.6	464.9	270.6	292.2	283.5	294.6	290.3
Medium- and Heavy-Duty Trucks <sup>c</sup>	37.7	33.9	38.5	39.8	39.3	40.4	41.3
Buses	0.3	0.4	0.8	0.9	0.9	0.9	0.9
Motorcycles	1.7	1.6	3.8	3.8	3.7	3.8	3.7
Recreational Boats <sup>d</sup>	14.3	13.8	10.8	10.6	10.6	10.7	10.7
<b>Distillate Fuel Oil (Diesel)<sup>b</sup></b>	<b>262.9</b>	<b>457.5</b>	<b>433.9</b>	<b>447.7</b>	<b>460.7</b>	<b>462.7</b>	<b>475.8</b>
Passenger Cars	7.9	4.2	4.1	4.1	4.3	4.3	4.4
Light-Duty Trucks	11.5	25.8	12.9	13.9	13.9	14.3	14.3
Medium- and Heavy-Duty Trucks <sup>c</sup>	190.5	360.2	350.0	361.2	369.4	376.4	388.3
Buses	8.0	10.6	15.5	16.9	17.3	17.0	17.9
Rail	35.5	45.5	40.1	41.6	39.9	36.7	37.9
Recreational Boats <sup>d</sup>	2.7	2.8	2.6	2.5	2.7	2.8	2.8
Ships and Non-Recreational Boats <sup>e</sup>	6.8	8.4	8.7	7.5	13.3	11.1	10.2
International Bunker Fuels <sup>f</sup>	11.7	9.4	5.6	6.1	8.4	8.7	9.0
<b>Jet Fuel</b>	<b>184.2</b>	<b>189.3</b>	<b>147.1</b>	<b>148.4</b>	<b>157.6</b>	<b>166.0</b>	<b>171.8</b>
Commercial Aircraft <sup>g</sup>	109.9	132.7	114.3	115.2	119.0	120.4	128.0
Military Aircraft	35.0	19.4	11.0	14.0	13.5	12.3	12.2
General Aviation Aircraft	39.4	37.3	21.8	19.2	25.1	33.4	31.5
International Bunker Fuels <sup>f</sup>	38.0	60.1	65.7	69.6	71.9	74.1	77.7
International Bunker Fuels from Commercial Aviation	30.0	55.6	62.8	66.3	68.6	70.8	74.5

<b>Aviation Gasoline</b>	<b>3.1</b>	<b>2.4</b>	<b>1.5</b>	<b>1.5</b>	<b>1.5</b>	<b>1.4</b>	<b>1.4</b>
General Aviation Aircraft	3.1	2.4	1.5	1.5	1.5	1.4	1.4
<b>Residual Fuel Oil</b>	<b>22.6</b>	<b>19.3</b>	<b>15.1</b>	<b>5.8</b>	<b>4.2</b>	<b>12.9</b>	<b>16.5</b>
Ships and Boats <sup>e</sup>	22.6	19.3	15.1	5.8	4.2	12.9	16.5
<i>International Bunker Fuels<sup>f</sup></i>	53.7	43.6	28.5	27.7	30.6	33.8	33.4
<b>Natural Gas<sup>j</sup></b>	<b>36.0</b>	<b>33.1</b>	<b>47.0</b>	<b>40.2</b>	<b>39.4</b>	<b>40.1</b>	<b>42.3</b>
Passenger Cars	+	+	+	+	+	+	+
Light-Duty Trucks	+	+	+	+	+	+	+
Medium- and Heavy-Duty Trucks	+	+	+	+	+	+	+
Buses	+	0.6	0.8	0.8	0.9	0.8	0.8
Pipeline <sup>h</sup>	36.0	32.4	46.2	39.4	38.5	39.2	41.4
<b>LPG<sup>j</sup></b>	<b>1.4</b>	<b>1.7</b>	<b>0.6</b>	<b>0.6</b>	<b>0.6</b>	<b>0.6</b>	<b>0.6</b>
Passenger Cars	+	+	+	+	0.1	+	+
Light-Duty Trucks	0.2	0.3	0.1	0.1	0.1	0.1	0.1
Medium- and Heavy-Duty Trucks <sup>e</sup>	1.1	1.3	0.4	0.4	0.4	0.4	0.4
Buses	0.1	0.1	0.1	0.1	0.1	0.1	0.1
<b>Electricity<sup>l</sup></b>	<b>3.0</b>	<b>4.7</b>	<b>4.3</b>	<b>4.5</b>	<b>4.3</b>	<b>4.2</b>	<b>4.3</b>
Passenger Cars	+	+	0.2	0.4	0.5	0.7	0.8
Light-Duty Trucks	+	+	+	+	+	0.1	0.1
Buses	+	+	+	+	+	+	+
Rail	3.0	4.7	4.0	4.0	3.7	3.5	3.4
<b>Total<sup>k</sup></b>	<b>1,472.1</b>	<b>1,861.7</b>	<b>1,686.9</b>	<b>1,726.0</b>	<b>1,738.2</b>	<b>1,783.2</b>	<b>1,804.9</b>
<b>Total (Including Bunkers)<sup>f</sup></b>	<b>1,575.6</b>	<b>1,974.9</b>	<b>1,786.7</b>	<b>1,829.4</b>	<b>1,849.1</b>	<b>1,899.8</b>	<b>1,925.0</b>
<i>Biofuels-Ethanol<sup>i</sup></i>	4.1	21.6	70.5	74.0	74.2	76.9	77.7
<i>Biofuels-Biodiesel<sup>i</sup></i>	+	0.9	13.5	13.3	14.1	19.6	18.7

+ Does not exceed 0.05 MMT CO<sub>2</sub> Eq.

<sup>a</sup> In 2011, FHWA changed its methods for estimating vehicle miles traveled (VMT) and related data. These methodological changes included how vehicles are classified, moving from a system based on body-type to one that is based on wheelbase. These changes were first incorporated for the 1990 through 2010 Inventory and apply to the 2007 through 2017 time period. This resulted in large changes in VMT and fuel consumption data by vehicle class, thus leading to a shift in emissions among on-road vehicle classes.

<sup>b</sup> Gasoline and diesel highway vehicle fuel consumption estimates are based on data from FHWA Highway Statistics Table MF-21, MF-27, and VM-1 (FHWA 1996 through 2017). Table VM-1 fuel consumption data for 2017 has not been published yet, therefore 2017 fuel consumption data is estimated using the percent change in VMT from 2016 to 2017. Data from Table VM-1 is used to estimate the share of consumption between each on-road vehicle class. These fuel consumption estimates are combined with estimates of fuel shares by vehicle type from DOE's TEDB Annex Tables A.1 through A.6 (DOE 1993 through 2017). TEDB data for 2017 has not been published yet, therefore 2016 data are used as a proxy.

<sup>c</sup> Includes medium- and heavy-duty trucks over 8,500 lbs.

<sup>d</sup> In 2014, EPA incorporated the NONROAD2008 model into MOVES2014. The current Inventory uses the NONROAD component of MOVES2014b for years 1999 through 2017.

<sup>e</sup> Note that large year over year fluctuations in emission estimates partially reflect nature of data collection for these sources.

<sup>f</sup> Official estimates exclude emissions from the combustion of both aviation and marine international bunker fuels; however, estimates including international bunker fuel-related emissions are presented for informational purposes.

<sup>g</sup> Commercial aircraft, as modeled in FAA's Aviation Environmental Design Tool (AEDT), consists of passenger aircraft, cargo, and other chartered flights.

<sup>h</sup> Pipelines reflect CO<sub>2</sub> emissions from natural gas-powered pipelines transporting natural gas.

<sup>i</sup> Ethanol and biodiesel estimates are presented for informational purposes only. See Section 3.11 of this chapter and the estimates in Land Use, Land-Use Change, and Forestry (see Chapter 6), in line with IPCC methodological guidance and UNFCCC reporting obligations, for more information on ethanol and biodiesel.

<sup>j</sup> Transportation sector natural gas and LPG consumption are based on data from EIA (2019b). Prior to the 1990 to 2015 Inventory, data from DOE TEDB were used to estimate each vehicle class's share of the total natural gas and LPG consumption. Since TEDB does not include estimates for natural gas use by medium and heavy-duty trucks or LPG use by passenger cars, EIA Alternative Fuel Vehicle Data (Browning 2017) is now used to determine each vehicle class's share of the total natural gas and LPG consumption. These changes were first incorporated in the 1990-2016 Inventory and apply to the 1990 to 2017 time period.

<sup>k</sup> Includes emissions from rail electricity.

<sup>l</sup> Electricity consumption by passenger cars, light-duty trucks (SUVs), and buses is based on plug-in electric vehicle sales and engine efficiency data, as outlined in Browning (2018a). In prior Inventory years, CO<sub>2</sub> emissions from electric vehicle



charging were allocated to the residential and commercial sectors. They are now allocated to the transportation sector. These changes were first incorporated in this year's Inventory and apply to the 2010 through 2017 time period.

Notes: This table does not include emissions from non-transportation mobile sources, such as agricultural equipment and construction/mining equipment; it also does not include emissions associated with electricity consumption by pipelines or lubricants used in transportation. In addition, this table does not include CO<sub>2</sub> emissions from U.S. Territories, since these are covered in a separate chapter of the Inventory. Totals may not sum due to independent rounding.

### *Mobile Fossil Fuel Combustion CH<sub>4</sub> and N<sub>2</sub>O Emissions*

Mobile combustion includes emissions of CH<sub>4</sub> and N<sub>2</sub>O from all transportation sources identified in the U.S. Inventory with the exception of pipelines and electric locomotives;<sup>29</sup> mobile sources also include non-transportation sources such as construction/mining equipment, agricultural equipment, vehicles used off-road, and other sources (e.g., snowmobiles, lawnmowers, etc.).<sup>30</sup> Annex 3.2 includes a summary of all emissions from both transportation and mobile sources. Table 3-14 and Table 3-15 provide mobile fossil fuel CH<sub>4</sub> and N<sub>2</sub>O emission estimates in MMT CO<sub>2</sub> Eq.<sup>31</sup>

Mobile combustion was responsible for a small portion of national CH<sub>4</sub> emissions (0.5 percent) but was the fourth largest source of U.S. N<sub>2</sub>O emissions (4.7 percent). From 1990 to 2017, mobile source CH<sub>4</sub> emissions declined by 75 percent, to 3.2 MMT CO<sub>2</sub> Eq. (128 kt CH<sub>4</sub>), due largely to control technologies employed in on-road vehicles since the mid-1990s to reduce CO, NO<sub>x</sub>, NMVOC, and CH<sub>4</sub> emissions. Mobile source emissions of N<sub>2</sub>O decreased by 60 percent, to 16.9 MMT CO<sub>2</sub> Eq. (57 kt N<sub>2</sub>O). Earlier generation control technologies initially resulted in higher N<sub>2</sub>O emissions, causing a 30 percent increase in N<sub>2</sub>O emissions from mobile sources between 1990 and 1997. Improvements in later-generation emission control technologies have reduced N<sub>2</sub>O output, resulting in a 69 percent decrease in mobile source N<sub>2</sub>O emissions from 1997 to 2017 (Figure 3-15). Overall, CH<sub>4</sub> and N<sub>2</sub>O emissions were predominantly from gasoline-fueled passenger cars and light-duty trucks. See also Annex 3.2 for data by vehicle mode and information on VMT and the share of new vehicles (in VMT).

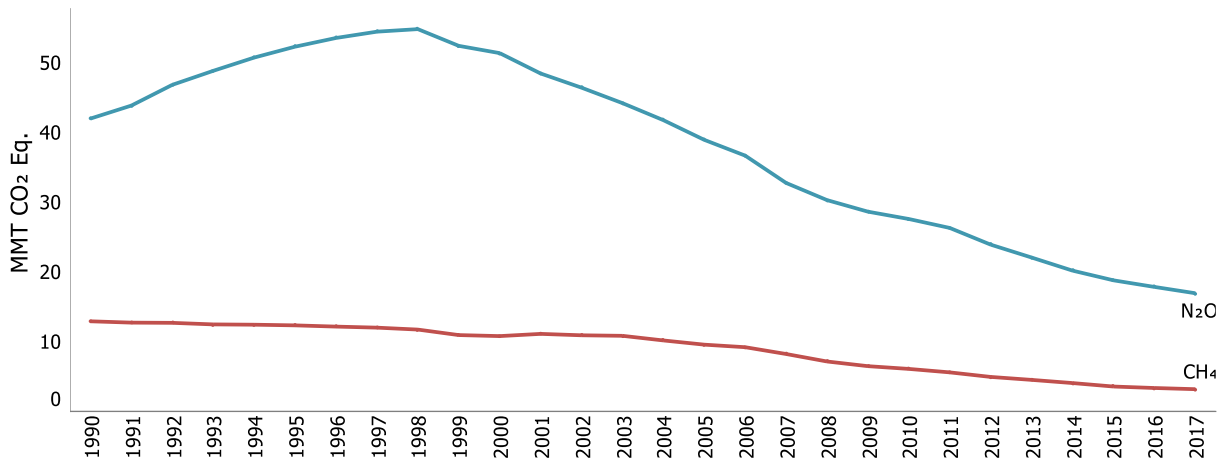
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<sup>29</sup> Emissions of CH<sub>4</sub> from natural gas systems are reported separately. More information on the methodology used to calculate these emissions are included in this chapter and Annex 3.4.

<sup>30</sup> See the methodology sub-sections of the CO<sub>2</sub> from Fossil Fuel Combustion and CH<sub>4</sub> and N<sub>2</sub>O from Mobile Combustion sections of this chapter. Note that N<sub>2</sub>O and CH<sub>4</sub> emissions are reported using different categories than CO<sub>2</sub>. CO<sub>2</sub> emissions are reported by end-use sector (Transportation, Industrial, Commercial, Residential, U.S. Territories), and generally adhere to a top-down approach to estimating emissions. CO<sub>2</sub> emissions from non-transportation sources (e.g., lawn and garden equipment, farm equipment, construction equipment) are allocated to their respective end-use sector (i.e., construction equipment CO<sub>2</sub> emissions are included in the Industrial end-use sector instead of the Transportation end-use sector). CH<sub>4</sub> and N<sub>2</sub>O emissions are reported using the "Mobile Combustion" category, which includes non-transportation mobile sources. CH<sub>4</sub> and N<sub>2</sub>O emission estimates are bottom-up estimates, based on total activity (fuel use, VMT) and emissions factors by source and technology type. These reporting schemes are in accordance with IPCC guidance. For informational purposes only, CO<sub>2</sub> emissions from non-transportation mobile sources are presented separately from their overall end-use sector in Annex 3.2.

<sup>31</sup> See Annex 3.2 for a complete time series of emission estimates for 1990 through 2017.

**Figure 3-15: Mobile Source CH<sub>4</sub> and N<sub>2</sub>O Emissions (MMT CO<sub>2</sub> Eq.)**



**Table 3-14: CH<sub>4</sub> Emissions from Mobile Combustion (MMT CO<sub>2</sub> Eq.)**

Fuel Type/Vehicle Type <sup>a</sup>	1990	2005	2013	2014	2015	2016	2017
<b>Gasoline On-Road<sup>b</sup></b>	<b>5.2</b>	<b>2.2</b>	<b>1.1</b>	<b>1.0</b>	<b>0.9</b>	<b>0.9</b>	<b>0.8</b>
Passenger Cars	3.2	1.3	0.8	0.7	0.6	0.6	0.5
Light-Duty Trucks	1.7	0.8	0.3	0.2	0.2	0.2	0.2
Medium- and Heavy-Duty Trucks and Buses	0.3	0.1	0.1	+	+	+	+
Motorcycles	+	+	+	+	+	+	+
<b>Diesel On-Road<sup>b</sup></b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>
Passenger Cars	+	+	+	+	+	+	+
Light-Duty Trucks	+	+	+	+	+	+	+
Medium- and Heavy-Duty Trucks and Buses	+	+	+	+	+	+	+
<b>Alternative Fuel On-Road</b>	<b>+</b>	<b>0.2</b>	<b>0.2</b>	<b>0.2</b>	<b>0.2</b>	<b>0.2</b>	<b>0.2</b>
<b>Non-Road<sup>c</sup></b>	<b>7.7</b>	<b>7.2</b>	<b>3.1</b>	<b>2.8</b>	<b>2.4</b>	<b>2.3</b>	<b>2.2</b>
Ships and Boats	0.6	0.5	0.3	0.3	0.3	0.3	0.3
Rail	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Aircraft	0.1	0.1	+	+	+	+	+
Agricultural Equipment <sup>d</sup>	0.6	0.6	0.2	0.2	0.1	0.1	0.1
Construction/Mining Equipment <sup>e</sup>	0.9	1.0	0.7	0.6	0.5	0.4	0.4
Other <sup>f</sup>	5.5	4.9	1.8	1.6	1.5	1.4	1.3
<b>Total</b>	<b>12.9</b>	<b>9.6</b>	<b>4.5</b>	<b>4.1</b>	<b>3.6</b>	<b>3.4</b>	<b>3.2</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub> Eq.

<sup>a</sup> See Annex 3.2 for definitions of on-road vehicle types.

<sup>b</sup> Gasoline and diesel highway vehicle mileage estimates are based on data from FHWA Highway Statistics Table VM-1 (FHWA 1996 through 2017). Table VM-1 fuel consumption data for 2017 has not been published yet, therefore 2017 fuel consumption data is estimated using the percent change in VMT from 2016 to 2017. These mileage estimates are combined with estimates of fuel shares by vehicle type from DOE's TEDB Annex Tables A.1 through A.6 (DOE 1993 through 2017). TEDB data for 2017 has not been published yet, therefore 2016 data are used as a proxy.

<sup>c</sup> Rail emissions do not include emissions from electric powered locomotives. Class II and Class III rail diesel consumption for 2014-2017 are not available, therefore 2013 data is used as a proxy. Commuter and intercity rail diesel consumption data for 2017 is not available yet, therefore 2016 data are used as a proxy.

<sup>d</sup> Includes equipment, such as tractors and combines, as well as fuel consumption from trucks that are used off-road in agriculture.

<sup>e</sup> Includes equipment, such as cranes, dumpers, and excavators, as well as fuel consumption from trucks that are used off-road in construction.

<sup>f</sup> “Other” includes snowmobiles and other recreational equipment, logging equipment, lawn and garden equipment, railroad equipment, airport equipment, commercial equipment, and industrial equipment, as well as fuel consumption from trucks that are used off-road for commercial/industrial purposes.

Notes: In 2011, FHWA changed its methods for estimating vehicle miles traveled (VMT) and related data. These methodological changes included how vehicles are classified, moving from a system based on body-type to one that is based on wheelbase. These changes were first incorporated for the 1990 through 2010 Inventory and apply to the 2007 through 2017 time period. This resulted in large changes in VMT and fuel consumption data by vehicle class, thus leading to a shift in emissions among on-road vehicle classes. Totals may not sum due to independent rounding.

**Table 3-15: N<sub>2</sub>O Emissions from Mobile Combustion (MMT CO<sub>2</sub> Eq.)**

Fuel Type/Vehicle Type <sup>a</sup>	1990	2005	2013	2014	2015	2016	2017
<b>Gasoline On-Road<sup>b</sup></b>	<b>37.5</b>	<b>33.5</b>	<b>17.2</b>	<b>15.4</b>	<b>14.0</b>	<b>12.8</b>	<b>11.7</b>
Passenger Cars	24.1	17.5	11.8	10.5	9.7	8.9	8.2
Light-Duty Trucks	12.8	15.0	4.7	4.4	3.8	3.5	3.1
Medium- and Heavy-Duty Trucks and Buses	0.5	0.9	0.6	0.5	0.5	0.4	0.4
Motorcycles	+	+	+	+	+	+	+
<b>Diesel On-Road<sup>b</sup></b>	<b>0.2</b>	<b>0.3</b>	<b>0.4</b>	<b>0.4</b>	<b>0.4</b>	<b>0.4</b>	<b>0.4</b>
Passenger Cars	+	+	+	+	+	+	+
Light-Duty Trucks	+	+	+	+	+	+	+
Medium- and Heavy-Duty Trucks and Buses	0.2	0.3	0.4	0.4	0.4	0.4	0.4
<b>Alternative Fuel On-Road</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>
<b>Non-Road</b>	<b>4.4</b>	<b>5.2</b>	<b>4.6</b>	<b>4.5</b>	<b>4.5</b>	<b>4.7</b>	<b>4.9</b>
Ships and Boats	0.6	0.6	0.5	0.3	0.4	0.5	0.5
Rail <sup>c</sup>	0.3	0.4	0.4	0.4	0.4	0.3	0.3
Aircraft	1.7	1.8	1.4	1.4	1.5	1.5	1.6
Agricultural Equipment <sup>d</sup>	0.5	0.6	0.6	0.6	0.6	0.6	0.5
Construction/Mining Equipment <sup>e</sup>	0.6	1.0	0.8	0.8	0.8	0.8	0.9
Other <sup>f</sup>	0.6	0.9	0.9	1.0	1.0	1.0	1.0
<b>Total</b>	<b>42.1</b>	<b>39.0</b>	<b>22.1</b>	<b>20.3</b>	<b>18.9</b>	<b>17.9</b>	<b>17.0</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub> Eq.

<sup>a</sup> See Annex 3.2 for definitions of on-road vehicle types.

<sup>b</sup> Gasoline and diesel highway vehicle mileage estimates are based on data from FHWA Highway Statistics Table VM-1 (FHWA 1996 through 2017). Table VM-1 fuel consumption data for 2017 has not been published yet, therefore 2017 fuel consumption data is estimated using the percent change in VMT from 2016 to 2017. These mileage estimates are combined with estimates of fuel shares by vehicle type from DOE’s TEDB Annex Tables A.1 through A.6 (DOE 1993 through 2017). TEDB data for 2017 has not been published yet, therefore 2016 data are used as a proxy.

<sup>c</sup> Rail emissions do not include emissions from electric powered locomotives. Class II and Class III rail diesel consumption for 2014-2017 are not available, therefore 2013 data is used as a proxy. Commuter and intercity rail diesel consumption data for 2017 is not available yet, therefore 2016 data are used as a proxy.

<sup>d</sup> Includes equipment, such as tractors and combines, as well as fuel consumption from trucks that are used off-road in agriculture.

<sup>e</sup> Includes equipment, such as cranes, dumpers, and excavators, as well as fuel consumption from trucks that are used off-road in construction.

<sup>f</sup> “Other” includes snowmobiles and other recreational equipment, logging equipment, lawn and garden equipment, railroad equipment, airport equipment, commercial equipment, and industrial equipment, as well as fuel consumption from trucks that are used off-road for commercial/industrial purposes.

Note: In 2011, FHWA changed its methods for estimating vehicle miles traveled (VMT) and related data. These methodological changes included how vehicles are classified, moving from a system based on body type to one that is based on wheelbase. These changes were first incorporated for the 1990 through 2010 Inventory and apply to the 2007 through 2017 time period. This resulted in large changes in VMT and fuel consumption data by vehicle class, thus leading to a shift in emissions among on-road vehicle classes. Totals may not sum due to independent rounding.

# CO<sub>2</sub> from Fossil Fuel Combustion

## Methodology

CO<sub>2</sub> emissions from fossil fuel combustion are estimated in line with a Tier 2 method described by the IPCC in the *2006 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC 2006) with some exceptions as discussed below.<sup>32</sup> A detailed description of the U.S. methodology is presented in Annex 2.1, and is characterized by the following steps:

1. *Determine total fuel consumption by fuel type and sector.* Total fossil fuel consumption for each year is estimated by aggregating consumption data by end-use sector (e.g., commercial, industrial, etc.), primary fuel type (e.g., coal, petroleum, gas), and secondary fuel category (e.g., motor gasoline, distillate fuel oil, etc.). Fuel consumption data for the United States were obtained directly from the EIA of the U.S. Department of Energy (DOE), primarily from the *Monthly Energy Review* (EIA 2019a). The EIA does not include territories in its national energy statistics, so fuel consumption data for territories were collected separately from EIA's International Energy Statistics (EIA 2017).<sup>33</sup>

For consistency of reporting, the IPCC has recommended that countries report energy data using the International Energy Agency (IEA) reporting convention and/or IEA data. Data in the IEA format are presented “top down”—that is, energy consumption for fuel types and categories are estimated from energy production data (accounting for imports, exports, stock changes, and losses). The resulting quantities are referred to as “apparent consumption.” The data collected in the United States by EIA on an annual basis and used in this Inventory are predominantly from mid-stream or conversion energy consumers such as refiners and electric power generators. These annual surveys are supplemented with end-use energy consumption surveys, such as the Manufacturing Energy Consumption Survey, that are conducted on a periodic basis (every four years). These consumption data sets help inform the annual surveys to arrive at the national total and sectoral breakdowns for that total.<sup>34</sup>

Also, note that U.S. fossil fuel energy statistics are generally presented using gross calorific values (GCV) (i.e., higher heating values). Fuel consumption activity data presented here have not been adjusted to correspond to international standards, which are to report energy statistics in terms of net calorific values (NCV) (i.e., lower heating values).<sup>35</sup>

2. *Subtract uses accounted for in the Industrial Processes and Product Use chapter.* Portions of the fuel consumption data for seven fuel categories—coking coal, distillate fuel, industrial other coal, petroleum coke, natural gas, residual fuel oil, and other oil—were reallocated to the Industrial Processes and Product Use chapter, as they were consumed during non-energy-related industrial activity. To make these adjustments, additional data were collected from AISI (2004 through 2018), Coffeyville (2012), U.S. Census Bureau (2001 through 2011), EIA (2019a, 2019b, 2018a), USAA (2008 through 2018), USGS (1991 through 2015a), (USGS 2018b), USGS (2014 through 2018b), USGS (2014 through 2017), USGS (1995 through 2013), USGS (1995, 1998, 2000, 2001, 2002, 2007), USGS (2018a), USGS (1991 through

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<sup>32</sup> The IPCC Tier 3B methodology is used for estimating emissions from commercial aircraft.

<sup>33</sup> Fuel consumption by U.S. Territories (i.e., American Samoa, Guam, Puerto Rico, U.S. Virgin Islands, Wake Island, and other U.S. Pacific Islands) is included in this report and contributed total emissions of 41.4 MMT CO<sub>2</sub> Eq. in 2017.

<sup>34</sup> See IPCC Reference Approach for Estimating CO<sub>2</sub> Emissions from Fossil Fuel Combustion in Annex 4 for a comparison of U.S. estimates using top-down and bottom-up approaches.

<sup>35</sup> A crude convention to convert between gross and net calorific values is to multiply the heat content of solid and liquid fossil fuels by 0.95 and gaseous fuels by 0.9 to account for the water content of the fuels. Biomass-based fuels in U.S. energy statistics, however, are generally presented using net calorific values.

2015c), USGS (1991 through 2017), USGS (2018b), USGS (2014 through 2018a), USGS (1996 through 2013), USGS (1991 through 2015b), USGS (2018c), USGS (1991 through 2015c).<sup>36</sup>

3. *Adjust for conversion of fuels and exports of CO<sub>2</sub>.* Fossil fuel consumption estimates are adjusted downward to exclude fuels created from other fossil fuels and exports of CO<sub>2</sub>.<sup>37</sup> Synthetic natural gas is created from industrial coal, and is currently included in EIA statistics for both coal and natural gas. Therefore, synthetic natural gas is subtracted from energy consumption statistics.<sup>38</sup> Since October 2000, the Dakota Gasification Plant has been exporting CO<sub>2</sub> to Canada by pipeline. Since this CO<sub>2</sub> is not emitted to the atmosphere in the United States, the associated fossil fuel burned to create the exported CO<sub>2</sub> is subtracted from fossil fuel consumption statistics. The associated fossil fuel is the total fossil fuel burned at the plant with the CO<sub>2</sub> capture system multiplied by the fraction of the plant's total site-generated CO<sub>2</sub> that is recovered by the capture system. To make these adjustments, additional data were collected from EIA (2019a), data for synthetic natural gas were collected from EIA (2018a), and data for CO<sub>2</sub> exports were collected from the Eastman Gasification Services Company (2011), Dakota Gasification Company (2006), Fitzpatrick (2002), Erickson (2003), EIA (2008) and DOE (2012).
4. *Adjust Sectoral Allocation of Distillate Fuel Oil and Motor Gasoline.* EPA had conducted a separate bottom-up analysis of transportation fuel consumption based on data from the Federal Highway Administration that indicated that the amount of distillate and motor gasoline consumption allocated to the transportation sector in the EIA statistics should be adjusted. Therefore, for these estimates, the transportation sector's distillate fuel and motor gasoline consumption was adjusted to match the value obtained from the bottom-up analysis. As the total distillate and motor gasoline consumption estimate from EIA are considered to be accurate at the national level, the distillate and motor gasoline consumption totals for the residential, commercial, and industrial sectors were adjusted proportionately. The data sources used in the bottom-up analysis of transportation fuel consumption include AAR (2008 through 2018), Benson (2002 through 2004), DOE (1993 through 2017), EIA (2007), EIA (1991 through 2018), EPA (2018b), and FHWA (1996 through 2017).<sup>39</sup>
5. *Adjust for fuels consumed for non-energy uses.* U.S. aggregate energy statistics include consumption of fossil fuels for non-energy purposes. These are fossil fuels that are manufactured into plastics, asphalt, lubricants, or other products. Depending on the end-use, this can result in storage of some or all of the C contained in the fuel for a period of time. As the emission pathways of C used for non-energy purposes are vastly different than fuel combustion (since the C in these fuels ends up in products instead of being combusted), these emissions are estimated separately in Section 3.2 – Carbon Emitted and Stored in Products from Non-Energy Uses of Fossil Fuels. Therefore, the amount of fuels used for non-energy purposes was subtracted from total fuel consumption. Data on non-fuel consumption was provided by EIA (2019a).
6. *Subtract consumption of international bunker fuels.* According to the UNFCCC reporting guidelines emissions from international transport activities, or bunker fuels, should not be included in national totals. U.S. energy consumption statistics include these bunker fuels (e.g., distillate fuel oil, residual fuel oil, and jet fuel) as part of consumption by the transportation end-use sector, however, so emissions from international transport activities were calculated separately following the same procedures used for emissions from consumption of all fossil fuels (i.e., estimation of consumption, and determination of C content).<sup>40</sup> The Office of the Under Secretary of Defense (Installations and Environment) and the Defense

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<sup>36</sup> See sections on Iron and Steel Production and Metallurgical Coke Production, Ammonia Production and Urea Consumption, Petrochemical Production, Titanium Dioxide Production, Ferroalloy Production, Aluminum Production, and Silicon Carbide Production and Consumption in the Industrial Processes and Product Use chapter.

<sup>37</sup> Energy statistics from EIA (2019a) are already adjusted downward to account for ethanol added to motor gasoline, biodiesel added to diesel fuel, and biogas in natural gas.

<sup>38</sup> These adjustments are explained in greater detail in Annex 2.1.

<sup>39</sup> Bottom-up gasoline and diesel highway vehicle fuel consumption estimates are based on data from FHWA Highway Statistics Table MF-21, MF-27, and VM-1 (FHWA 1996 through 2017).

<sup>40</sup> See International Bunker Fuels section in this chapter for a more detailed discussion.

Logistics Agency Energy (DLA Energy) of the U.S. Department of Defense (DoD) (DLA Energy 2018) supplied data on military jet fuel and marine fuel use. Commercial jet fuel use was obtained from FAA (2019); residual and distillate fuel use for civilian marine bunkers was obtained from DOC (1991 through 2018) for 1990 through 2001 and 2007 through 2017, and DHS (2008) for 2003 through 2006.<sup>41</sup> Consumption of these fuels was subtracted from the corresponding fuels in the transportation end-use sector. Estimates of international bunker fuel emissions for the United States are discussed in detail in Section 3.10 – International Bunker Fuels.

7. *Determine the total C content of fuels consumed.* Total C was estimated by multiplying the amount of fuel consumed by the amount of C in each fuel. This total C estimate defines the maximum amount of C that could potentially be released to the atmosphere if all of the C in each fuel was converted to CO<sub>2</sub>. The C content coefficients used by the United States were obtained from EIA's *Emissions of Greenhouse Gases in the United States 2008* (EIA 2009a), and an EPA analysis of C content coefficients developed for the GHGRP (EPA 2010). A discussion of the methodology used to develop the C content coefficients are presented in Annexes 2.1 and 2.2.
8. *Estimate CO<sub>2</sub> Emissions.* Total CO<sub>2</sub> emissions are the product of the adjusted energy consumption (from the previous methodology steps 1 through 6), the C content of the fuels consumed, and the fraction of C that is oxidized. The fraction oxidized was assumed to be 100 percent for petroleum, coal, and natural gas based on guidance in IPCC (2006) (see Annex 2.1).
9. *Allocate transportation emissions by vehicle type.* This report provides a more detailed accounting of emissions from transportation because it is such a large consumer of fossil fuels in the United States. For fuel types other than jet fuel, fuel consumption data by vehicle type and transportation mode were used to allocate emissions by fuel type calculated for the transportation end-use sector. Heat contents and densities were obtained from EIA (2019a) and USAF (1998).<sup>42</sup>
  - For on-road vehicles, annual estimates of combined motor gasoline and diesel fuel consumption by vehicle category were obtained from FHWA (1996 through 2017); for each vehicle category, the percent gasoline, diesel, and other (e.g., CNG, LPG) fuel consumption are estimated using data from DOE (1993 through 2017).<sup>43,44</sup>
  - For non-road vehicles, activity data were obtained from AAR (2008 through 2018), APTA (2007 through 2017), APTA (2006), BEA (2018), Benson (2002 through 2004), DOE (1993 through 2017), DLA Energy (2018), DOC (1991 through 2018), DOT (1991 through 2017), EIA (2009a), EIA (2019a), EIA (2018c), EIA (1991 through 2018), EPA (2018b),<sup>45</sup> and Gaffney (2007).
  - For jet fuel used by aircraft, CO<sub>2</sub> emissions from commercial aircraft were developed by the U.S.

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<sup>41</sup> Data for 2002 were interpolated due to inconsistencies in reported fuel consumption data.

<sup>42</sup> For a more detailed description of the data sources used for the analysis of the transportation end use sector see the Mobile Combustion (excluding CO<sub>2</sub>) and International Bunker Fuels sections of the Energy chapter, Annex 3.2, and Annex 3.8, respectively.

<sup>43</sup> Data from FHWA's Table VM-1 is used to estimate the share of fuel consumption between each on-road vehicle class. These fuel consumption estimates are combined with estimates of fuel shares by vehicle type from DOE's TEDB Annex Tables A.1 through A.6 (DOE 1993 through 2017). In 2011, FHWA changed its methods for estimating data in the VM-1 table. These methodological changes included how vehicles are classified, moving from a system based on body-type to one that is based on wheelbase. These changes were first incorporated for the 1990 through 2010 Inventory and apply to the time period from 2007 through 2015. This resulted in large changes in VMT and fuel consumption data by vehicle class, thus leading to a shift in emissions among on-road vehicle classes.

<sup>44</sup> Transportation sector natural gas and LPG consumption are based on data from EIA (2019a). In previous Inventory years, data from DOE TEDB was used to estimate each vehicle class's share of the total natural gas and LPG consumption. Since TEDB does not include estimates for natural gas use by medium- and heavy-duty trucks or LPG use by passenger cars, EIA Alternative Fuel Vehicle Data (Browning 2017) is now used to determine each vehicle class's share of the total natural gas and LPG consumption. These changes were first incorporated in the previous Inventory and apply to the time period from 1990 to 2015.

<sup>45</sup> In 2014, EPA incorporated the NONROAD2008 model into MOVES2014. The current Inventory uses the NONROAD component of MOVES2014b for years 1999 through 2017.

Federal Aviation Administration (FAA) using a Tier 3B methodology, consistent IPCC (2006) (see Annex 3.3). Carbon dioxide emissions from other aircraft were calculated directly based on reported consumption of fuel as reported by EIA. Allocation to domestic military uses was made using DoD data (see Annex 3.8). General aviation jet fuel consumption is calculated as the remainder of total jet fuel use (as determined by EIA) nets all other jet fuel use as determined by FAA and DoD. For more information, see Annex 3.2.

#### **Box 3-4: Uses of Greenhouse Gas Reporting Program Data and Improvements in Reporting Emissions from Industrial Sector Fossil Fuel Combustion**

As described in the calculation methodology, total fossil fuel consumption for each year is based on aggregated end-use sector consumption published by the EIA. The availability of facility-level combustion emissions through EPA's GHGRP has provided an opportunity to better characterize the industrial sector's energy consumption and emissions in the United States, through a disaggregation of EIA's industrial sector fuel consumption data from select industries.

For GHGRP 2010 through 2017 reporting years, facility-level fossil fuel combustion emissions reported through EPA's GHGRP were categorized and distributed to specific industry types by utilizing facility-reported NAICS codes (as published by the U.S. Census Bureau). As noted previously in this report, the definitions and provisions for reporting fuel types in EPA's GHGRP include some differences from the Inventory's use of EIA national fuel statistics to meet the UNFCCC reporting guidelines. The IPCC has provided guidance on aligning facility-level reported fuels and fuel types published in national energy statistics, which guided this exercise.<sup>46</sup>

As with previous Inventory reports, the current effort represents an attempt to align, reconcile, and coordinate the facility-level reporting of fossil fuel combustion emissions under EPA's GHGRP with the national-level approach presented in this report. Consistent with recommendations for reporting the Inventory to the UNFCCC, progress was made on certain fuel types for specific industries and has been included in the CRF tables that are submitted to the UNFCCC along with this report.<sup>47</sup> The efforts in reconciling fuels focus on standard, common fuel types (e.g., natural gas, distillate fuel oil, etc.) where the fuels in EIA's national statistics aligned well with facility-level GHGRP data. For these reasons, the current information presented in the CRF tables should be viewed as an initial attempt at this exercise. Additional efforts will be made for future Inventory reports to improve the mapping of fuel types, and examine ways to reconcile and coordinate any differences between facility-level data and national statistics. The current analysis includes the full time series presented in the CRF tables. Analyses were conducted linking GHGRP facility-level reporting with the information published by EIA in its MECS data in order to disaggregate the full 1990 through 2017 time period in the CRF tables. It is believed that the current analysis has led to improvements in the presentation of data in the Inventory, but further work will be conducted, and future improvements will be realized in subsequent Inventory reports. This includes incorporating the latest MECS data as it becomes available.

#### **Box 3-5: Carbon Intensity of U.S. Energy Consumption**

The amount of C emitted from the combustion of fossil fuels is dependent upon the C content of the fuel and the fraction of that C that is oxidized. Fossil fuels vary in their average C content, ranging from about 53 MMT CO<sub>2</sub> Eq./Qbtu for natural gas to upwards of 95 MMT CO<sub>2</sub> Eq./Qbtu for coal and petroleum coke. In general, the C content per unit of energy of fossil fuels is the highest for coal products, followed by petroleum, and then natural gas. The overall C intensity of the U.S. economy is thus dependent upon the quantity and combination of fuels and other energy sources employed to meet demand.

Table 3-16 provides a time series of the C intensity of direct emissions for each sector of the U.S. economy. The time series incorporates only the energy from the direct combustion of fossil fuels in each sector. For example, the C

<sup>46</sup> See Section 4 "Use of Facility-Level Data in Good Practice National Greenhouse Gas Inventories" of the IPCC meeting report, and specifically the section on using facility-level data in conjunction with energy data, at <[http://www.ipcc-nggip.iges.or.jp/public/tb/TFI\\_Technical\\_Bulletin\\_1.pdf](http://www.ipcc-nggip.iges.or.jp/public/tb/TFI_Technical_Bulletin_1.pdf)>.

<sup>47</sup> See <<https://www.epa.gov/ghgemissions/inventory-us-greenhouse-gas-emissions-and-sinks>>.

intensity for the residential sector does not include the energy from or emissions related to the use of electricity for lighting, as it is instead allocated to the electric power sector. For the purposes of maintaining the focus of this section, renewable energy and nuclear energy are not included in the energy totals used in Table 3-16 in order to focus attention on fossil fuel combustion as detailed in this chapter. Looking only at this direct consumption of fossil fuels, the residential sector exhibited the lowest C intensity, which is related to the large percentage of its energy derived from natural gas for heating. The C intensity of the commercial sector has predominantly declined since 1990 as commercial businesses shift away from petroleum to natural gas. The industrial sector was more dependent on petroleum and coal than either the residential or commercial sectors, and thus had higher C intensities over this period. The C intensity of the transportation sector was closely related to the C content of petroleum products (e.g., motor gasoline and jet fuel, both around 70 MMT CO<sub>2</sub> Eq./EJ), which were the primary sources of energy.<sup>48</sup> Lastly, the electric power sector had the highest C intensity due to its heavy reliance on coal for generating electricity.

**Table 3-16: Carbon Intensity from Direct Fossil Fuel Combustion by Sector (MMT CO<sub>2</sub> Eq./QBTu)**

Sector	1990	2005	2013	2014	2015	2016	2017
Residential <sup>a</sup>	57.4	56.6	55.3	55.4	55.5	55.1	55.0
Commercial <sup>a</sup>	59.6	57.7	56.0	55.7	57.2	56.8	56.6
Industrial <sup>a</sup>	64.4	64.5	62.0	61.5	61.2	60.8	60.5
Transportation <sup>a</sup>	71.1	71.4	71.4	71.5	71.5	71.5	71.5
Electric Power <sup>b</sup>	87.3	85.8	81.3	81.2	78.1	76.8	77.3
U.S. Territories <sup>c</sup>	73.0	73.5	71.9	72.3	72.2	72.2	72.2
<b>All Sectors<sup>c</sup></b>	<b>73.0</b>	<b>73.5</b>	<b>70.9</b>	<b>70.7</b>	<b>69.7</b>	<b>69.2</b>	<b>69.2</b>

<sup>a</sup> Does not include electricity or renewable energy consumption.

<sup>b</sup> Does not include electricity produced using nuclear or renewable energy.

<sup>c</sup> Does not include nuclear or renewable energy consumption.

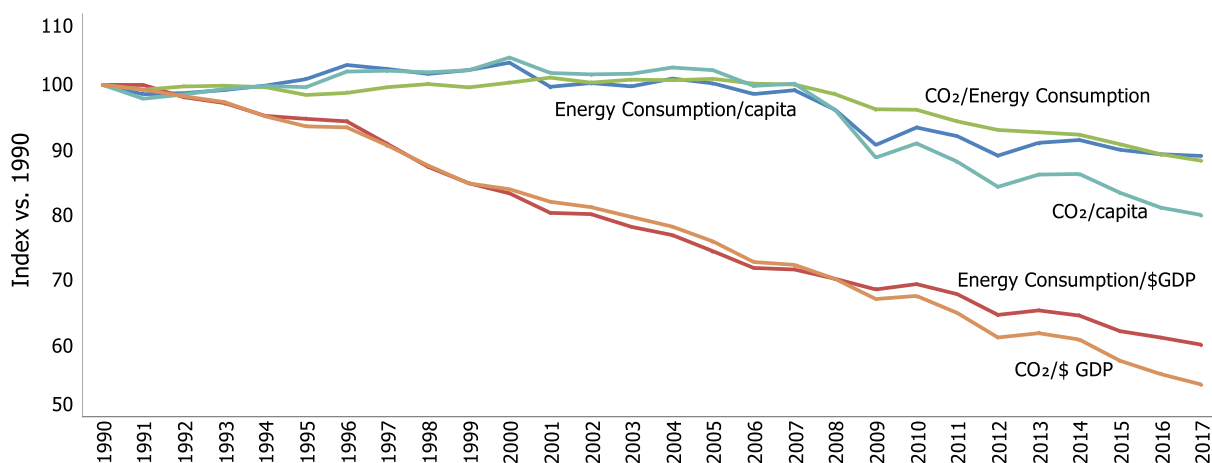
Note: Excludes non-energy fuel use emissions and consumption.

For the time period of 1990 through about 2008, the C intensity of U.S. energy consumption was fairly constant, as the proportion of fossil fuels used by the individual sectors did not change significantly over that time. Starting in 2008 the C intensity has decreased, reflecting the shift from coal to natural gas in the electric power sector during that time period. Per capita energy consumption fluctuated little from 1990 to 2007, but then started decreasing after 2007 and, in 2017, was approximately 10.9 percent below levels in 1990 (see Figure 3-16). To differentiate these estimates from those of Table 3-16, the C intensity trend shown in Figure 3-16 and described below includes nuclear and renewable energy EIA data to provide a comprehensive economy-wide picture of energy consumption. Due to a general shift from a manufacturing-based economy to a service-based economy, as well as overall increases in efficiency, energy consumption and energy-related CO<sub>2</sub> emissions per dollar of gross domestic product (GDP) have both declined since 1990 (BEA 2018).

<sup>48</sup> One exajoule (EJ) is equal to 10<sup>18</sup> joules or 0.9478 QBTu.



**Figure 3-16: U.S. Energy Consumption and Energy-Related CO<sub>2</sub> Emissions Per Capita and Per Dollar GDP**



C intensity estimates were developed using nuclear and renewable energy data from EIA (2019a), EPA (2010), and fossil fuel consumption data as discussed above and presented in Annex 2.1.

## Uncertainty and Time-Series Consistency

For estimates of CO<sub>2</sub> from fossil fuel combustion, the amount of CO<sub>2</sub> emitted is directly related to the amount of fuel consumed, the fraction of the fuel that is oxidized, and the carbon content of the fuel. Therefore, a careful accounting of fossil fuel consumption by fuel type, average carbon contents of fossil fuels consumed, and production of fossil fuel-based products with long-term carbon storage should yield an accurate estimate of CO<sub>2</sub> emissions.

Nevertheless, there are uncertainties in the consumption data, carbon content of fuels and products, and carbon oxidation efficiencies. For example, given the same primary fuel type (e.g., coal, petroleum, or natural gas), the amount of carbon contained in the fuel per unit of useful energy can vary. For the United States, however, the impact of these uncertainties on overall CO<sub>2</sub> emission estimates is believed to be relatively small. See, for example, Marland and Pippin (1990).

Although statistics of total fossil fuel and other energy consumption are relatively accurate, the allocation of this consumption to individual end-use sectors (i.e., residential, commercial, industrial, and transportation) is less certain. For example, for some fuels the sectoral allocations are based on price rates (i.e., tariffs), but a commercial establishment may be able to negotiate an industrial rate or a small industrial establishment may end up paying an industrial rate, leading to a misallocation of emissions. Also, the deregulation of the natural gas industry and the more recent deregulation of the electric power industry have likely led to some minor problems in collecting accurate energy statistics as firms in these industries have undergone significant restructuring.

To calculate the total CO<sub>2</sub> emission estimate from energy-related fossil fuel combustion, the amount of fuel used in non-energy production processes were subtracted from the total fossil fuel consumption. The amount of CO<sub>2</sub> emissions resulting from non-energy related fossil fuel use has been calculated separately and reported in the Carbon Emitted from Non-Energy Uses of Fossil Fuels section of this report (Section 3.2). These factors all contribute to the uncertainty in the CO<sub>2</sub> estimates. Detailed discussions on the uncertainties associated with C emitted from Non-Energy Uses of Fossil Fuels can be found within that section of this chapter.

Various sources of uncertainty surround the estimation of emissions from international bunker fuels, which are subtracted from the U.S. totals (see the detailed discussions on these uncertainties provided in Section 3.10 – International Bunker Fuels). Another source of uncertainty is fuel consumption by U.S. Territories. The United States does not collect energy statistics for its territories at the same level of detail as for the fifty states and the

District of Columbia. Therefore, estimating both emissions and bunker fuel consumption by these territories is difficult.

Uncertainties in the emission estimates presented above also result from the data used to allocate CO<sub>2</sub> emissions from the transportation end-use sector to individual vehicle types and transport modes. In many cases, bottom-up estimates of fuel consumption by vehicle type do not match aggregate fuel-type estimates from EIA. Further research is planned to improve the allocation into detailed transportation end-use sector emissions.

The uncertainty analysis was performed by primary fuel type for each end-use sector, using the IPCC-recommended Approach 2 uncertainty estimation methodology, Monte Carlo Stochastic Simulation technique, with @RISK software. For this uncertainty estimation, the inventory estimation model for CO<sub>2</sub> from fossil fuel combustion was integrated with the relevant variables from the inventory estimation model for International Bunker Fuels, to realistically characterize the interaction (or endogenous correlation) between the variables of these two models. About 120 input variables were modeled for CO<sub>2</sub> from energy-related Fossil Fuel Combustion (including about 10 for non-energy fuel consumption and about 20 for International Bunker Fuels).

In developing the uncertainty estimation model, uniform distributions were assumed for all activity-related input variables and emission factors, based on the SAIC/EIA (2001) report.<sup>49</sup> Triangular distributions were assigned for the oxidization factors (or combustion efficiencies). The uncertainty ranges were assigned to the input variables based on the data reported in SAIC/EIA (2001) and on conversations with various agency personnel.<sup>50</sup>

The uncertainty ranges for the activity-related input variables were typically asymmetric around their inventory estimates; the uncertainty ranges for the emissions factors were symmetric. Bias (or systematic uncertainties) associated with these variables accounted for much of the uncertainties associated with these variables (SAIC/EIA 2001).<sup>51</sup> For purposes of this uncertainty analysis, each input variable was simulated 10,000 times through Monte Carlo sampling.

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-17. Fossil fuel combustion CO<sub>2</sub> emissions in 2017 were estimated to be between 4,806.4 and 5,135.3 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. This indicates a range of 2 percent below to 5 percent above the 2017 emission estimate of 4,912.0 MMT CO<sub>2</sub> Eq.

**Table 3-17: Approach 2 Quantitative Uncertainty Estimates for CO<sub>2</sub> Emissions from Energy-Related Fossil Fuel Combustion by Fuel Type and Sector (MMT CO<sub>2</sub> Eq. and Percent)**

Fuel/Sector	2017 Emission Estimate (MMT CO <sub>2</sub> Eq.)	Uncertainty Range Relative to Emission Estimate <sup>a</sup>			
		(MMT CO <sub>2</sub> Eq.)		(%)	
		Lower Bound	Upper Bound	Lower Bound	Upper Bound
<b>Coal<sup>b</sup></b>	<b>1,267.5</b>	<b>1,223.5</b>	<b>1,386.1</b>	<b>-3%</b>	<b>9%</b>
Residential	NE	NE	NE	NE	NE
Commercial	2.0	1.9	2.3	-5%	15%
Industrial	54.4	51.8	63.0	-5%	16%
Transportation	NE	NE	NE	NE	NE
Electric Power	1,207.1	1,159.9	1,321.9	-4%	10%

<sup>49</sup> SAIC/EIA (2001) characterizes the underlying probability density function for the input variables as a combination of uniform and normal distributions (the former to represent the bias component and the latter to represent the random component). However, for purposes of the current uncertainty analysis, it was determined that uniform distribution was more appropriate to characterize the probability density function underlying each of these variables.

<sup>50</sup> In the SAIC/EIA (2001) report, the quantitative uncertainty estimates were developed for each of the three major fossil fuels used within each end-use sector; the variations within the sub-fuel types within each end-use sector were not modeled. However, for purposes of assigning uncertainty estimates to the sub-fuel type categories within each end-use sector in the current uncertainty analysis, SAIC/EIA (2001)-reported uncertainty estimates were extrapolated.

<sup>51</sup> Although, in general, random uncertainties are the main focus of statistical uncertainty analysis, when the uncertainty estimates are elicited from experts, their estimates include both random and systematic uncertainties. Hence, both these types of uncertainties are represented in this uncertainty analysis.

U.S. Territories	4.0	3.5	4.8	-12%	19%
<b>Natural Gas<sup>b</sup></b>	<b>1,450.3</b>	<b>1,433.8</b>	<b>1,517.1</b>	<b>-1%</b>	<b>5%</b>
Residential	241.5	234.7	258.5	-3%	7%
Commercial	173.2	168.3	185.2	-3%	7%
Industrial	484.7	470.3	519.6	-3%	7%
Transportation	42.3	41.1	45.2	-3%	7%
Electric Power	505.6	491.2	531.5	-3%	5%
U.S. Territories	3.0	2.6	3.5	-13%	17%
<b>Petroleum<sup>b</sup></b>	<b>2,193.7</b>	<b>2,061.2</b>	<b>2,324.4</b>	<b>-6%</b>	<b>6%</b>
Residential	53.0	50.1	55.8	-6%	5%
Commercial	57.7	54.4	60.8	-6%	5%
Industrial	271.5	215.4	322.7	-21%	19%
Transportation	1,758.3	1,645.4	1,870.4	-6%	6%
Electric Power	18.9	18.0	20.5	-5%	9%
U.S. Territories	34.3	31.7	38.0	-8%	11%
<b>Total (excluding Geothermal)<sup>b</sup></b>	<b>4,911.6</b>	<b>4,805.9</b>	<b>5,134.7</b>	<b>-2%</b>	<b>5%</b>
Geothermal	0.4	NE	NE	NE	NE
<b>Total (including Geothermal)<sup>b,c</sup></b>	<b>4,912.0</b>	<b>4,806.4</b>	<b>5,135.3</b>	<b>-2%</b>	<b>5%</b>

NE (Not Estimated)

Note: Totals may not sum due to independent rounding.

<sup>a</sup> Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval.

<sup>b</sup> The low and high estimates for total emissions were calculated separately through simulations and, hence, the low and high emission estimates for the sub-source categories do not sum to total emissions.

<sup>c</sup> Geothermal emissions added for reporting purposes, but an uncertainty analysis was not performed for CO<sub>2</sub> emissions from geothermal production.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2017. Details on the emission trends through time are described in more detail in the Methodology section, above. As discussed in Annex 5, data are unavailable to include estimates of CO<sub>2</sub> emissions from any liquid fuel used in pipeline transport or non-hazardous industrial waste incineration, but those emissions are assumed to be insignificant.

## QA/QC and Verification

In order to ensure the quality of the emission estimates from fossil fuel combustion CO<sub>2</sub>, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and methodology used for estimating CO<sub>2</sub> emissions from fossil fuel combustion in the United States. Emission totals for the different sectors and fuels were compared and trends were investigated to determine whether any corrective actions were needed. Minor corrective actions were taken.

## Recalculations Discussion

The Energy Information Administration (EIA 2019a) updated energy consumption statistics across the time series relative to the previous Inventory. EIA revised LPG consumption in the residential, commercial, industrial, and transportation sectors, and lubricants consumption in the industrial and transportation sectors, for the years 2010 through 2016. EIA revised kerosene consumption in the commercial and industrial sectors, and jet fuel and lubricants consumption in the transportation sector, for the year 1995. EIA revised distillate fuel consumption in the residential, commercial, industrial, and transportation sectors for years 2014 through 2016. EIA updated heat contents of motor gasoline, which changed motor gasoline consumption in the commercial, industrial, and transportation sectors for the years 1990 through 2016. EIA also revised 1990 and 1993 natural gas consumption in the residential, commercial, industrial, and transportation sectors, and 2016 natural gas consumption in all sectors. In addition, the number of significant figures increased for industrial coking coal, industrial “other” coal, and industrial “other” petroleum consumption data obtained from EIA, which decreased total energy consumption in the industrial sector by less than 0.05 percent but increased the precision of the data.

Annual carbon contents were updated for some fuels based on the availability of new data. Annual coal carbon contents were updated across the time series to incorporate domestic coal production data obtained from EIA (2018b), as well as state-specific coal sample data for Montana, Illinois, and Indiana. Annual natural gas carbon contents were also updated across the time series to incorporate annual heat content data for natural gas obtained from EIA (2019a). See Annex 2.2 for more detail on how coal and natural gas carbon contents are calculated.

Within the transportation sector, electricity use from electric vehicle charging in commercial and residential locations was re-allocated from the Residential and Commercial sectors to the Transportation sector, starting in 2010. See the recalculations discussion of the CH<sub>4</sub> and N<sub>2</sub>O from Mobile Combustion section below for more detail on how EV electricity use was estimated.

EPA also updated the methodology used to calculate emissions from geothermal electricity production. Technology-specific emission factors for geothermal electricity production obtained from EPA (2018c) were applied to total net electricity generation by geotype (i.e., binary, flash steam, dry steam) obtained from EIA (2019c).

Revisions to LPG, lubricants, kerosene, jet fuel, distillate fuel, and motor gasoline consumption resulted in an average annual decrease of 1.7 MMT CO<sub>2</sub> Eq. (0.1 percent) in CO<sub>2</sub> emissions from petroleum. Revisions to natural gas consumption resulted in an average annual decrease of 0.2 MMT CO<sub>2</sub> Eq. (less than 0.05 percent), while updates to annually variable natural gas carbon contents resulted in an average annual increase of 0.2 MMT CO<sub>2</sub> Eq. (less than 0.05 percent), in CO<sub>2</sub> emissions from natural gas. In aggregate, these changes resulted in an average annual increase of 0.1 MMT CO<sub>2</sub> Eq. (less than 0.05 percent) in CO<sub>2</sub> emissions from natural gas. Revisions to annually variable coal carbon contents resulted in an average annual decrease of 0.6 MMT CO<sub>2</sub> Eq. (less than 0.05 percent) in CO<sub>2</sub> emissions from coal. Updates to the methodology for estimating emissions from geothermal electricity production resulted in an average annual increase of 0.1 MMT CO<sub>2</sub> Eq. (20.6 percent) in CO<sub>2</sub> emissions from geothermal electricity production. Overall, these changes resulted in an average annual decrease of 2.1 MMT CO<sub>2</sub> Eq. (less than 0.05 percent) in CO<sub>2</sub> emissions from fossil fuel combustion for the period 1990 through 2016, relative to the previous Inventory.

## Planned Improvements

To reduce uncertainty of CO<sub>2</sub> from fossil fuel combustion estimates for U.S. Territories, efforts will be made to improve the quality of the U.S. Territories data, including through work with EIA and other agencies. This improvement is part of an ongoing analysis and efforts to continually improve the CO<sub>2</sub> from fossil fuel combustion estimates. In addition, further expert elicitation may be conducted to better quantify the total uncertainty associated with emissions from this source.

The availability of facility-level combustion emissions through EPA's GHGRP will continue to be examined to help better characterize the industrial sector's energy consumption in the United States, and further classify total industrial sector fossil fuel combustion emissions by business establishments according to industrial economic activity type. Most methodologies used in EPA's GHGRP are consistent with IPCC, though for EPA's GHGRP, facilities collect detailed information specific to their operations according to detailed measurement standards, which may differ with the more aggregated data collected for the Inventory to estimate total, national U.S. emissions. In addition, and unlike the reporting requirements for this chapter under the UNFCCC reporting guidelines, some facility-level fuel combustion emissions reported under the GHGRP may also include industrial process emissions.<sup>52</sup> In line with UNFCCC reporting guidelines, fuel combustion emissions are included in this chapter, while process emissions are included in the Industrial Processes and Product Use chapter of this report. In examining data from EPA's GHGRP that would be useful to improve the emission estimates for the CO<sub>2</sub> from fossil fuel combustion category, particular attention will also be made to ensure time-series consistency, as the facility-level reporting data from EPA's GHGRP are not available for all inventory years as reported in this Inventory.

Additional analyses will be conducted to align reported facility-level fuel types and IPCC fuel types per the national energy statistics. For example, efforts will be taken to incorporate updated industrial fuel consumption data from EIA's Manufacturing Energy Consumption Survey (MECS), with updated data for 2014. Additional work will look at CO<sub>2</sub> emissions from biomass to ensure they are separated in the facility-level reported data, and maintaining

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<sup>52</sup> See <<https://unfccc.int/resource/docs/2013/cop19/eng/10a03.pdf#page=2>>.

consistency with national energy statistics provided by EIA. In implementing improvements and integration of data from EPA's GHGRP, the latest guidance from the IPCC on the use of facility-level data in national inventories will continue to be relied upon.<sup>53</sup>

An ongoing planned improvement is to develop improved estimates of domestic waterborne fuel consumption. The Inventory estimates for residual and distillate fuel used by ships and boats is based in part on data on bunker fuel use from the U.S. Department of Commerce. Domestic fuel consumption is estimated by subtracting fuel sold for international use from the total sold in the United States. It may be possible to more accurately estimate domestic fuel use and emissions by using detailed data on marine ship activity. The feasibility of using domestic marine activity data to improve the estimates will continue to be investigated.

EPA will evaluate and potentially update methods for allocating motor gasoline consumption to the transportation, industrial, and commercial sectors. In 2016, FHWA changed its methods for estimating the share of gasoline used in on-road and non-road applications, creating a time-series inconsistency in the current Inventory between 2015 and previous years.<sup>54</sup> EPA will continue to explore approaches to address this inconsistency, including using MOVES on-road fuel consumption output to define the percentage of the FHWA consumption totals (from MF-21) that are attributable to transportation, and applying that percentage to the EIA total. This would define gasoline consumption from transportation, such that the remainder would be defined as consumption by the industrial and commercial sectors.

EPA will continue to evaluate updates to the annual coal carbon content coefficients, such as continuing to integrate new information from state-level geological surveys. EPA will also explore potential updates to annual variability in carbon contents for petroleum fuels developed by EIA, such as potential updates to the data sources used to develop transportation CO<sub>2</sub> factors for motor gasoline and low-sulfur diesel fuel.

EPA is also evaluating the methods used to adjust for conversion of fuels and exports of CO<sub>2</sub>. EPA is exploring the approach used to account for CO<sub>2</sub> transport, injection, and geologic storage, as part of this there may be changes made to accounting for CO<sub>2</sub> exports. EPA is also exploring the data provided by EIA in terms of tracking supplemental natural gas which may impact the treatment of adjustments for synthetic fuels.

## CH<sub>4</sub> and N<sub>2</sub>O from Stationary Combustion

### Methodology

Methane and N<sub>2</sub>O emissions from stationary combustion were estimated by multiplying fossil fuel and wood consumption data by emission factors (by sector and fuel type for industrial, residential, commercial, and U.S. Territories; and by fuel and technology type for the electric power sector). The electric power sector utilizes a Tier 2 methodology, whereas all other sectors utilize a Tier 1 methodology. The activity data and emission factors used are described in the following subsections.

#### *Industrial, Residential, Commercial, and U.S. Territories*

National coal, natural gas, fuel oil, and wood consumption data were grouped by sector: industrial, commercial, residential, and U.S. Territories. For the CH<sub>4</sub> and N<sub>2</sub>O emission estimates, consumption data for each fuel were obtained from EIA's *Monthly Energy Review* (EIA 2019). Because the United States does not include territories in its national energy statistics, fuel consumption data for territories were provided separately by EIA's International Energy Statistics (EIA 2017).<sup>55</sup> Fuel consumption for the industrial sector was adjusted to subtract out mobile source construction and agricultural use, which is reported under mobile sources. Construction and agricultural

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<sup>53</sup> See <[http://www.ipcc-nggip.iges.or.jp/public/tb/TFI\\_Technical\\_Bulletin\\_1.pdf](http://www.ipcc-nggip.iges.or.jp/public/tb/TFI_Technical_Bulletin_1.pdf)>.

<sup>54</sup> The previous and new FHWA methodologies for estimating non-road gasoline are described in *Off-Highway and Public-Use Gasoline Consumption Estimation Models Used in the Federal Highway Administration*, Publication Number FHWA-PL-17-012. <<https://www.fhwa.dot.gov/policyinformation/pubs/pl17012.pdf>>

<sup>55</sup> U.S. Territories data also include combustion from mobile activities because data to allocate territories' energy use were unavailable. For this reason, CH<sub>4</sub> and N<sub>2</sub>O emissions from combustion by U.S. Territories are only included in the stationary combustion totals.

mobile source fuel use was obtained from EPA (2018b) and FHWA (1996 through 2017). Estimates for wood biomass consumption for fuel combustion do not include municipal solid waste, tires, etc., that are reported as biomass by EIA. Non-CO<sub>2</sub> emissions from combustion of the biogenic portion of municipal solid waste and tires is included under waste incineration. Tier 1 default emission factors for the industrial, commercial, and residential end-use sectors were provided by the *2006 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC 2006). U.S. Territories' emission factors were estimated using the U.S. emission factors for the primary sector in which each fuel was combusted.

### *Electric Power Sector*

The electric power sector uses a Tier 2 emission estimation methodology as fuel consumption for the electric power sector by control-technology type was based on EPA's Acid Rain Program Dataset (EPA 2018a). Total fuel consumption in the electric power sector from EIA (2019) was apportioned to each combustion technology type and fuel combination using a ratio of fuel consumption by technology type derived from EPA (2018a) data. The combustion technology and fuel use data by facility obtained from EPA (2018a) were only available from 1996 to 2017, so the consumption estimates from 1990 to 1995 were estimated by applying the 1996 consumption ratio by combustion technology type from EPA (2018a) to the total EIA (2019) consumption for each year from 1990 to 1995.

Emissions were estimated by multiplying fossil fuel and wood consumption by technology-, fuel-, and country-specific Tier 2 emission factors. The Tier 2 emission factors used are based in part on emission factors published by EPA, and EPA's Compilation of Air Pollutant Emission Factors, AP-42 (EPA 1997) for coal wall-fired boilers, natural gas-fired turbines, and combined cycle natural gas units.<sup>56</sup>

More detailed information on the methodology for calculating emissions from stationary combustion, including emission factors and activity data, is provided in Annex 3.1.

## **Uncertainty and Time-Series Consistency**

Methane emission estimates from stationary sources exhibit high uncertainty, primarily due to difficulties in calculating emissions from wood combustion (i.e., fireplaces and wood stoves). The estimates of CH<sub>4</sub> and N<sub>2</sub>O emissions presented are based on broad indicators of emissions (i.e., fuel use multiplied by an aggregate emission factor for different sectors), rather than specific emission processes (i.e., by combustion technology and type of emission control).

An uncertainty analysis was performed by primary fuel type for each end-use sector, using the IPCC-recommended Approach 2 uncertainty estimation methodology, Monte Carlo Stochastic Simulation technique, with @RISK software.

The uncertainty estimation model for this source category was developed by integrating the CH<sub>4</sub> and N<sub>2</sub>O stationary source inventory estimation models with the model for CO<sub>2</sub> from fossil fuel combustion to realistically characterize the interaction (or endogenous correlation) between the variables of these three models. About 55 input variables were simulated for the uncertainty analysis of this source category (about 20 from the CO<sub>2</sub> emissions from fossil fuel combustion inventory estimation model and about 35 from the stationary source inventory models).

In developing the uncertainty estimation model, uniform distribution was assumed for all activity-related input variables and N<sub>2</sub>O emission factors, based on the SAIC/EIA (2001) report.<sup>57</sup> For these variables, the uncertainty

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<sup>56</sup> Several of the U.S. Tier 2 emission factors were used in IPCC (2006) as Tier 1 emission factors. See Table A-92 in Annex 3.1 for emission factors by technology type and fuel type for the electric power sector.

<sup>57</sup> SAIC/EIA (2001) characterizes the underlying probability density function for the input variables as a combination of uniform and normal distributions (the former distribution to represent the bias component and the latter to represent the random component). However, for purposes of the current uncertainty analysis, it was determined that uniform distribution was more appropriate to characterize the probability density function underlying each of these variables.

ranges were assigned to the input variables based on the data reported in SAIC/EIA (2001).<sup>58</sup> However, the CH<sub>4</sub> emission factors differ from those used by EIA. These factors and uncertainty ranges are based on IPCC default uncertainty estimates (IPCC 2006).

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-18. Stationary combustion CH<sub>4</sub> emissions in 2017 (including biomass) were estimated to be between 5.2 and 17.5 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. This indicates a range of 33 percent below to 124 percent above the 2017 emission estimate of 7.8 MMT CO<sub>2</sub> Eq.<sup>59</sup> Stationary combustion N<sub>2</sub>O emissions in 2017 (including biomass) were estimated to be between 20.6 and 43.4 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. This indicates a range of 28 percent below to 52 percent above the 2017 emission estimate of 28.6 MMT CO<sub>2</sub> Eq.

**Table 3-18: Approach 2 Quantitative Uncertainty Estimates for CH<sub>4</sub> and N<sub>2</sub>O Emissions from Energy-Related Stationary Combustion, Including Biomass (MMT CO<sub>2</sub> Eq. and Percent)**

Source	Gas	2017 Emission Estimate (MMT CO <sub>2</sub> Eq.)	Uncertainty Range Relative to Emission Estimate <sup>a</sup>			
			Lower Bound	Upper Bound	Lower Bound (%)	Upper Bound (%)
Stationary Combustion	CH <sub>4</sub>	7.8	5.2	17.5	-33%	+124%
Stationary Combustion	N <sub>2</sub> O	28.6	20.6	43.4	-28%	+52%

<sup>a</sup> Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval.

The uncertainties associated with the emission estimates of CH<sub>4</sub> and N<sub>2</sub>O are greater than those associated with estimates of CO<sub>2</sub> from fossil fuel combustion, which mainly rely on the carbon content of the fuel combusted. Uncertainties in both CH<sub>4</sub> and N<sub>2</sub>O estimates are due to the fact that emissions are estimated based on emission factors representing only a limited subset of combustion conditions. For the indirect greenhouse gases, uncertainties are partly due to assumptions concerning combustion technology types, age of equipment, emission factors used, and activity data projections.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2017 as discussed below. Details on the emission trends through time are described in more detail in the Methodology section, above. As discussed in Annex 5, data are unavailable to include estimates of CH<sub>4</sub> and N<sub>2</sub>O emissions from biomass use in territories, but those emissions are assumed to be insignificant.

For more information on the general QA/QC process applied to this source category, consistent with Volume 1, Chapter 6 of the *2006 IPCC Guidelines*, see QA/QC and Verification Procedures section in the introduction of the IPPU Chapter.

## QA/QC and Verification

In order to ensure the quality of the emission estimates from stationary combustion non-CO<sub>2</sub>, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and emission factor sources and methodology used for estimating CH<sub>4</sub>, N<sub>2</sub>O, and the greenhouse gas precursors from stationary combustion in the United States. Emission totals for the different sectors and fuels were compared and trends were investigated.

<sup>58</sup> In the SAIC/EIA (2001) report, the quantitative uncertainty estimates were developed for each of the three major fossil fuels used within each end-use sector; the variations within the sub-fuel types within each end-use sector were not modeled. However, for purposes of assigning uncertainty estimates to the sub-fuel type categories within each end-use sector in the current uncertainty analysis, SAIC/EIA (2001)-reported uncertainty estimates were extrapolated.

<sup>59</sup> The low emission estimates reported in this section have been rounded down to the nearest integer values and the high emission estimates have been rounded up to the nearest integer values.

## Recalculations Discussion

Methane and N<sub>2</sub>O emissions from stationary sources (excluding CO<sub>2</sub>) across the entire time series were revised due to revised data from EIA (2019) and EPA (2018a) relative to the previous Inventory. Most notably, EIA (2019) updated wood biomass consumption statistics in the residential sector from 2009 to 2016 and the commercial sector from 2014 to 2016. EPA revised the methodology for estimating CH<sub>4</sub> and N<sub>2</sub>O electric power emissions due to differences between total fuel consumption in the electric power sector reported by EIA (2019) and EPA (2018a). In addition, nitrous oxide emission factors for coal wall-fired boilers used in the electric power sector were updated from 0.5 kg/TJ to 5.8 kg/TJ to be consistent with EPA's Compilation of Air Pollutant Emission Factors, AP-42 (EPA 1997). The historical data changes and methodology updates resulted in an average annual increase of 0.1 MMT CO<sub>2</sub> Eq. (1.2 percent) in CH<sub>4</sub> emissions, and an average annual increase of 15.3 MMT CO<sub>2</sub> Eq. (107.2 percent) in N<sub>2</sub>O emissions from stationary combustion for the 1990 through 2016 period.

## Planned Improvements

Several items are being evaluated to improve the CH<sub>4</sub> and N<sub>2</sub>O emission estimates from stationary combustion and to reduce uncertainty for U.S. Territories. Efforts will be taken to work with EIA and other agencies to improve the quality of the U.S. Territories data. Because these data are not broken out by stationary and mobile uses, further research will be aimed at trying to allocate consumption appropriately. In addition, the uncertainty of biomass emissions will be further investigated since it was expected that the exclusion of biomass from the estimates would reduce the uncertainty; and in actuality the exclusion of biomass increases the uncertainty. These improvements are not all-inclusive, but are part of an ongoing analysis and efforts to continually improve these stationary combustion estimates from U.S. Territories.

Fuel use was adjusted for the industrial sector to subtract out construction and agricultural use, which is reported under mobile sources. Mobile source CH<sub>4</sub> and N<sub>2</sub>O also include emissions from sources that may be captured as part of the commercial sector. Future research will look into the need to adjust commercial sector fuel consumption to account for sources included elsewhere.

## CH<sub>4</sub> and N<sub>2</sub>O from Mobile Combustion

### Methodology

Estimates of CH<sub>4</sub> and N<sub>2</sub>O emissions from mobile combustion were calculated by multiplying emission factors by measures of activity for each fuel and vehicle type (e.g., light-duty gasoline trucks). Activity data included vehicle miles traveled (VMT) for on-road vehicles and fuel consumption for non-road mobile sources. The activity data and emission factors used are described in the subsections that follow. A complete discussion of the methodology used to estimate CH<sub>4</sub> and N<sub>2</sub>O emissions from mobile combustion and the emission factors used in the calculations is provided in Annex 3.2.

### *On-Road Vehicles*

Estimates of CH<sub>4</sub> and N<sub>2</sub>O emissions from gasoline and diesel on-road vehicles are based on VMT and emission factors by vehicle type, fuel type, model year, and emission control technology. Emission estimates for alternative fuel vehicles (AFVs) are based on VMT and emission factors by vehicle and fuel type.<sup>60</sup>

Emissions factors for N<sub>2</sub>O from newer on-road gasoline vehicles were calculated based upon a regression analysis done by EPA (ICF 2017a). Methane emission factors were calculated based on the ratio of NMOG emission standards for newer vehicles. Older gasoline vehicles on-road emissions factors were developed by ICF (2004). These factors were derived from EPA, California Air Resources Board (CARB) and Environment Canada laboratory test results of different vehicle and control technology types. The EPA, CARB and Environment Canada tests were designed following the Federal Test Procedure (FTP), which covers three separate driving segments, since vehicles

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<sup>60</sup> Alternative fuel and advanced technology vehicles are those that can operate using a motor fuel other than gasoline or diesel. This includes electric or other bi-fuel or dual-fuel vehicles that may be partially powered by gasoline or diesel.



emit varying amounts of greenhouse gases depending on the driving segment. These driving segments are: (1) a transient driving cycle that includes cold start and running emissions, (2) a cycle that represents running emissions only, and (3) a transient driving cycle that includes hot start and running emissions. For each test run, a bag was affixed to the tailpipe of the vehicle and the exhaust was collected; the content of this bag was then analyzed to determine quantities of gases present. The emissions characteristics of segment 2 were used to define running emissions, and subtracted from the total FTP emissions to determine start emissions. These were then recombined based upon the ratio of start to running emissions for each vehicle class from MOBILE6.2, an EPA emission factor model that predicts gram per mile emissions of CO<sub>2</sub>, CO, HC, NO<sub>x</sub>, and PM from vehicles under various conditions, to approximate average driving characteristics.<sup>61</sup> Diesel on-road vehicle emission factors were developed by ICF (2006b).

CH<sub>4</sub> and N<sub>2</sub>O emission factors for AFVs were developed based on the 2017 GREET model. For light-duty trucks, EPA used a curve fit of 1999 through 2011 travel fractions for LDT1 and LDT2 (MOVES Source Type 31 for LDT1 and MOVES Source Type 32 for LDT2). For medium-duty vehicles, EPA used emission factors for light heavy-duty vocational trucks. For heavy-duty vehicles, EPA used emission factors for long haul combination trucks. For buses, EPA used emission factors for transit buses. These values represent vehicle operation only (tank-to-wheels); well-to-tank emissions are calculated elsewhere in the Inventory.

Annual VMT data for 1990 through 2017 were obtained from the Federal Highway Administration's (FHWA) Highway Performance Monitoring System database as reported in Highway Statistics (FHWA 1996 through 2017).<sup>62</sup> VMT estimates were then allocated from FHWA's vehicle categories to fuel-specific vehicle categories using the calculated shares of vehicle fuel use for each vehicle category by fuel type reported in DOE (1993 through 2017) and information on total motor vehicle fuel consumption by fuel type from FHWA (1996 through 2017). VMT for AFVs were estimated based on Browning (2017 and 2018a). The age distributions of the U.S. vehicle fleet were obtained from EPA (2018b, 2000), and the average annual age-specific vehicle mileage accumulation of U.S. vehicles were obtained from EPA (2018b).

Control technology and standards data for on-road vehicles were obtained from EPA's Office of Transportation and Air Quality (EPA 2017a, 2017b, 2000, 1998, and 1997) and Browning (2005). These technologies and standards are defined in Annex 3.2, and were compiled from EPA (1994a, 1994b, 1998, 1999a) and IPCC (2006).

### *Non-Road Mobile Sources*

To estimate emissions from non-road mobile sources, fuel consumption data were employed as a measure of activity, and multiplied by fuel-specific emission factors (in grams of N<sub>2</sub>O and CH<sub>4</sub> per kilogram of fuel consumed).<sup>63</sup> Activity data were obtained from AAR (2008 through 2018), APTA (2007 through 2017), APTA (2006), BEA (1991 through 2015), Benson (2002 through 2004), DHS (2008), DLA Energy (2018), DOC (1991 through 2018), DOE (1993 through 2017), DOT (1991 through 2017), EIA (2002, 2007, 2019a), EIA (2019b), EIA (1991 through 2018), EPA (2018b), Esser (2003 through 2004), FAA (2019), FHWA (1996 through 2017),<sup>64</sup>

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<sup>61</sup> Additional information regarding the MOBILE model can be found online at <<https://www.epa.gov/moves/description-and-history-mobile-highway-vehicle-emission-factor-model>>.

<sup>62</sup> The source of VMT is FHWA Highway Statistics Table VM-1. In 2011, FHWA changed its methods for estimating data in the VM-1 table. These methodological changes included how vehicles are classified, moving from a system based on body-type to one that is based on wheelbase. These changes were first incorporated for the 1990 through 2010 Inventory and apply to the 2007 through 2017 time period. This resulted in large changes in VMT by vehicle class, thus leading to a shift in emissions among on-road vehicle classes. For example, the category "Passenger Cars" has been replaced by "Light-duty Vehicles-Short Wheelbase" and "Other 2 axle-4 Tire Vehicles" has been replaced by "Light-duty Vehicles, Long Wheelbase." This change in vehicle classification has moved some smaller trucks and sport utility vehicles from the light truck category to the passenger vehicle category in the current Inventory. These changes are reflected in a large drop in light-truck emissions between 2006 and 2007.

<sup>63</sup> The consumption of international bunker fuels is not included in these activity data, but is estimated separately under the International Bunker Fuels source category.

<sup>64</sup> This Inventory uses FHWA's Agriculture, Construction, and Commercial/Industrial MF-24 fuel volumes along with the MOVES NONROAD model gasoline volumes to estimate non-road mobile source CH<sub>4</sub> and N<sub>2</sub>O emissions for these categories. For agriculture, the MF-24 gasoline volume is used directly because it includes both off-road trucks and equipment. For

Gaffney (2007), and Whorton (2006 through 2014). Emission factors for non-road modes were taken from IPCC (2006) and Browning (2018b).

## Uncertainty and Time-Series Consistency

A quantitative uncertainty analysis was conducted for the mobile source sector using the IPCC-recommended Approach 2 uncertainty estimation methodology, Monte Carlo Stochastic Simulation technique, using @RISK software. The uncertainty analysis was performed on 2017 estimates of CH<sub>4</sub> and N<sub>2</sub>O emissions, incorporating probability distribution functions associated with the major input variables. For the purposes of this analysis, the uncertainty was modeled for the following four major sets of input variables: (1) VMT data, by on-road vehicle and fuel type and (2) emission factor data, by on-road vehicle, fuel, and control technology type, (3) fuel consumption, data, by non-road vehicle and equipment type, and (4) emission factor data, by non-road vehicle and equipment type.

Uncertainty analyses were not conducted for NO<sub>x</sub>, CO, or NMVOC emissions. Emission factors for these gases have been extensively researched since emissions of these gases from motor vehicles are regulated in the United States, and the uncertainty in these emission estimates is believed to be relatively low. For more information, see Section 3.9 – Uncertainty Analysis of Emission Estimates. However, a much higher level of uncertainty is associated with CH<sub>4</sub> and N<sub>2</sub>O emission factors due to limited emission test data, and because, unlike CO<sub>2</sub> emissions, the emission pathways of CH<sub>4</sub> and N<sub>2</sub>O are highly complex.

Mobile combustion CH<sub>4</sub> emissions from all mobile sources in 2017 were estimated to be between 2.9 and 4.1 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. This indicates a range of 8 percent below to 27 percent above the corresponding 2017 emission estimate of 3.2 MMT CO<sub>2</sub> Eq. Also at a 95 percent confidence level, mobile combustion N<sub>2</sub>O emissions from mobile sources in 2017 were estimated to be between 15.5 and 19.3 MMT CO<sub>2</sub> Eq., indicating a range of 8 percent below to 14 percent above the corresponding 2017 emission estimate of 16.9 MMT CO<sub>2</sub> Eq.

**Table 3-19: Approach 2 Quantitative Uncertainty Estimates for CH<sub>4</sub> and N<sub>2</sub>O Emissions from Mobile Sources (MMT CO<sub>2</sub> Eq. and Percent)**

Source	Gas	2017 Emission Estimate <sup>a</sup> (MMT CO <sub>2</sub> Eq.)	Uncertainty Range Relative to Emission Estimate <sup>a</sup> (MMT CO <sub>2</sub> Eq.)			
			Lower Bound	Upper Bound	Lower Bound (%)	Upper Bound (%)
Mobile Sources	CH <sub>4</sub>	3.2	2.9	4.1	-8%	+27%
Mobile Sources	N <sub>2</sub> O	16.9	15.5	19.3	-8%	+14%

<sup>a</sup> Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval.

This uncertainty analysis is a continuation of a multi-year process for developing quantitative uncertainty estimates for this source category using the IPCC Approach 2 uncertainty analysis. As a result, as new information becomes available, uncertainty characterization of input variables may be improved and revised. For additional information regarding uncertainty in emission estimates for CH<sub>4</sub> and N<sub>2</sub>O please refer to the Uncertainty Annex. As discussed in Annex 5, data are unavailable to include estimates of CH<sub>4</sub> and N<sub>2</sub>O emissions from any liquid fuel used in pipeline transport or some biomass used in transportation sources, but those emissions are assumed to be insignificant.

## QA/QC and Verification

In order to ensure the quality of the emission estimates from mobile combustion, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The specific plan used for mobile combustion was

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construction and commercial/industrial gasoline estimates, the 2014 and older MF-24 volumes represented off-road trucks only; therefore, the MOVES NONROAD gasoline volumes for construction and commercial/industrial are added to the respective categories in the Inventory. Beginning in 2015, this addition is no longer necessary since the FHWA updated its methods for estimating on-road and non-road gasoline consumption. Among the method updates, FHWA now incorporates MOVES NONROAD equipment gasoline volumes in the construction and commercial/industrial categories.

updated prior to collection and analysis of this current year of data. The Tier 2 procedures focused on the emission factor and activity data sources, as well as the methodology used for estimating emissions. These procedures included a qualitative assessment of the emission estimates to determine whether they appear consistent with the most recent activity data and emission factors available. A comparison of historical emissions between the current Inventory and the previous Inventory was also conducted to ensure that the changes in estimates were consistent with the changes in activity data and emission factors.

## Recalculations Discussion

Updates were made to the non-road CH<sub>4</sub> and N<sub>2</sub>O emissions calculations for the current Inventory, resulting in decreases across the time series in emissions from alternative fuel highway vehicles and non-highway sources, such as construction and farm equipment. The collective result of these changes was a net increase in CH<sub>4</sub> emissions and a decrease in N<sub>2</sub>O emissions from mobile combustion relative to the previous Inventory. Methane emissions increased by 0.2 percent. Nitrous oxide emissions decreased by 0.3 percent. Each of these changes is described below.

Inventory calculations now reflect MOVES2014b, the latest version of EPA's MOVES model, released in the summer of 2018. This update affected emissions from non-highway mobile sources across the time series. Amongst some more minor updates, MOVES2014b includes updated non-road engine population growth rates, resulting in generally lower equipment populations, fuel consumption, and emissions. For this year's Inventory, new non-road CH<sub>4</sub> and N<sub>2</sub>O emission factors were calculated using the updated 2006 IPCC Tier 3 guidance and EPA's MOVES2014b model. Methane emission factors were calculated directly from MOVES. Nitrous oxide emission factors were calculated using MOVES activity and emission factors by fuel type from the European Environment Agency. Gasoline engines were broken out by 2- and 4-stroke engine types using MOVES2014b.

Alternative fuel vehicle CH<sub>4</sub> and N<sub>2</sub>O emissions are estimated using VMT data developed by Browning (2017 and 2018a) and are based on Energy Information Administration (EIA) Alternative Fuel Vehicle Data. EIA recently updated their historical data for vehicle counts and fuel consumption which decreased overall AFV emissions across this Inventory year's 1990-2016 time series. This year's Inventory also includes updated and corrected methane emissions factors for biodiesel use in heavy-duty vehicles from 2007 and onwards.

An updated methodology (Browning 2018a) was used to estimate VMT from battery electric vehicles (BEVs), plug-in hybrid vehicles (PHEVs), neighborhood electric vehicles (NEVs), and electric buses in this year's Inventory. Monthly vehicle sales by make and model are now obtained from hybridcars.com, and fuel consumption by vehicle type, make, and model is supplied by fueleconomy.gov. Average annual mileage estimates by vehicle type are sourced from Federal Highway Administration (FHWA) Highway Statistics VM-1 table. PHEVs use both electricity and gasoline. Miles driven in all-electric mode depends upon the vehicle's all-electric range. A fleet utility factor (SAE 2010) is now used to estimate the average percentage of miles that were all-electric based upon the all-electric range (AER) in miles of a vehicle model. Similar to the approach for BEVs, PHEV energy consumption in all-electric mode was calculated from vehicle counts, fuel consumption rates, and vehicle miles travelled.

Populations of NEVs from 2003 to 2010 are estimated from EIA data tables of total vehicle counts and the electric vehicle fleet counts for all but low-speed vehicles. Fleet vehicle counts are subtracted from the total electric vehicle counts to estimate the population of NEVs. These values are then extrapolated to calendar years after 2010 using a regression analysis. Fuel consumption is estimated from EIA fleet data by dividing the energy used (in GGE) by the number of vehicles and then multiplying by the number of vehicles estimated above. Fuel economy for NEVs is assumed to be the same as light-duty automobile BEVs. All-electric bus vehicle population counts and fuel consumption (in GGE) are directly supplied by EIA fleet data. More detail on the methods to account for emissions from BEVs, PHEVs, NEVs, and electric buses is provided in Annex 3.2 Mobile Combustion and Browning (2018a).

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2017 with one recent notable exception. An update by FHWA to the method for estimating on-road VMT created an inconsistency in on-road CH<sub>4</sub> and N<sub>2</sub>O for the time periods 1990 to 2006 and 2007 to 2017. Details on the emission trends and methodological inconsistencies through time are described in the Methodology section, above.

## Planned Improvements

While the data used for this report represent the most accurate information available, several areas have been identified that could potentially be improved in the near term given available resources.

- Evaluate and potentially update EPA’s method for estimating motor gasoline consumption for non-road mobile sources to improve accuracy and create a more consistent time series. As discussed in the Methodology section above and in Annex 3.2, CH<sub>4</sub> and N<sub>2</sub>O estimates for gasoline-powered non-road sources in this Inventory are based on a variety of inputs, including FHWA Highway Statistics Table MF-24. In 2016, FHWA changed its methods for estimating the share of gasoline used in on-road and non-road applications.<sup>65</sup> These method changes created a time-series inconsistency in the current Inventory between 2015 and previous years in CH<sub>4</sub> and N<sub>2</sub>O estimates for agricultural, construction, commercial, and industrial non-road mobile sources. In the current Inventory EPA has implemented one approach to address this inconsistency. EPA will test other approaches including using MOVES on-road fuel consumption output to define the percentage of the FHWA consumption totals (from MF-21) that are attributable to on-highway transportation sources. This percentage would then be applied to the EIA total, thereby defining gasoline consumption from on-highway transportation sources, such that the remainder would be defined as consumption by the industrial and commercial sectors.
- Explore updates to on-road diesel emissions factors for CH<sub>4</sub> and N<sub>2</sub>O to incorporate diesel after treatment technology for light-duty vehicles.
- Continue to explore potential improvements to estimates of domestic waterborne fuel consumption for future Inventories. The Inventory estimates for residual and distillate fuel used by ships and boats is based in part on data on bunker fuel use from the U.S. Department of Commerce. Domestic fuel consumption is estimated by subtracting fuel sold for international use from the total sold in the United States. It may be possible to more accurately estimate domestic fuel use and emissions by using detailed data on marine ship activity. The feasibility of using domestic marine activity data to improve the estimates continues to be investigated. Additionally, the feasibility of including data from a broader range of domestic and international sources for domestic bunker fuels, including data from studies such as the *Third IMO GHG Study 2014*, continues to be explored.
- Update the methodology for estimating Class II and Class III rail diesel fuel consumption. For many years, the American Short-line and Regional Railroad Association (ASLRRA) supplied annual data on Class II and Class III rail diesel fuel consumption (national totals), but is no longer able to do so. One alternative approach would be to estimate fuel use based on rail car loadings. EPA will explore potential updates to annual variability in carbon contents for petroleum fuels developed by EIA, such as potential updates to the data sources used to develop transportation CO<sub>2</sub> factors for motor gasoline and low-sulfur diesel fuel.

## 3.2 Carbon Emitted from Non-Energy Uses of Fossil Fuels (CRF Source Category 1A5)

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In addition to being combusted for energy, fossil fuels are also consumed for non-energy uses in the United States. The fuels used for these purposes are diverse, including natural gas, liquefied petroleum gases (LPG), asphalt (a viscous liquid mixture of heavy crude oil distillates), petroleum coke (manufactured from heavy oil), and coal (metallurgical) coke (manufactured from coking coal). The non-energy applications of these fuels are equally diverse, including feedstocks for the manufacture of plastics, rubber, synthetic fibers and other materials; reducing agents for the production of various metals and inorganic products; and products such as lubricants, waxes, and

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<sup>65</sup> The previous and new FHWA methodologies for estimating non-road gasoline are described in *Off-Highway and Public-Use Gasoline Consumption Estimation Models Used in the Federal Highway Administration*, Publication Number FHWA-PL-17-012. <<https://www.fhwa.dot.gov/policyinformation/pubs/pl17012.pdf>>.

asphalt (IPCC 2006). Emissions from a portion of non-energy uses of fossil fuels are reported in the Energy sector, as opposed to the Industrial Processes and Product Use (IPPU) sector, to reflect national circumstances in its choice of methodology and to increase transparency of this source category’s unique country-specific data sources and methodology (see Box 3-6).

Carbon dioxide emissions arise from non-energy uses via several pathways. Emissions may occur during the manufacture of a product, as is the case in producing plastics or rubber from fuel-derived feedstocks. Additionally, emissions may occur during the product’s lifetime, such as during solvent use. Overall, throughout the time series and across all uses, about 62 percent of the total C consumed for non-energy purposes was stored in products, and not released to the atmosphere; the remaining 38 percent was emitted.

There are several areas in which non-energy uses of fossil fuels are closely related to other parts of this Inventory. For example, some of the non-energy use products release CO<sub>2</sub> at the end of their commercial life when they are combusted after disposal; these emissions are reported separately within the Energy chapter in the Incineration of Waste source category. In addition, there is some overlap between fossil fuels consumed for non-energy uses and the fossil-derived CO<sub>2</sub> emissions accounted for in the IPPU chapter, especially for fuels used as reducing agents. To avoid double counting, the “raw” non-energy fuel consumption data reported by EIA are modified to account for these overlaps. There are also net exports of petrochemicals that are not completely accounted for in the EIA data, and the Inventory calculations adjust for the effect of net exports on the mass of C in non-energy applications.

As shown in Table 3-20, fossil fuel emissions in 2017 from the non-energy uses of fossil fuels were 123.2 MMT CO<sub>2</sub> Eq., which constituted approximately 2 percent of overall fossil fuel emissions. In 2017, the consumption of fuels for non-energy uses (after the adjustments described above) was 5,017.5 TBtu (see Table 3-21). A portion of the C in the 5,017.5 TBtu of fuels was stored (217.9 MMT CO<sub>2</sub> Eq.), while the remaining portion was emitted (123.2 MMT CO<sub>2</sub> Eq.). Non-energy use emissions increased 8.4 percent from 2016 to 2017 mainly due to increases in coking coal and petrochemical feedstock use, both of which are driven by changes in economic activity and changes in the industrial sector, see Annex 2.3 for more details.

**Table 3-20: CO<sub>2</sub> Emissions from Non-Energy Use Fossil Fuel Consumption (MMT CO<sub>2</sub> Eq. and Percent)**

Year	1990	2005	2013	2014	2015	2016	2017
Potential Emissions	312.1	377.5	328.9	325.1	340.5	329.9	341.1
C Stored	192.5	237.9	205.4	205.2	213.6	216.2	217.9
Emissions as a % of Potential	38%	37%	38%	37%	37%	34%	36%
<b>Emissions</b>	<b>119.6</b>	<b>139.6</b>	<b>123.5</b>	<b>119.9</b>	<b>126.9</b>	<b>113.7</b>	<b>123.2</b>

## Methodology

The first step in estimating C stored in products was to determine the aggregate quantity of fossil fuels consumed for non-energy uses. The C content of these feedstock fuels is equivalent to potential emissions, or the product of consumption and the fuel-specific C content values. Both the non-energy fuel consumption and C content data were supplied by the EIA (2019) (see Annex 2.1). Consumption values for industrial coking coal, petroleum coke, other oils, and natural gas in Table 3-21 and Table 3-22 have been adjusted to subtract non-energy uses that are included in the source categories of the Industrial Processes and Product Use chapter.<sup>66,67</sup> Consumption of natural gas, LPG, pentanes plus, naphthas, other oils, and special naphtha were adjusted to subtract out net exports of these products

<sup>66</sup> These source categories include Iron and Steel Production, Lead Production, Zinc Production, Ammonia Manufacture, Carbon Black Manufacture (included in Petrochemical Production), Titanium Dioxide Production, Ferroalloy Production, Silicon Carbide Production, and Aluminum Production.

<sup>67</sup> Some degree of double counting may occur between these estimates of non-energy use of fuels and process emissions from petrochemical production presented in the Industrial Processes and Produce Use sector. Data integration is not feasible at this time as feedstock data from EIA used to estimate non-energy uses of fuels are aggregated by fuel type, rather than disaggregated by both fuel type and particular industries (e.g., petrochemical production) as currently collected through EPA’s GHGRP and used for the petrochemical production category.

that are not reflected in the raw data from EIA. Consumption values were also adjusted to subtract net exports of intermediary chemicals.

For the remaining non-energy uses, the quantity of C stored was estimated by multiplying the potential emissions by a storage factor.

- For several fuel types—petrochemical feedstocks (including natural gas for non-fertilizer uses, LPG, pentanes plus, naphthas, other oils, still gas, special naphtha, and industrial other coal), asphalt and road oil, lubricants, and waxes—U.S. data on C stocks and flows were used to develop C storage factors, calculated as the ratio of (a) the C stored by the fuel’s non-energy products to (b) the total C content of the fuel consumed. A lifecycle approach was used in the development of these factors in order to account for losses in the production process and during use. Because losses associated with municipal solid waste management are handled separately in the Energy sector under the Incineration of Waste source category, the storage factors do not account for losses at the disposal end of the life cycle.
- For industrial coking coal and distillate fuel oil, storage factors were taken from IPCC (2006), which in turn draws from Marland and Rotty (1984).
- For the remaining fuel types (petroleum coke, miscellaneous products, and other petroleum), IPCC (2006) does not provide guidance on storage factors, and assumptions were made based on the potential fate of C in the respective non-energy use products. Carbon dioxide emissions from carbide production are implicitly accounted for in the storage factor calculation for the non-energy use of petroleum coke.

**Table 3-21: Adjusted Consumption of Fossil Fuels for Non-Energy Uses (Tbtu)**

Year	1990	2005	2013	2014	2015	2016	2017
<b>Industry</b>	<b>4,215.8</b>	<b>5,110.7</b>	<b>4,607.8</b>	<b>4,602.9</b>	<b>4,764.6</b>	<b>4,634.6</b>	<b>4,797.9</b>
Industrial Coking Coal	NO	80.4	119.3	48.8	121.8	89.2	112.5
Industrial Other Coal	8.2	11.9	10.3	10.3	10.3	10.3	10.3
Natural Gas to Chemical Plants	281.6	260.9	296.8	323.5	321.9	308.9	306.9
Asphalt & Road Oil	1,170.2	1,323.2	783.3	792.6	831.7	853.4	849.2
LPG	1,120.5	1,610.0	2,062.9	2,109.8	2,157.5	2,118.9	2,186.8
Lubricants	186.3	160.2	125.1	130.7	142.1	135.1	124.6
Pentanes Plus	117.6	95.5	45.4	43.5	78.4	53.1	81.5
Naphtha (<401 °F)	326.3	679.5	498.8	435.2	417.8	396.9	410.9
Other Oil (>401 °F)	662.1	499.5	209.1	236.2	216.8	204.0	241.7
Still Gas	36.7	67.7	166.7	164.5	162.2	166.1	163.8
Petroleum Coke	27.2	105.2	NO	NO	NO	NO	NO
Special Naphtha	100.9	60.9	96.6	104.5	97.0	88.7	94.9
Distillate Fuel Oil	7.0	11.7	5.8	5.8	5.8	5.8	5.8
Waxes	33.3	31.4	16.5	14.8	12.4	12.8	10.2
Miscellaneous Products	137.8	112.8	171.2	182.7	188.9	191.3	198.8
<b>Transportation</b>	<b>176.0</b>	<b>151.3</b>	<b>143.4</b>	<b>149.4</b>	<b>162.8</b>	<b>154.4</b>	<b>142.3</b>
Lubricants	176.0	151.3	143.4	149.4	162.8	154.4	142.3
<b>U.S. Territories</b>	<b>85.6</b>	<b>123.2</b>	<b>82.4</b>	<b>77.3</b>	<b>77.3</b>	<b>77.3</b>	<b>77.3</b>
Lubricants	0.7	4.6	1.0	1.0	1.0	1.0	1.0
Other Petroleum (Misc. Prod.)	84.9	118.6	81.4	76.2	76.2	76.2	76.2
<b>Total</b>	<b>4,477.4</b>	<b>5,385.2</b>	<b>4,833.6</b>	<b>4,829.6</b>	<b>5,004.7</b>	<b>4,866.2</b>	<b>5,017.5</b>

NO (Not Occurring)

**Table 3-22: 2017 Adjusted Non-Energy Use Fossil Fuel Consumption, Storage, and Emissions**

Adjusted Non-Energy Use <sup>a</sup>	Carbon Content Coefficient	Potential Carbon	Storage Factor	Carbon Stored	Carbon Emissions	Carbon Emissions
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Sector/Fuel Type	(TBtu)	(MMT C/QBtu)			(MMT CO <sub>2</sub> Eq.)		
		(MMT C)	(MMT C)	(MMT C)	(MMT C)	(MMT C)	(MMT CO <sub>2</sub> Eq.)
<b>Industry</b>	<b>4,797.9</b>	<b>NA</b>	<b>88.6</b>	<b>NA</b>	<b>59.0</b>	<b>29.6</b>	<b>108.5</b>
Industrial Coking Coal	112.5	31.00	3.5	0.10	0.3	3.1	11.5
Industrial Other Coal	10.3	26.06	0.3	0.67	0.2	0.1	0.3
Natural Gas to							
Chemical Plants	306.9	14.47	4.4	0.67	3.0	1.5	5.3
Asphalt & Road Oil	849.2	20.55	17.5	1.00	17.4	0.1	0.3
LPG	2,186.8	17.06	37.3	0.67	25.1	12.2	44.8
Lubricants	124.6	20.20	2.5	0.09	0.2	2.3	8.4
Pentanes Plus	81.5	19.10	1.6	0.67	1.0	0.5	1.9
Naphtha (<401° F)	410.9	18.55	7.6	0.67	5.1	2.5	9.2
Other Oil (>401° F)	241.7	20.17	4.9	0.67	3.3	1.6	5.9
Still Gas	163.8	17.51	2.9	0.67	1.9	0.9	3.4
Petroleum Coke	NO	27.85	NO	0.30	NO	NO	NO
Special Naphtha	94.9	19.74	1.9	0.67	1.3	0.6	2.2
Distillate Fuel Oil	5.8	20.17	0.1	0.50	0.1	0.1	0.2
Waxes	10.2	19.80	0.2	0.58	0.1	0.1	0.3
Miscellaneous Products	198.8	20.31	4.0	0.00	0.0	4.0	14.8
<b>Transportation</b>	<b>142.3</b>	<b>NA</b>	<b>2.9</b>	<b>NA</b>	<b>0.3</b>	<b>2.6</b>	<b>9.6</b>
Lubricants	142.3	20.20	2.9	0.09	0.3	2.6	9.6
<b>U.S. Territories</b>	<b>77.3</b>	<b>NA</b>	<b>1.5</b>	<b>NA</b>	<b>0.2</b>	<b>1.4</b>	<b>5.1</b>
Lubricants	1.0	20.20	0.0	0.09	+	+	0.1
Other Petroleum (Misc. Prod.)	76.2	20.00	1.5	0.10	0.2	1.4	5.0
<b>Total</b>	<b>5,017.5</b>		<b>93.0</b>		<b>59.4</b>	<b>33.6</b>	<b>123.2</b>

+ Does not exceed 0.05 TBtu, MMT C, MMT CO<sub>2</sub> Eq.

NA (Not Applicable)

NO (Not Occurring)

<sup>a</sup> To avoid double counting, net exports have been deducted.

Note: Totals may not sum due to independent rounding.

Lastly, emissions were estimated by subtracting the C stored from the potential emissions (see Table 3-20). More detail on the methodology for calculating storage and emissions from each of these sources is provided in Annex 2.3.

Where storage factors were calculated specifically for the United States, data were obtained on (1) products such as asphalt, plastics, synthetic rubber, synthetic fibers, cleansers (soaps and detergents), pesticides, food additives, antifreeze and deicers (glycols), and silicones; and (2) industrial releases including energy recovery, Toxics Release Inventory (TRI) releases, hazardous waste incineration, and volatile organic compound, solvent, and non-combustion CO emissions. Data were taken from a variety of industry sources, government reports, and expert communications. Sources include EPA reports and databases such as compilations of air emission factors (EPA 2001), *National Emissions Inventory (NEI) Air Pollutant Emissions Trends Data* (EPA 2018a), *Toxics Release Inventory, 1998* (EPA 2000b), *Biennial Reporting System* (EPA 2000a, 2009), *Resource Conservation and Recovery Act Information System* (EPA 2013b, 2015, 2016b, 2018c), pesticide sales and use estimates (EPA 1998, 1999, 2002, 2004, 2011, 2017), and the Chemical Data Access Tool (EPA 2014b); the EIA Manufacturer's Energy Consumption Survey (MECS) (EIA 1994, 1997, 2001, 2005, 2010, 2013, 2017); the National Petrochemical & Refiners Association (NPRO 2002); the U.S. Census Bureau (1999, 2004, 2009, 2014); Bank of Canada (2012, 2013, 2014, 2016, 2017, 2018); Financial Planning Association (2006); INEGI (2006); the United States International Trade Commission (1990 through 2017); Gosselin, Smith, and Hodge (1984); EPA's *Municipal Solid Waste (MSW) Facts and Figures* (EPA 2013, 2014a, 2016a, 2018b); the Rubber Manufacturers' Association (RMA 2009, 2011, 2014, 2016, 2018); the International Institute of Synthetic Rubber Products (IISRP 2000, 2003); the Fiber Economics Bureau (FEB 2001, 2003, 2005, 2007, 2009, 2010, 2011, 2012, 2013); Chemical and Engineering News (C&EN 2017); the Independent Chemical Information Service (ICIS 2008, 2016); the EPA Chemical Data Access Tool (CDAT) (EPA 2014b); the American Chemistry Council (ACC 2003 through 2011, 2013, 2014, 2015a, 2016b, 2017b, 2018b); and the *Guide to the Business of Chemistry* (ACC 2012, 2015b, 2016a, 2017a, 2018a). Specific data sources are listed in full detail in Annex 2.3.

## Uncertainty and Time-Series Consistency

An uncertainty analysis was conducted to quantify the uncertainty surrounding the estimates of emissions and storage factors from non-energy uses. This analysis, performed using @RISK software and the IPCC-recommended Approach 2 methodology (Monte Carlo Stochastic Simulation technique), provides for the specification of probability density functions for key variables within a computational structure that mirrors the calculation of the inventory estimate. The results presented below provide the 95 percent confidence interval, the range of values within which emissions are likely to fall, for this source category.

As noted above, the non-energy use analysis is based on U.S.-specific storage factors for (1) feedstock materials (natural gas, LPG, pentanes plus, naphthas, other oils, still gas, special naphthas, and other industrial coal), (2) asphalt, (3) lubricants, and (4) waxes. For the remaining fuel types (the “other” category in Table 3-21 and Table 3-22), the storage factors were taken directly from IPCC (2006), where available, and otherwise assumptions were made based on the potential fate of carbon in the respective NEU products. To characterize uncertainty, five separate analyses were conducted, corresponding to each of the five categories. In all cases, statistical analyses or expert judgments of uncertainty were not available directly from the information sources for all the activity variables; thus, uncertainty estimates were determined using assumptions based on source category knowledge.

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-23 (emissions) and Table 3-24 (storage factors). Carbon emitted from non-energy uses of fossil fuels in 2017 was estimated to be between 94.3 and 168.5 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. This indicates a range of 23 percent below to 37 percent above the 2017 emission estimate of 123.2 MMT CO<sub>2</sub> Eq. The uncertainty in the emission estimates is a function of uncertainty in both the quantity of fuel used for non-energy purposes and the storage factor.

**Table 3-23: Approach 2 Quantitative Uncertainty Estimates for CO<sub>2</sub> Emissions from Non-Energy Uses of Fossil Fuels (MMT CO<sub>2</sub> Eq. and Percent)**

Source	Gas	2017 Emission Estimate (MMT CO <sub>2</sub> Eq.)	Uncertainty Range Relative to Emission Estimate <sup>a</sup>			
			(MMT CO <sub>2</sub> Eq.)		(%)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Feedstocks	CO <sub>2</sub>	73.0	51.2	122.6	-30%	+68%
Asphalt	CO <sub>2</sub>	0.3	0.1	0.6	-59%	+118%
Lubricants	CO <sub>2</sub>	18.0	14.9	21.0	-18%	+16%
Waxes	CO <sub>2</sub>	0.3	0.2	0.6	-24%	+93%
Other	CO <sub>2</sub>	31.6	18.7	34.2	-41%	+8%
<b>Total</b>	<b>CO<sub>2</sub></b>	<b>123.2</b>	<b>94.3</b>	<b>168.5</b>	<b>-23%</b>	<b>37%</b>

<sup>a</sup> Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval.

Note: Totals may not sum due to independent rounding.

**Table 3-24: Approach 2 Quantitative Uncertainty Estimates for Storage Factors of Non-Energy Uses of Fossil Fuels (Percent)**

Source	Gas	2017 Storage Factor (%)	Uncertainty Range Relative to Emission Estimate <sup>a</sup>			
			(%)		(% , Relative)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Feedstocks	CO <sub>2</sub>	67.2%	54.8%	71.5%	-18%	+6%
Asphalt	CO <sub>2</sub>	99.6%	99.1%	99.8%	-0.5%	+0.3%
Lubricants	CO <sub>2</sub>	9.2%	3.9%	17.5%	-58%	+91%
Waxes	CO <sub>2</sub>	57.8%	47.4%	67.3%	-18%	+16%
Other	CO <sub>2</sub>	6.1%	6.0%	41.3%	-2%	+576%

<sup>a</sup> Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval, as a percentage of the inventory value (also expressed in percent terms).



As shown in Table 3-24, feedstocks and asphalt contribute least to overall storage factor uncertainty on a percentage basis. Although the feedstocks category—the largest use category in terms of total carbon flows—appears to have tight confidence limits, this is to some extent an artifact of the way the uncertainty analysis was structured. As discussed in Annex 2.3, the storage factor for feedstocks is based on an analysis of six fates that result in long-term storage (e.g., plastics production), and eleven that result in emissions (e.g., volatile organic compound emissions). Rather than modeling the total uncertainty around all of these fate processes, the current analysis addresses only the storage fates, and assumes that all C that is not stored is emitted. As the production statistics that drive the storage values are relatively well-characterized, this approach yields a result that is probably biased toward understating uncertainty.

As is the case with the other uncertainty analyses discussed throughout this document, the uncertainty results above address only those factors that can be readily quantified. More details on the uncertainty analysis are provided in Annex 2.3.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2017 as discussed below. Details on the emission trends through time are described in more detail in the Methodology section, above.

## QA/QC and Verification

In order to ensure the quality of the emission estimates from non-energy uses of fossil fuels, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. This effort included a general analysis, as well as portions of a category specific analysis for non-energy uses involving petrochemical feedstocks and for imports and exports. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and methodology for estimating the fate of C (in terms of storage and emissions) across the various end-uses of fossil C. Emission and storage totals for the different subcategories were compared, and trends across the time series were analyzed to determine whether any corrective actions were needed. Corrective actions were taken to rectify minor errors and to improve the transparency of the calculations, facilitating future QA/QC.

For petrochemical import and export data, special attention was paid to NAICS numbers and titles to verify that none had changed or been removed. Import and export totals were compared with 2016 totals as well as their trends across the time series.

Petrochemical input data reported by EIA will continue to be investigated in an attempt to address an input/output discrepancy in the NEU model. Prior to 2001, the C balance inputs exceeded outputs, then starting in 2001 through 2009, outputs exceeded inputs. Inputs exceeded outputs in 2010, 2011, 2013 through 2015, and 2017, but outputs exceeded inputs in 2012 and 2016. A portion of this discrepancy has been reduced and two strategies have been developed to address the remaining portion (see the Planned Improvements section, below).

## Recalculations Discussion

The Energy Information Administration (EIA 2019) updated energy consumption statistics across the time series relative to the previous Inventory. EPA released updated hazardous waste incineration data (EPA 2018c), which included minor updates to 2015 values. Overall, these changes resulted in an average annual increase of 0.4 MMT CO<sub>2</sub> Eq. (0.3 percent) in carbon emissions from non-energy uses of fossil fuels for the period 1990 through 2016, relative to the previous Inventory.

## Planned Improvements

There are several future improvements planned:

- Analyzing the fuel and feedstock data from EPA's GHGRP subpart X (Petrochemical Production) to better disaggregate CO<sub>2</sub> emissions in NEU model and CO<sub>2</sub> process emissions from petrochemical production.
- More accurate accounting of C in petrochemical feedstocks. EPA has worked with EIA to determine the

cause of input/output discrepancies in the C mass balance contained within the NEU model. In the future, two strategies to reduce or eliminate this discrepancy will continue to be pursued. First, accounting of C in imports and exports will be improved. The import/export adjustment methodology will be examined to ensure that net exports of intermediaries such as ethylene and propylene are fully accounted for. Second, the use of top-down C input calculation in estimating emissions will be reconsidered. Alternative approaches that rely more substantially on the bottom-up C output calculation will be considered instead.

- Improving the uncertainty analysis. Most of the input parameter distributions are based on professional judgment rather than rigorous statistical characterizations of uncertainty.
- Better characterizing flows of fossil C. Additional fates may be researched, including the fossil C load in organic chemical wastewaters, plasticizers, adhesives, films, paints, and coatings. There is also a need to further clarify the treatment of fuel additives and backflows (especially methyl tert-butyl ether, MTBE).
- Reviewing the trends in fossil fuel consumption for non-energy uses. Annual consumption for several fuel types is highly variable across the time series, including industrial coking coal and other petroleum (miscellaneous products). A better understanding of these trends will be pursued to identify any mischaracterized or misreported fuel consumption for non-energy uses. For example, “miscellaneous products” category includes miscellaneous products that are not reported elsewhere in the EIA data set. The EIA does not have firm data concerning the amounts of various products that are being reported in the “miscellaneous products” category; however, EIA has indicated that recovered sulfur from petroleum and natural gas processing, and potentially also C black feedstock could be reported in this category. Recovered sulfur would not be reported in the NEU calculation or elsewhere in the Inventory.
- Updating the average C content of solvents was researched, since the entire time series depends on one year’s worth of solvent composition data. The data on C emissions from solvents that were readily available do not provide composition data for all categories of solvent emissions and also have conflicting definitions for volatile organic compounds, the source of emissive C in solvents. Additional sources of solvents data will be investigated in order to update the C content assumptions.
- Updating the average C content of cleansers (soaps and detergents) was researched; although production and consumption data for cleansers are published every 5 years by the Census Bureau, the composition (C content) of cleansers has not been recently updated. Recently available composition data sources may facilitate updating the average C content for this category.
- Revising the methodology for consumption, production, and C content of plastics was researched; because of recent changes to the type of data publicly available for plastics, the NEU model for plastics applies data obtained from personal communications. Potential revisions to the plastics methodology to account for the recent changes in published data will be investigated.
- Although U.S.-specific storage factors have been developed for feedstocks, asphalt, lubricants, and waxes, default values from IPCC are still used for two of the non-energy fuel types (industrial coking coal, distillate oil), and broad assumptions are being used for miscellaneous products and other petroleum. Over the long term, there are plans to improve these storage factors by analyzing C fate similar to those described in Annex 2.3 or deferring to more updated default storage factors from IPCC where available.
- Reviewing the storage of carbon black across various sectors in the Inventory; in particular, the carbon black abraded and stored in tires.

### Box 3-6: Reporting of Lubricants, Waxes, and Asphalt and Road Oil Product Use in Energy Sector

IPCC (2006) provides methodological guidance to estimate emissions from the first use of fossil fuels as a product for primary purposes other than combustion for energy purposes (including lubricants, paraffin waxes, bitumen / asphalt, and solvents) under the IPPU sector.<sup>68</sup> In this Inventory, C storage and C emissions from product use of

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<sup>68</sup> See for example Volume 3: Industrial Processes and Product Use, and Chapter 5: Non-Energy Products from Fuels and Solvent Use of the *2006 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC 2006).

lubricants, waxes, and asphalt and road oil are reported under the Energy sector in the Carbon Emitted from Non-Energy Uses of Fossil Fuels source category (CRF Source Category 1A5).<sup>69</sup>

The emissions are reported in the Energy sector, as opposed to the IPPU sector, to reflect national circumstances in its choice of methodology and to increase transparency of this source category's unique country-specific data sources and methodology. The country-specific methodology used for the Carbon Emitted from Non-Energy Uses of Fossil Fuels source category is based on a carbon balance (i.e., C inputs-outputs) calculation of the aggregate amount of fossil fuels used for non-energy uses, including inputs of lubricants, waxes, asphalt and road oil (see Section 3.2, Table 3-22). For those inputs, U.S. country-specific data on C stocks and flows are used to develop carbon storage factors, which are calculated as the ratio of the C stored by the fossil fuel non-energy products to the total C content of the fuel consumed, taking into account losses in the production process and during product use.<sup>70</sup> The country-specific methodology to reflect national circumstances starts with the aggregate amount of fossil fuels used for non-energy uses and applies a C balance calculation, breaking out the C emissions from non-energy use of lubricants, waxes, and asphalt and road oil. Due to U.S. national circumstances, reporting these C emissions separately under IPPU would involve making artificial adjustments to allocate both the C inputs and C outputs of the non-energy use C balance. These artificial adjustments would also result in the C emissions for lubricants, waxes, and asphalt and road oil being reported under IPPU, while the C storage for lubricants, waxes, and asphalt and road oil would be reported under Energy. To avoid presenting an incomplete C balance and a less transparent approach for the Carbon Emitted from Non-Energy Uses of Fossil Fuels source category calculation, the entire calculation of C storage and C emissions is therefore conducted in the Non-Energy Uses of Fossil Fuels category calculation methodology, and both the C storage and C emissions for lubricants, waxes, and asphalt and road oil are reported under the Energy sector.

However, portions of the fuel consumption data for seven fuel categories—coking coal, distillate fuel, industrial other coal, petroleum coke, natural gas, residual fuel oil, and other oil—were reallocated to the IPPU chapter, as they were consumed during non-energy related industrial activity. Emissions from uses of fossil fuels as feedstocks or reducing agents (e.g., petrochemical production, aluminum production, titanium dioxide and zinc production) are reported in the IPPU chapter, unless otherwise noted due to specific national circumstances.

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## 3.3 Incineration of Waste (CRF Source Category 1A5)

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Incineration is used to manage about 7 to 19 percent of the solid wastes generated in the United States, depending on the source of the estimate and the scope of materials included in the definition of solid waste (EPA 2000; EPA 2018a; Goldstein and Madtes 2001; Kaufman et al. 2004; Simmons et al. 2006; van Haaren et al. 2010). In the context of this section, waste includes all municipal solid waste (MSW) as well as scrap tires. In the United States, incineration of MSW tends to occur at waste-to-energy facilities or industrial facilities where useful energy is recovered, and thus emissions from waste incineration are accounted for in the Energy chapter. Similarly, scrap tires are combusted for energy recovery in industrial and utility boilers, pulp and paper mills, and cement kilns. Incineration of waste results in conversion of the organic inputs to CO<sub>2</sub>. According to the *2006 IPCC Guidelines*, when the CO<sub>2</sub> emitted is of fossil origin, it is counted as a net anthropogenic emission of CO<sub>2</sub> to the atmosphere. Thus, the emissions from waste incineration are calculated by estimating the quantity of waste combusted and the fraction of the waste that is C derived from fossil sources.

Most of the organic materials in MSW are of biogenic origin (e.g., paper, yard trimmings), and have their net C flows accounted for under the Land Use, Land-Use Change, and Forestry chapter. However, some components—

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<sup>69</sup> Non-methane volatile organic compound (NMVOC) emissions from solvent use are reported separately in the IPPU sector, following Chapter 5 of the *2006 IPCC Guidelines*.

<sup>70</sup> Data and calculations for lubricants and waxes and asphalt and road oil are in Annex 2.3 – Methodology for Estimating Carbon Emitted from Non-Energy Uses of Fossil Fuels.

plastics, synthetic rubber, synthetic fibers, and carbon black in scrap tires—are of fossil origin. Plastics in the U.S. waste stream are primarily in the form of containers, packaging, and durable goods. Rubber is found in durable goods, such as carpets, and in non-durable goods, such as clothing and footwear. Fibers in MSW are predominantly from clothing and home furnishings. As noted above, scrap tires (which contain synthetic rubber and carbon black) are also considered a “non-hazardous” waste and are included in the waste incineration estimate, though waste disposal practices for tires differ from MSW. Estimates on emissions from hazardous waste incineration can be found in Annex 2.3 and are accounted for as part of the C mass balance for non-energy uses of fossil fuels.

Approximately 20.8 million metric tons of MSW were incinerated in 2011 (van Haaren et al. 2010). Updated data were not available for 2012 through 2017 from this source so the data were proxied to the 2011 estimate. Carbon dioxide emissions from incineration of waste increased 27 percent since 1990, to an estimated 10.8 MMT CO<sub>2</sub> (10,790 kt) in 2017, as the volume of scrap tires and other fossil C-containing materials in waste increased (see Table 3-25 and Table 3-26). Waste incineration is also a source of CH<sub>4</sub> and N<sub>2</sub>O emissions (De Soete 1993; IPCC 2006). Methane emissions from the incineration of waste were estimated to be less than 0.05 MMT CO<sub>2</sub> Eq. (less than 0.5 kt CH<sub>4</sub>) in 2017, and have decreased by 32 percent since 1990. Nitrous oxide emissions from the incineration of waste were estimated to be 0.3 MMT CO<sub>2</sub> Eq. (1 kt N<sub>2</sub>O) in 2017, and have decreased by 32 percent since 1990.

**Table 3-25: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from the Incineration of Waste (MMT CO<sub>2</sub> Eq.)**

Gas/Waste Product	1990	2005	2013	2014	2015	2016	2017
<b>CO<sub>2</sub></b>	<b>8.0</b>	<b>12.5</b>	<b>10.3</b>	<b>10.4</b>	<b>10.7</b>	<b>10.8</b>	<b>10.8</b>
Plastics	5.6	6.9	5.8	5.9	6.2	6.2	6.2
Synthetic Rubber in Tires	0.3	1.6	1.2	1.2	1.1	1.2	1.2
Carbon Black in Tires	0.4	2.0	1.4	1.4	1.4	1.4	1.4
Synthetic Rubber in MSW	0.9	0.8	0.7	0.7	0.7	0.7	0.7
Synthetic Fibers	0.8	1.2	1.2	1.2	1.3	1.3	1.3
<b>CH<sub>4</sub></b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>
<b>N<sub>2</sub>O</b>	<b>0.5</b>	<b>0.4</b>	<b>0.3</b>	<b>0.3</b>	<b>0.3</b>	<b>0.3</b>	<b>0.3</b>
<b>Total</b>	<b>8.4</b>	<b>12.9</b>	<b>10.6</b>	<b>10.7</b>	<b>11.1</b>	<b>11.1</b>	<b>11.1</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub> Eq.

**Table 3-26: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from the Incineration of Waste (kt)**

Gas/Waste Product	1990	2005	2013	2014	2015	2016	2017
<b>CO<sub>2</sub></b>	<b>7,950</b>	<b>12,469</b>	<b>10,333</b>	<b>10,429</b>	<b>10,742</b>	<b>10,765</b>	<b>10,790</b>
Plastics	5,588	6,919	5,823	5,928	6,184	6,184	6,184
Synthetic Rubber in Tires	308	1,599	1,158	1,154	1,149	1,160	1,171
Carbon Black in Tires	385	1,958	1,412	1,406	1,401	1,415	1,430
Synthetic Rubber in MSW	854	766	692	692	703	703	703
Synthetic Fibers	816	1,227	1,247	1,249	1,305	1,303	1,303
<b>CH<sub>4</sub></b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>
<b>N<sub>2</sub>O</b>	<b>2</b>	<b>1</b>	<b>1</b>	<b>1</b>	<b>1</b>	<b>1</b>	<b>1</b>

+ Does not exceed 0.5 kt.

## Methodology

Emissions of CO<sub>2</sub> from the incineration of waste include CO<sub>2</sub> generated by the incineration of plastics, synthetic fibers, and synthetic rubber in MSW, as well as the incineration of synthetic rubber and carbon black in scrap tires. The emission estimates are calculated for all four sources on a mass-basis based on the data available. These emissions were estimated by multiplying the mass of each material incinerated by the C content of the material and the fraction oxidized (98 percent). Plastics incinerated in MSW were categorized into seven plastic resin types, each material having a discrete C content. Similarly, synthetic rubber is categorized into three product types, and synthetic fibers were categorized into four product types, each having a discrete C content. Scrap tires contain several types of synthetic rubber, carbon black, and synthetic fibers. Each type of synthetic rubber has a discrete C content, and carbon black is 100 percent C. Emissions of CO<sub>2</sub> were calculated based on the amount of scrap tires

used for fuel and the synthetic rubber and carbon black content of scrap tires. More detail on the methodology for calculating emissions from each of these waste incineration sources is provided in Annex 3.7.

For each of the methods used to calculate CO<sub>2</sub> emissions from the incineration of waste, data on the quantity of product combusted and the C content of the product are needed. For plastics, synthetic rubber, and synthetic fibers in MSW, the amount of specific materials discarded as MSW (i.e., the quantity generated minus the quantity recycled) was taken from *Municipal Solid Waste Generation, Recycling, and Disposal in the United States: Facts and Figures* (EPA 2000 through 2003, 2005 through 2014), and *Advancing Sustainable Materials Management: Facts and Figures: Assessing Trends in Material Generation, Recycling and Disposal in the United States* (EPA 2015; EPA 2016; EPA 2018a) and detailed unpublished backup data for some years not shown in the reports (Schneider 2007). For 2012 through 2017, data on total waste incinerated were assumed to equal to the 2011 value from Shin (2014) for 2012 through 2017. For synthetic rubber and carbon black in scrap tires, information was obtained biannually from U.S. Scrap Tire Management Summary for 2005 through 2017 data (RMA 2018). Average C contents for the “Other” plastics category and synthetic rubber in MSW were calculated from 1998 and 2002 production statistics; C content for 1990 through 1998 is based on the 1998 value; C content for 1999 through 2001 is the average of 1998 and 2002 values; and C content for 2002 to date is based on the 2002 value. Carbon content for synthetic fibers was calculated from a weighted average of production statistics from 1990 to date. Information about scrap tire composition was taken from the Rubber Manufacturers’ Association internet site (RMA 2012a). The mass of incinerated material is multiplied by its C content to calculate the total amount of carbon stored.

The assumption that 98 percent of organic C is oxidized (which applies to all waste incineration categories for CO<sub>2</sub> emissions) was reported in EPA’s life cycle analysis of greenhouse gas emissions and sinks from management of solid waste (EPA 2006). This percentage is multiplied by the carbon stored to estimate the amount of carbon emitted.

Incineration of waste, including MSW, also results in emissions of CH<sub>4</sub> and N<sub>2</sub>O. These emissions were calculated as a function of the total estimated mass of waste incinerated and emission factors. As noted above, CH<sub>4</sub> and N<sub>2</sub>O emissions are a function of total waste incinerated in each year; for 1990 through 2008, these data were derived from the information published in *BioCycle* (van Haaren et al. 2010). Data for 2009 and 2010 were interpolated between 2008 and 2011 values. Data for 2011 were derived from Shin (2014). Data on total waste incinerated was not available in the *BioCycle* data set for 2012 through 2017, so these values were assumed to equal the 2011 *BioCycle* dataset value.

Table 3-27 provides data on MSW discarded and percentage combusted for the total waste stream. The emission factors of N<sub>2</sub>O and CH<sub>4</sub> emissions per quantity of MSW combusted are default emission factors for the default continuously-fed stoker unit MSW incineration technology type and were taken from IPCC (2006).

**Table 3-27: Municipal Solid Waste Generation (Metric Tons) and Percent Combusted (BioCycle dataset)**

Year	Waste Discarded	Waste Incinerated	Incinerated (% of Discards)
1990	235,733,657	30,632,057	13.0%
2005	259,559,787	25,973,520	10.0%
2013	273,116,704 <sup>a</sup>	20,756,879	7.6%
2014	273,116,704 <sup>a</sup>	20,756,879	7.6%
2015	273,116,704 <sup>a</sup>	20,756,879	7.6%
2016	273,116,704 <sup>a</sup>	20,756,879	7.6%
2017	273,116,704 <sup>a</sup>	20,756,879	7.6%

<sup>a</sup> Assumed equal to 2011 value.  
Source: van Haaren et al. (2010)

## Uncertainty and Time-Series Consistency

An Approach 2 Monte Carlo analysis was performed to determine the level of uncertainty surrounding the estimates of CO<sub>2</sub> emissions and N<sub>2</sub>O emissions from the incineration of waste (given the very low emissions for CH<sub>4</sub>, no uncertainty estimate was derived). IPCC Approach 2 analysis allows the specification of probability density functions for key variables within a computational structure that mirrors the calculation of the Inventory estimate. Uncertainty estimates and distributions for waste generation variables (i.e., plastics, synthetic rubber, and textiles generation) were obtained through a conversation with one of the authors of the Municipal Solid Waste in the United States reports. Statistical analyses or expert judgments of uncertainty were not available directly from the information sources for the other variables; thus, uncertainty estimates for these variables were determined using assumptions based on source category knowledge and the known uncertainty estimates for the waste generation variables.

The uncertainties in the waste incineration emission estimates arise from both the assumptions applied to the data and from the quality of the data. Key factors include MSW incineration rate; fraction oxidized; missing data on waste composition; average C content of waste components; assumptions on the synthetic/biogenic C ratio; and combustion conditions affecting N<sub>2</sub>O emissions. The highest levels of uncertainty surround the variables that are based on assumptions (e.g., percent of clothing and footwear composed of synthetic rubber); the lowest levels of uncertainty surround variables that were determined by quantitative measurements (e.g., combustion efficiency, C content of C black).

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-28. Waste incineration CO<sub>2</sub> emissions in 2017 were estimated to be between 9.6 and 12.4 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. This indicates a range of 11 percent below to 15 percent above the 2017 emission estimate of 10.8 MMT CO<sub>2</sub> Eq. Also at a 95 percent confidence level, waste incineration N<sub>2</sub>O emissions in 2017 were estimated to be between 0.2 and 1.2 MMT CO<sub>2</sub> Eq. This indicates a range of 47 percent below to 301 percent above the 2017 emission estimate of 0.3 MMT CO<sub>2</sub> Eq. Differences observed in comparison to last year were due to a reevaluation and refinement of assumptions on scrap tire weights of light and heavy-duty tires.

**Table 3-28: Approach 2 Quantitative Uncertainty Estimates for CO<sub>2</sub> and N<sub>2</sub>O from the Incineration of Waste (MMT CO<sub>2</sub> Eq. and Percent)**

Source	Gas	2017 Emission Estimate (MMT CO <sub>2</sub> Eq.)	Uncertainty Range Relative to Emission Estimate <sup>a</sup> (MMT CO <sub>2</sub> Eq.) (%)			
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Incineration of Waste	CO <sub>2</sub>	10.8	9.6	12.4	-11%	+15%
Incineration of Waste	N <sub>2</sub> O	0.3	0.2	1.2	-47%	+301%

<sup>a</sup> Range of emission estimates predicted by Monte Carlo Simulation for a 95 percent confidence interval.

## QA/QC and Verification

In order to ensure the quality of the emission estimates from waste incineration, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and specifically focused on the emission factor and activity data sources and methodology used for estimating emissions from incineration of waste. Trends across the time series were analyzed to determine whether any corrective actions were needed. Corrective actions were taken to rectify minor errors in use of activity data.

## Recalculations Discussion

The total generation and recovery of textiles in MSW in 2015 was updated to reflect the tonnage reported in the newest *Advancing Sustainable Materials Management: Facts and Figures: Assessing Trends in Material Generation, Recycling and Disposal in the United States* (EPA 2018a), which impacted CO<sub>2</sub> emissions from synthetic fibers.

## Planned Improvements

The waste incineration estimates have recently relied on MSW mass flow (i.e., tonnage) data that has not been updated since 2011. These values previously came from *BioCycle* (Shin 2014) and *EPA Facts and Figures* (EPA 2015). EPA performed an examination of facility-level MSW tonnage data availability, primarily focusing on EPA's GHGRP data, Energy Information Administration (EIA) waste-to-energy data, and other sources. EPA concluded that the GHGRP data were more complete (i.e., included more facilities), but did not contain data for all inventory years (1990 through 2016). The EIA data can be used to supplement years not available in the GHGRP dataset. In addition, the GHGRP data do not include specific waste components outside of an assumed biogenic and fossil component, which is necessary for CO<sub>2</sub> emission calculations. Data from EPA's GHGRP on fossil CO<sub>2</sub> emissions can be used to benchmark results for other waste components in the Inventory.

Additional improvements will focus on investigating new methods and sources for CO<sub>2</sub> emission estimates, and investigating new data sources for MSW incinerated values (i.e., tonnage) for estimating CO<sub>2</sub> and non-CO<sub>2</sub> (CH<sub>4</sub>, N<sub>2</sub>O) emissions.

Proposed improvements to the current CO<sub>2</sub> emissions estimation methodology include opportunities for either incorporating total CO<sub>2</sub> emissions from existing waste incineration datasets (i.e., EIA and GHGRP data that provide CO<sub>2</sub> emission estimates) or updating emission factors (i.e., MSW carbon content) and continuing to use the *Facts and Figures* disposal data for fossil-based products. Further research is required to compare the emission factors (i.e., MSW carbon content, heating values) used across waste incineration CO<sub>2</sub> emissions approaches, including the current Inventory, EIA, and EPA's GHGRP. In addition, the currently used *BioCycle* percent combusted assumption could be updated with *Facts and Figures* product tonnage combusted data.

Non-CO<sub>2</sub> improvements will focus on research of potential data sources for updating emission factors. EPA is also researching potential data sources for incinerated MSW tonnage that can be used for future inventory years instead of applying an incineration rate to generated MSW tonnage. EPA is analyzing the *Facts and Figures* non-tire MSW combusted tonnage and previously compiled EIA and GHGRP tonnage data to compare organic and non-organic components of these MSW tonnage data where available.

Additional improvements will be conducted to improve the transparency in the current reporting of waste incineration. Currently, hazardous industrial waste incineration is included within the overall calculations for the Carbon Emitted from Non-Energy Uses of Fossil Fuels source category. Waste incineration activities that do not include energy recovery will be examined. Synthetic fibers within scrap tires are not included in this analysis and will be explored for future Inventories. The C content of fibers within scrap tires will be used to calculate the associated incineration emissions. Updated fiber content data from the Fiber Economics Bureau will also be explored.

## 3.4 Coal Mining (CRF Source Category 1B1a)

Three types of coal mining-related activities release CH<sub>4</sub> to the atmosphere: underground mining, surface mining, and post-mining (i.e., coal-handling) activities. While surface mines account for the majority of U.S. coal production, underground coal mines contribute the largest share of CH<sub>4</sub> emissions (see Table 3-30 and Table 3-31) due to the higher CH<sub>4</sub> content of coal in the deeper underground coal seams. In 2017, 237 underground coal mines and 434 surface mines were operating in the United States (EIA 2018). In recent years the total number of active coal mines in the United States has declined. In 2017, the United States was the third largest coal producer in the world (702 MMT), after China (3,376 MMT) and India (730 MMT) (IEA 2018).

**Table 3-29: Coal Production (kt)**

Year	Underground		Surface		Total	
	Number of Mines	Production	Number of Mines	Production	Number of Mines	Production
1990	1,683	384,244	1,656	546,808	3,339	931,052
2005	586	334,398	789	691,448	1,398	1,025,846

2013	395	309,546	637	581,270	1,032	890,815
2014	345	321,783	613	583,974	958	905,757
2015	305	278,342	529	534,092	834	812,435
2016	251	228,707	439	431,285	690	659,991
2017	237	247,779	434	454,303	671	702,082

Underground mines liberate CH<sub>4</sub> from ventilation systems and from degasification systems. Ventilation systems pump air through the mine workings to dilute noxious gases and ensure worker safety; these systems can exhaust significant amounts of CH<sub>4</sub> to the atmosphere in low concentrations. Degasification systems are wells drilled from the surface or boreholes drilled inside the mine that remove large, often highly concentrated volumes of CH<sub>4</sub> before, during, or after mining. Some mines recover and use CH<sub>4</sub> generated from ventilation and degasification systems, thereby reducing emissions to the atmosphere.

Surface coal mines liberate CH<sub>4</sub> as the overburden is removed and the coal is exposed to the atmosphere. Methane emissions are normally a function of coal rank (a classification related to the percentage of carbon in the coal) and depth. Surface coal mines typically produce lower-rank coals and remove less than 250 feet of overburden, so their level of emissions is much lower than from underground mines.

In addition, CH<sub>4</sub> is released during post-mining activities, as the coal is processed, transported, and stored for use.

Total CH<sub>4</sub> emissions in 2017 were estimated to be 2,227 kt (55.7 MMT CO<sub>2</sub> Eq.), a decline of 42 percent since 1990 (see Table 3-30 and Table 3-31). In 2017, underground mines accounted for approximately 75 percent of total emissions, surface mines accounted for 13 percent, and post-mining activities accounted for 12 percent. In 2017, total CH<sub>4</sub> emissions from coal mining increased by three percent relative to the previous year. This increase was due to a modest increase in coal production and a slight reduction in CH<sub>4</sub> recovered and used. The amount of CH<sub>4</sub> recovered and used in 2017 decreased by approximately six percent compared to 2016 levels. This decrease is primarily attributed to a decrease in reported CH<sub>4</sub> recovery and use at three mines.

**Table 3-30: CH<sub>4</sub> Emissions from Coal Mining (MMT CO<sub>2</sub> Eq.)**

Activity	1990	2005	2013	2014	2015	2016	2017
Underground (UG) Mining	74.2	42.0	46.2	46.1	44.9	40.7	41.5
Liberated	80.8	59.7	64.5	63.0	61.2	57.0	56.7
Recovered & Used	(6.6)	(17.7)	(18.3)	(16.9)	(16.3)	(16.3)	(15.2)
Surface Mining	10.8	11.9	9.7	9.6	8.7	6.8	7.2
Post-Mining (UG)	9.2	7.6	6.6	6.7	5.8	4.8	5.3
Post-Mining (Surface)	2.3	2.6	2.1	2.1	1.9	1.5	1.6
<b>Total</b>	<b>96.5</b>	<b>64.1</b>	<b>64.6</b>	<b>64.6</b>	<b>61.2</b>	<b>53.8</b>	<b>55.7</b>

Notes: Totals may not sum due to independent rounding. Parentheses indicate negative values.

**Table 3-31: CH<sub>4</sub> Emissions from Coal Mining (kt)**

Activity	1990	2005	2013	2014	2015	2016	2017
UG Mining	2,968	1,682	1,849	1,844	1,796	1,629	1,662
Liberated	3,234	2,390	2,580	2,522	2,448	2,279	2,270
Recovered & Used	(266)	(708)	(730)	(677)	(652)	(650)	(608)
Surface Mining	430	475	388	386	347	273	290
Post-Mining (UG)	368	306	263	270	231	193	213
Post-Mining (Surface)	93	103	84	84	75	59	63
<b>Total</b>	<b>3,860</b>	<b>2,565</b>	<b>2,584</b>	<b>2,583</b>	<b>2,449</b>	<b>2,154</b>	<b>2,227</b>

Notes: Totals may not sum due to independent rounding. Parentheses indicate negative values.

## Methodology

The methodology for estimating CH<sub>4</sub> emissions from coal mining consists of two steps:

- Estimate emissions from underground mines. These emissions have two sources: ventilation systems and degasification systems. They are estimated using mine-specific data, then summed to determine total CH<sub>4</sub> liberated. The CH<sub>4</sub> recovered and used is then subtracted from this total, resulting in an estimate of net emissions to the atmosphere.



- Estimate CH<sub>4</sub> emissions from surface mines and post-mining activities. Unlike the methodology for underground mines, which uses mine-specific data, the methodology for estimating emissions from surface mines and post-mining activities consists of multiplying basin-specific coal production by basin-specific gas content and an emission factor.

## Step 1: Estimate CH<sub>4</sub> Liberated and CH<sub>4</sub> Emitted from Underground Mines

Underground mines generate CH<sub>4</sub> from ventilation systems and degasification systems. Some mines recover and use the liberated CH<sub>4</sub>, thereby reducing emissions to the atmosphere. Total CH<sub>4</sub> emitted from underground mines equals the CH<sub>4</sub> liberated from ventilation systems, plus the CH<sub>4</sub> liberated from degasification systems, minus the CH<sub>4</sub> recovered and used.

### *Step 1.1: Estimate CH<sub>4</sub> Liberated from Ventilation Systems*

To estimate CH<sub>4</sub> liberated from ventilation systems, EPA uses data collected through its Greenhouse Gas Reporting Program (GHGRP)<sup>71</sup> (subpart FF, “Underground Coal Mines”), data provided by the U.S. Mine Safety and Health Administration (MSHA), and occasionally data collected from other sources on a site-specific level (e.g., state gas production databases). Since 2011, the nation’s “gassiest” underground coal mines—those that liberate more than 36,500,000 actual cubic feet of CH<sub>4</sub> per year (about 17,525 MT CO<sub>2</sub> Eq.)—have been required to report to EPA’s GHGRP (EPA 2018).<sup>72</sup> Mines that report to EPA’s GHGRP must report quarterly measurements of CH<sub>4</sub> emissions from ventilation systems; they have the option of recording their own measurements, or using the measurements taken by MSHA as part of that agency’s quarterly safety inspections of all mines in the United States with detectable CH<sub>4</sub> concentrations.<sup>73</sup>

Since 2013, ventilation CH<sub>4</sub> emission estimates have been calculated based on both GHGRP data submitted by underground mines, and on quarterly measurement data obtained directly from MSHA for the remaining mines. The quarterly measurements are used to determine the average daily emission rate for the reporting year quarter. Because not all mines report under EPA’s GHGRP, the emissions of the mines that do not report must be calculated using MSHA data. The MSHA data also serves as a quality assurance tool for validating GHGRP data.

### *Step 1.2: Estimate CH<sub>4</sub> Liberated from Degasification Systems*

Particularly gassy underground mines also use degasification systems (e.g., wells or boreholes) to remove CH<sub>4</sub> before, during, or after mining. This CH<sub>4</sub> can then be collected for use or vented to the atmosphere. Nineteen mines used degasification systems in 2017, and the CH<sub>4</sub> removed through these systems was reported to EPA’s GHGRP under subpart FF (EPA 2018). Based on the weekly measurements reported to EPA’s GHGRP, degasification data summaries for each mine were added to estimate the CH<sub>4</sub> liberated from degasification systems. Twelve of the 19 mines with degasification systems had operational CH<sub>4</sub> recovery and use projects (see step 1.3 below), and EPA’s GHGRP reports show the remaining seven mines vented CH<sub>4</sub> from degasification systems to the atmosphere.<sup>74</sup>

Degasification data reported to EPA’s GHGRP by underground coal mines is the primary source of data used to develop estimates of CH<sub>4</sub> liberated from degasification systems. Data reported to EPA’s GHGRP were used exclusively to estimate CH<sub>4</sub> liberated from degasification systems at 15 of the 19 mines that used degasification systems in 2017.

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<sup>71</sup> In implementing improvements and integrating data from EPA’s GHGRP, EPA followed the latest guidance from the IPCC on the use of facility-level data in national inventories (IPCC 2011).

<sup>72</sup> Underground coal mines report to EPA under subpart FF of the GHGRP (40 CFR part 98). In 2017, 78 underground coal mines reported to the program.

<sup>73</sup> MSHA records coal mine CH<sub>4</sub> readings with concentrations of greater than 50 ppm (parts per million) CH<sub>4</sub>. Readings below this threshold are considered non-detectable.

<sup>74</sup> Several of the mines venting CH<sub>4</sub> from degasification systems use a small portion of the gas to fuel gob well blowers in remote locations where electricity is not available. However, this CH<sub>4</sub> use is not considered to be a formal recovery and use project.

For pre-mining wells, cumulative degasification volumes that occur prior to the well being mined through are attributed to the mine in the inventory year in which the well is mined through.<sup>75</sup> EPA’s GHGRP does not require gas production from virgin coal seams (coalbed methane) to be reported by coal mines under subpart FF.<sup>76</sup> Most pre-mining wells drilled from the surface are considered coalbed methane wells prior to mine-through and associated CH<sub>4</sub> emissions are reported under another subpart of the GHGRP (subpart W, “Petroleum and Natural Gas Systems”). As a result, GHGRP data must be supplemented to estimate cumulative degasification volumes that occurred prior to well mine-through. There were five mines with degasification systems that include pre-mining wells that were mined through in 2017. For four of these mines, GHGRP data were supplemented with historical data from state gas well production databases (GSA 2018; WVGES 2018), as well as with mine-specific information regarding the locations and dates on which the pre-mining wells were mined through (JWR 2010; El Paso 2009). For the remaining mine, data from a state gas well production database were used (DMME 2018).

### *Step 1.3: Estimate CH<sub>4</sub> Recovered from Ventilation and Degasification Systems, and Utilized or Destroyed (Emissions Avoided)*

Thirteen mines had CH<sub>4</sub> recovery and use projects in place in 2017. Twelve of these projects involved degasification systems and one involved a ventilation air methane abatement project (VAM). Eleven of these mines sold the recovered CH<sub>4</sub> to a pipeline, including one that also used CH<sub>4</sub> to fuel a thermal coal dryer. One mine used recovered CH<sub>4</sub> to heat mine ventilation air.

EPA’s GHGRP data was exclusively used to estimate the CH<sub>4</sub> recovered and used from eight of the 12 mines that deployed degasification systems in 2017. State sales data were used to estimate CH<sub>4</sub> recovered and used from the remaining four mines that deployed degasification systems in 2017 (DMME 2018; GSA 2018). Based on weekly measurements, the GHGRP degasification destruction data summaries for each mine were added together to estimate the CH<sub>4</sub> recovered and used from degasification systems. For the single mine that employed VAM for CH<sub>4</sub> recovery and use, the estimates of CH<sub>4</sub> recovered and used were obtained from the mine’s offset verification statement (OVS) submitted to the California Air Resources Board (CARB) (McElroy OVS 2018).

Of the 13 mines with CH<sub>4</sub> recovery in 2017, five intersected pre-mining wells in 2017. EPA’s GHGRP and supplemental data were used to estimate CH<sub>4</sub> recovered and used at these mines. Supplemental information was used for these mines because estimating CH<sub>4</sub> recovery and use from pre-mining wells requires additional data not reported under subpart FF of EPA’s GHGRP (see discussion in step 1.2 above) to account for the emissions avoided. The supplemental data came from state gas production databases as well as mine-specific information on the timing of mined-through pre-mining wells.

## **Step 2: Estimate CH<sub>4</sub> Emitted from Surface Mines and Post-Mining Activities**

Mine-specific data are not available for estimating CH<sub>4</sub> emissions from surface coal mines or for post-mining activities. For surface mines, basin-specific coal production obtained from the Energy Information Administration’s *Annual Coal Report* (EIA 2018) was multiplied by basin-specific CH<sub>4</sub> contents (EPA 1996, 2005) and a 150 percent emission factor (to account for CH<sub>4</sub> from over- and under-burden) to estimate CH<sub>4</sub> emissions (King 1994; Saghafi 2013). For post-mining activities, basin-specific coal production was multiplied by basin-specific gas contents and a mid-range 32.5 percent emission factor for CH<sub>4</sub> desorption during coal transportation and storage (Creedy 1993). Basin-specific in situ gas content data were compiled from AAPG (1984) and USBM (1986).

## **Uncertainty and Time-Series Consistency**

A quantitative uncertainty analysis was conducted for the coal mining source category using the IPCC-recommended Approach 2 uncertainty estimation methodology. Because emission estimates from underground ventilation systems were based on actual measurement data from EPA’s GHGRP or from MSHA, uncertainty is relatively low. A degree of imprecision was introduced because the ventilation air measurements used were not

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<sup>75</sup> A well is “mined through” when coal mining development or the working face intersects the borehole or well.

<sup>76</sup> This applies for pre-drainage in years prior to the well being mined through. Beginning with the year the well is mined through, the annual volume of CH<sub>4</sub> liberated from a pre-drainage well is reported under subpart FF of EPA’s GHGRP.

continuous but rather quarterly instantaneous readings that were used to determine the average daily emission rate for the quarter. Additionally, the measurement equipment used can be expected to have resulted in an average of 10 percent overestimation of annual CH<sub>4</sub> emissions (Mutmansky & Wang 2000). Equipment measurement uncertainty is applied to both GHGRP and MSHA data.

Estimates of CH<sub>4</sub> liberated and recovered by degasification systems are relatively certain for utilized CH<sub>4</sub> because of the availability of EPA’s GHGRP data and gas sales information. Many of the liberation and recovery estimates use data on wells within 100 feet of a mined area. However, uncertainty exists concerning the radius of influence of each well. The number of wells counted, and thus the liberated CH<sub>4</sub> and avoided emissions may vary if the drainage area is found to be larger or smaller than estimated.

EPA’s GHGRP requires weekly CH<sub>4</sub> monitoring of mines that report degasification systems, and continuous CH<sub>4</sub> monitoring is required for CH<sub>4</sub> utilized on- or off-site. Since 2012, GHGRP data have been used to estimate CH<sub>4</sub> emissions from vented degasification wells, reducing the uncertainty associated with prior MSHA estimates used for this sub-source. Beginning in 2013, GHGRP data were also used for determining CH<sub>4</sub> recovery and use at mines without publicly available gas usage or sales records, which has reduced the uncertainty from previous estimation methods that were based on information from coal industry contacts.

In 2015, 2016, and 2017 a small level of uncertainty was introduced by using estimated rather than measured values of recovered CH<sub>4</sub> from two of the mines with degasification systems. An increased level of uncertainty was applied to these two sub-sources, but the change had little impact on the overall uncertainty.

Surface mining and post-mining emissions are associated with considerably more uncertainty than underground mines, because of the difficulty in developing accurate emission factors from field measurements. However, since underground emissions constitute the majority of total coal mining emissions, the uncertainty associated with underground emissions is the primary factor that determines overall uncertainty.

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-32. Coal mining CH<sub>4</sub> emissions in 2017 were estimated to be between 50.5 and 66.2 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. This indicates a range of 9.3 percent below to 18.8 percent above the 2017 emission estimate of 55.7 MMT CO<sub>2</sub> Eq.

**Table 3-32: Approach 2 Quantitative Uncertainty Estimates for CH<sub>4</sub> Emissions from Coal Mining (MMT CO<sub>2</sub> Eq. and Percent)**

Source	Gas	2017 Emission Estimate (MMT CO <sub>2</sub> Eq.)	Uncertainty Range Relative to Emission Estimate <sup>a</sup>			
			Lower Bound	Upper Bound	Lower Bound (%)	Upper Bound (%)
Coal Mining	CH <sub>4</sub>	55.7	50.5	66.2	-9.3%	+18.8%

<sup>a</sup> Range of emission estimates predicted by Monte Carlo stochastic simulation for a 95 percent confidence interval.

## QA/QC and Verification

In order to ensure the quality of the emission estimates for coal mining, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and reported emissions data used for estimating emissions from coal mining. Trends across the time series were analyzed to determine whether any corrective actions were needed.

Emission estimates for coal mining rely in large part on data reported by coal mines to EPA’s GHGRP. EPA verifies annual facility-level reports through a multi-step process to identify potential errors and ensure that data submitted to EPA are accurate, complete, and consistent. All reports submitted to EPA are evaluated by electronic validation and verification checks. If potential errors are identified, EPA will notify the reporter, who can resolve the issue either by providing an acceptable response describing why the flagged issue is not an error or by correcting the flagged issue and resubmitting their annual greenhouse gas report. Additional QA/QC and verification procedures occur for each GHGRP subpart.

## Recalculations Discussion

For the current Inventory, minor revisions were made to the 2016 annual coal production quantities for underground and surface mines. These revisions to the 2016 activity data were based on the EIA 2017 *Annual Coal Report* (EIA 2018). The revisions to the underground coal production quantities resulted in an emission increase of approximately 0.5 percent for the 2016 emissions from post-mining activities for underground mining. The revisions to the surface coal production quantities resulted in an insignificant increase (0.001 percent) for the 2016 emissions from surface mining and post-surface mining activities.

### 3.5 Abandoned Underground Coal Mines (CRF Source Category 1B1a)

Underground coal mines contribute the largest share of coal mine methane (CMM) emissions, with active underground mines the leading source of underground emissions. However, mines also continue to release CH<sub>4</sub> after closure. As mines mature and coal seams are mined through, mines are closed and abandoned. Many are sealed and some flood through intrusion of groundwater or surface water into the void. Shafts or portals are generally filled with gravel and capped with a concrete seal, while vent pipes and boreholes are plugged in a manner similar to oil and gas wells. Some abandoned mines are vented to the atmosphere to prevent the buildup of CH<sub>4</sub> that may find its way to surface structures through overburden fractures. As work stops within the mines, CH<sub>4</sub> liberation decreases but it does not stop completely. Following an initial decline, abandoned mines can liberate CH<sub>4</sub> at a near-steady rate over an extended period of time, or if flooded, produce gas for only a few years. The gas can migrate to the surface through the conduits described above, particularly if they have not been sealed adequately. In addition, diffuse emissions can occur when CH<sub>4</sub> migrates to the surface through cracks and fissures in the strata overlying the coal mine. The following factors influence abandoned mine emissions:

- Time since abandonment;
- Gas content and adsorption characteristics of coal;
- CH<sub>4</sub> flow capacity of the mine;
- Mine flooding;
- Presence of vent holes; and
- Mine seals.

Annual gross abandoned mine CH<sub>4</sub> emissions ranged from 7.2 to 10.8 MMT CO<sub>2</sub> Eq. from 1990 through 2017, varying, in general, by less than one percent to approximately 19 percent from year to year. Fluctuations were due mainly to the number of mines closed during a given year, as well as the magnitude of the emissions from those mines when active. Gross abandoned mine CH<sub>4</sub> emissions peaked in 1996 (10.8 MMT CO<sub>2</sub> Eq.) due to the large number of gassy mine<sup>77</sup> closures from 1994 to 1996 (72 gassy mines closed during the three-year period). In spite of this rapid rise, abandoned mine CH<sub>4</sub> emissions have been generally on the decline since 1996. Since 2002, there have been fewer than twelve gassy mine closures each year. There were no gassy mine closures in 2017. In 2017, gross abandoned mine emissions decreased slightly from 9.5 to 9.2 MMT CO<sub>2</sub> Eq. (see Table 3-33 and Table 3-34). Gross emissions are reduced by CH<sub>4</sub> recovered and used at 45 mines, resulting in net emissions in 2017 of 6.4 MMT CO<sub>2</sub> Eq.

**Table 3-33: CH<sub>4</sub> Emissions from Abandoned Coal Mines (MMT CO<sub>2</sub> Eq.)**

Activity	1990	2005	2013	2014	2015	2016	2017
Abandoned Underground Mines	7.2	8.4	8.8	8.7	9.0	9.5	9.2
Recovered & Used	+	(1.8)	(2.6)	(2.4)	(2.6)	(2.8)	(2.7)
<b>Total</b>	<b>7.2</b>	<b>6.6</b>	<b>6.2</b>	<b>6.3</b>	<b>6.4</b>	<b>6.7</b>	<b>6.4</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub> Eq.

<sup>77</sup> A mine is considered a “gassy” mine if it emits more than 100 thousand cubic feet of CH<sub>4</sub> per day (100 mcf/d).

Notes: Totals may not sum due to independent rounding. Parentheses indicate negative values.

**Table 3-34: CH<sub>4</sub> Emissions from Abandoned Coal Mines (kt)**

Activity	1990	2005	2013	2014	2015	2016	2017
Abandoned Underground Mines	288	334	353	350	359	380	367
Recovered & Used	+	(70)	(104)	(97)	(102)	(112)	(109)
<b>Total</b>	<b>288</b>	<b>264</b>	<b>249</b>	<b>253</b>	<b>256</b>	<b>268</b>	<b>257</b>

+ Does not exceed 0.5 kt

Notes: Totals may not sum due to independent rounding. Parentheses indicate negative values.

## Methodology

Estimating CH<sub>4</sub> emissions from an abandoned coal mine requires predicting the emissions of a mine from the time of abandonment through the inventory year of interest. The flow of CH<sub>4</sub> from the coal to the mine void is primarily dependent on the mine's emissions when active and the extent to which the mine is flooded or sealed. The CH<sub>4</sub> emission rate before abandonment reflects the gas content of the coal, the rate of coal mining, and the flow capacity of the mine in much the same way as the initial rate of a water-free conventional gas well reflects the gas content of the producing formation and the flow capacity of the well. A well or a mine that produces gas from a coal seam and the surrounding strata will produce less gas through time as the reservoir of gas is depleted. Depletion of a reservoir will follow a predictable pattern depending on the interplay of a variety of natural physical conditions imposed on the reservoir. The depletion of a reservoir is commonly modeled by mathematical equations and mapped as a type curve. Type curves, which are referred to as decline curves, have been developed for abandoned coal mines. Existing data on abandoned coal mine emissions through time, although sparse, appear to fit the hyperbolic type of decline curve used in forecasting production from natural gas wells.

In order to estimate CH<sub>4</sub> emissions over time for a given abandoned mine, it is necessary to apply a decline function, initiated upon abandonment, to that mine. In the analysis, mines were grouped by coal basin with the assumption that they will generally have the same initial pressures, permeability and isotherm. As CH<sub>4</sub> leaves the system, the reservoir pressure (Pr) declines as described by the isotherm's characteristics. The emission rate declines because the mine pressure (Pw) is essentially constant at atmospheric pressure for a vented mine, and the productivity index (PI), which is expressed as the flow rate per unit of pressure change, is essentially constant at the pressures of interest (atmospheric to 30 psia). The CH<sub>4</sub> flow rate is determined by the laws of gas flow through porous media, such as Darcy's Law. A rate-time equation can be generated that can be used to predict future emissions. This decline through time is hyperbolic in nature and can be empirically expressed as:

$$q = q_i (1 + bD_i t)^{-1/b}$$

where,

q	=	Gas flow rate at time t in million cubic feet per day (mmcf/d)
q <sub>i</sub>	=	Initial gas flow rate at time zero (t <sub>0</sub> ), mmcf/d
b	=	The hyperbolic exponent, dimensionless
D <sub>i</sub>	=	Initial decline rate, 1/year
t	=	Elapsed time from t <sub>0</sub> (years)

This equation is applied to mines of various initial emission rates that have similar initial pressures, permeability and adsorption isotherms (EPA 2004).

The decline curves created to model the gas emission rate of coal mines must account for factors that decrease the rate of emissions after mining activities cease, such as sealing and flooding. Based on field measurement data, it was assumed that most U.S. mines prone to flooding will become completely flooded within eight years and therefore will no longer have any measurable CH<sub>4</sub> emissions. Based on this assumption, an average decline rate for flooded mines was established by fitting a decline curve to emissions from field measurements. An exponential equation was developed from emissions data measured at eight abandoned mines known to be filling with water located in two of the five basins. Using a least squares, curve-fitting algorithm, emissions data were matched to the exponential

equation shown below. There was not enough data to establish basin-specific equations as was done with the vented, non-flooding mines (EPA 2004).

$$q = q_i e^{(-Dt)}$$

where,

q	=	Gas flow rate at time t in mmcf/d
q <sub>i</sub>	=	Initial gas flow rate at time zero (t <sub>0</sub> ), mmcf/d
D	=	Decline rate, 1/year
t	=	Elapsed time from t <sub>0</sub> (years)

Seals have an inhibiting effect on the rate of flow of CH<sub>4</sub> into the atmosphere compared to the flow rate that would exist if the mine had an open vent. The total volume emitted will be the same, but emissions will occur over a longer period of time. The methodology, therefore, treats the emissions prediction from a sealed mine similarly to the emissions prediction from a vented mine, but uses a lower initial rate depending on the degree of sealing. A computational fluid dynamics simulator was used with the conceptual abandoned mine model to predict the decline curve for inhibited flow. The percent sealed is defined as 100 × (1 – [initial emissions from sealed mine / emission rate at abandonment prior to sealing]). Significant differences are seen between 50 percent, 80 percent and 95 percent closure. These decline curves were therefore used as the high, middle, and low values for emissions from sealed mines (EPA 2004).

For active coal mines, those mines producing over 100 thousand cubic feet per day (mcf/d) of CH<sub>4</sub> account for about 98 percent of all CH<sub>4</sub> emissions. This same relationship is assumed for abandoned mines. It was determined that the 532 abandoned mines closed after 1972 produced CH<sub>4</sub> emissions greater than 100 mcf/d when active. Further, the status of 304 of the 532 mines (or 57 percent) is known to be either: 1) vented to the atmosphere; 2) sealed to some degree (either earthen or concrete seals); or 3) flooded (enough to inhibit CH<sub>4</sub> flow to the atmosphere). The remaining 43 percent of the mines whose status is unknown were placed in one of these three categories by applying a probability distribution analysis based on the known status of other mines located in the same coal basin (EPA 2004).

**Table 3-35: Number of Gassy Abandoned Mines Present in U.S. Basins in 2017, Grouped by Class According to Post-Abandonment State**

Basin	Sealed	Vented	Flooded	Total		Total Mines
				Known	Unknown	
Central Appl.	40	26	52	118	148	266
Illinois	34	3	14	51	31	82
Northern Appl.	47	22	16	85	39	124
Warrior Basin	0	0	16	16	0	16
Western Basins	28	4	2	34	10	44
<b>Total</b>	<b>149</b>	<b>55</b>	<b>100</b>	<b>304</b>	<b>228</b>	<b>532</b>

Inputs to the decline equation require the average CH<sub>4</sub> emission rate prior to abandonment and the date of abandonment. Generally, these data are available for mines abandoned after 1971; however, such data are largely unknown for mines closed before 1972. Information that is readily available, such as coal production by state and county, is helpful but does not provide enough data to directly employ the methodology used to calculate emissions from mines abandoned before 1972. It is assumed that pre-1972 mines are governed by the same physical, geologic, and hydrologic constraints that apply to post-1971 mines; thus, their emissions may be characterized by the same decline curves.

During the 1970s, 78 percent of CH<sub>4</sub> emissions from coal mining came from seventeen counties in seven states. Mine closure dates were obtained for two of these states, Colorado and Illinois, for the hundred-year period extending from 1900 through 1999. The data were used to establish a frequency of mine closure histogram (by decade) and applied to the other five states with gassy mine closures. As a result, basin-specific decline curve equations were applied to the 145 gassy coal mines estimated to have closed between 1920 and 1971 in the United States, representing 78 percent of the emissions. State-specific, initial emission rates were used based on average coal mine CH<sub>4</sub> emissions rates during the 1970s (EPA 2004).

Abandoned mine emission estimates are based on all closed mines known to have active mine CH<sub>4</sub> ventilation emission rates greater than 100 mcf/d at the time of abandonment. For example, for 1990 the analysis included 145 mines closed before 1972 and 258 mines closed between 1972 and 1990. Initial emission rates are based on MSHA reports, time of abandonment, and basin-specific decline curves (MSHA 2018). Coal mine degasification data are not available for years prior to 1990, thus the initial emission rates used reflect ventilation emissions only for pre-1990 closures. CH<sub>4</sub> degasification amounts were added to the quantity of CH<sub>4</sub> vented to determine the total CH<sub>4</sub> liberation rate for all mines that closed between 1992 and 2017. Because the available pre-1972 mine data are assumed to account for 78 percent of the pre-1972 abandoned mine CH<sub>4</sub> emissions, the modeled results were multiplied by 1.22 to account for all U.S. pre-1972 abandoned mine emissions. The post-1971 mine data are assumed to represent 98 percent of the post-1971 abandoned mine CH<sub>4</sub> emissions, and therefore the modeled results were multiplied by 1.02 to account for all U.S. post-1971 abandoned mine emissions.

From 1993 through 2017, emission totals were downwardly adjusted to reflect CH<sub>4</sub> emissions avoided from those abandoned mines with CH<sub>4</sub> recovery and use or destruction systems. The Inventory totals were not adjusted for abandoned mine CH<sub>4</sub> emission reductions from 1990 through 1992, because no data was reported for abandoned coal mine CH<sub>4</sub> recovery and use or destruction projects during that time.

## Uncertainty and Time-Series Consistency

A quantitative uncertainty analysis was conducted to estimate the uncertainty surrounding the estimates of CH<sub>4</sub> emissions from abandoned underground coal mines. The uncertainty analysis provides for the specification of probability density functions for key variables within a computational structure that mirrors the calculation of the inventory estimate. The results provide the range within which, with 95 percent certainty, emissions from this source category are likely to fall.

As discussed above, the parameters for which values must be estimated for each mine in order to predict its decline curve are: 1) the coal's adsorption isotherm; 2) CH<sub>4</sub> flow capacity as expressed by permeability; and 3) pressure at abandonment. Because these parameters are not available for each mine, a methodological approach to estimating emissions was used that generates a probability distribution of potential outcomes based on the most likely value and the probable range of values for each parameter. The range of values is not meant to capture the extreme values, but rather values that represent the highest and lowest quartile of the cumulative probability density function of each parameter. Once the low, mid, and high values are selected, they are applied to a probability density function.

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-36. Annual abandoned coal mine CH<sub>4</sub> emissions in 2017 were estimated to be between 5.1 and 7.7 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. This indicates a range of 21 percent below to 19 percent above the 2017 emission estimate of 6.4 MMT CO<sub>2</sub> Eq. One of the reasons for the relatively narrow range is that mine-specific data is available for use in the methodology for mines closed after 1972. Emissions from mines closed prior to 1972 have the largest degree of uncertainty because no mine-specific CH<sub>4</sub> liberation rates exist.

**Table 3-36: Approach 2 Quantitative Uncertainty Estimates for CH<sub>4</sub> Emissions from Abandoned Underground Coal Mines (MMT CO<sub>2</sub> Eq. and Percent)**

Source	Gas	2017 Emission Estimate (MMT CO <sub>2</sub> Eq.)	Uncertainty Range Relative to Emission Estimate <sup>a</sup> (MMT CO <sub>2</sub> Eq.)			
			Lower Bound	Upper Bound	Lower Bound (%)	Upper Bound (%)
Abandoned Underground Coal Mines	CH <sub>4</sub>	6.4	5.1	7.7	-21%	+19%

<sup>a</sup> Range of emission estimates predicted by Monte Carlo Simulation for a 95 percent confidence interval.

## QA/QC and Verification

In order to ensure the quality of the emission estimates for abandoned coal mines, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focused on the emission factor and activity data sources and methodology used for estimating

emissions from abandoned coal mines. Trends across the time series were analyzed to determine whether any corrective actions were needed.

## Recalculations Discussion

No recalculations were applied to the abandoned underground coal mine emission estimates for 1990 through 2017.

### 3.6 Petroleum Systems (CRF Source Category 1B2a)

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Methane emissions from petroleum systems are primarily associated with onshore and offshore crude oil production, transportation, and refining operations. During these activities, CH<sub>4</sub> is released to the atmosphere as leak emissions, vented emissions (including emissions from operational upsets) and emissions from flaring. Carbon dioxide emissions from petroleum systems are primarily associated with crude oil production and refining operations. Total CH<sub>4</sub> emissions from petroleum systems in 2017 were 37.7 MMT CO<sub>2</sub> Eq. (1,506 kt), a decrease of 10 percent from 1990. Total CO<sub>2</sub> emissions from petroleum systems in 2017 were 23.3 MMT CO<sub>2</sub> Eq. (23,336 kt), an increase of 161 percent from 1990. Total N<sub>2</sub>O emissions from petroleum systems in 2017 were 0.02 MMT CO<sub>2</sub> Eq. (0.08 kt), an increase of 77 percent from 1990.

Each year, some estimates in the Inventory are recalculated with improved methods and/or data. These improvements are implemented consistently across the previous Inventory's time series (i.e., 1990 to 2016) to ensure that the trend is accurate. Recalculations in petroleum systems in this year's Inventory include:

- Revised hydraulically fractured (HF) oil well completions and workovers methodology
- Newly calculated N<sub>2</sub>O emissions from flaring
- Newly calculated CO<sub>2</sub> emissions from crude oil transportation
- Recalculations due to GHGRP submission revisions

The Recalculations Discussion section below provides more details on the updated methods.

*Exploration.* Exploration includes well drilling, testing, and completions. Exploration accounts for approximately 1 percent of total CH<sub>4</sub> emissions from petroleum systems. The predominant sources of emissions from exploration are hydraulically fractured oil well completions and well testing. Other sources include well completions without hydraulic fracturing, and well drilling. Since 1990, exploration CH<sub>4</sub> emissions have decreased 88 percent, and while the number of hydraulically fractured wells completed increased by a factor of nearly 3, there were decreases in the fraction of such completions without reduced emissions completions (RECs) or flaring (from 90 percent in 1990 to 2 percent in 2017). Emissions of CH<sub>4</sub> from exploration were highest in 2012, over 20 times higher than in 2017, and lowest in 2017. Emissions of CH<sub>4</sub> from exploration decreased 24 percent from 2016 to 2017. Exploration accounts for 7 percent of total CO<sub>2</sub> emissions from petroleum systems in 2017. Emissions of CO<sub>2</sub> from exploration in 2017 increased by a factor of 4.2 from 1990, and 38 percent from 2016, due to an increase in hydraulically fractured oil well completions with flaring (from 10 percent of completions in 1990 to 58 percent in 2017). Emissions of CO<sub>2</sub> from exploration were highest in 2014, around 1.8 times as high as in 2017. Exploration accounts for 3 percent of total N<sub>2</sub>O emissions from petroleum systems in 2017. Emissions of N<sub>2</sub>O from exploration in 2017 increased by a factor of 3.4 from 1990, and 22 percent from 2016, due to an increase in hydraulically fractured oil well completions with flaring (from 10 percent of completions in 1990 to 58 percent in 2017).

*Production.* Production accounts for approximately 97 percent of total CH<sub>4</sub> emissions from petroleum systems. The predominant sources of emissions from production field operations are pneumatic controllers, offshore oil platforms, gas engines, chemical injection pumps, leaks from oil wellheads, and oil tanks. These six sources together account for 91 percent of the CH<sub>4</sub> emissions from production. Since 1990, CH<sub>4</sub> emissions from production have decreased by 5 percent, due to decreases in emissions from tanks, hydraulically fractured oil well workovers, and offshore platforms. Overall, production segment methane emissions decreased by 1 percent from 2016 levels. Production emissions account for 77 percent of the total CO<sub>2</sub> emissions from petroleum systems in 2017. The principal sources of CO<sub>2</sub> emissions are associated gas flaring, oil tanks with flares, and miscellaneous production flaring. These three



sources together account for 98 percent of the CO<sub>2</sub> emissions from production. Since 1990, CO<sub>2</sub> emissions from production have increased by 236 percent, due to increases in flaring emissions from associated gas flaring, tanks, and miscellaneous production flaring. Overall, production segment CO<sub>2</sub> emissions increased by 6 percent from 2016 levels due to an increase in associated gas flaring and miscellaneous production flaring. Production emissions account for 52 percent of the total N<sub>2</sub>O emissions from petroleum systems. The principal sources of N<sub>2</sub>O emissions are associated gas flaring, oil tanks with flares, and miscellaneous production flaring. Since 1990, N<sub>2</sub>O emissions from production have increased by 186 percent.

*Crude Oil Transportation.* Crude oil transportation activities account for less than 1 percent of total CH<sub>4</sub> emissions from petroleum systems. Emissions from tanks, marine loading, and truck loading operations account for 73 percent of CH<sub>4</sub> emissions from crude oil transportation. Since 1990, CH<sub>4</sub> emissions from transportation have increased by 17 percent. Methane emissions from transportation in 2017 decreased 5 percent from 2016 levels. Crude oil transportation activities account for less than 0.01 percent of total CO<sub>2</sub> emissions from petroleum systems. Emissions from tanks, marine loading, and truck loading operations account for 73 percent of CO<sub>2</sub> emissions from crude oil transportation. Emissions from crude oil transportation account for a very small percentage of the total emissions from petroleum systems and have little impact on the overall emissions.

*Crude Oil Refining.* Crude oil refining processes and systems account for 2 percent of total CH<sub>4</sub> emissions from petroleum systems. This low share is because most of the CH<sub>4</sub> in crude oil is removed or escapes before the crude oil is delivered to the refineries. There is an insignificant amount of CH<sub>4</sub> in all refined products. Within refineries, flaring accounts for 41 percent of the CH<sub>4</sub> emissions, while uncontrolled blowdowns and process vents account for 19 and 16 percent, respectively. Methane emissions from refining of crude oil have increased by 16 percent since 1990, and increased less than 1 percent since 2016; however, similar to the transportation subcategory, this increase has had little effect on the overall emissions of CH<sub>4</sub>. Crude oil refining processes and systems account for 16 percent of total CO<sub>2</sub> emissions from petroleum systems. Almost all (about 97 percent) of the CO<sub>2</sub> from refining is from flaring. Refinery CO<sub>2</sub> emissions increased by 14 percent from 1990 to 2017, and decreased by 7 percent from 2016 levels. Flaring occurring at crude oil refining processes and systems accounts for 45 percent of total N<sub>2</sub>O emissions from the oil industry. Refinery N<sub>2</sub>O emissions increased by 19 percent from 1990 to 2017, and decreased by 6 percent from 2016 levels.

**Table 3-37: CH<sub>4</sub> Emissions from Petroleum Systems (MMT CO<sub>2</sub> Eq.)**

Activity	1990	2005	2013	2014	2015	2016	2017
<b>Exploration<sup>a</sup></b>	<b>3.0</b>	<b>4.5</b>	<b>6.3</b>	<b>5.0</b>	<b>2.1</b>	<b>0.5</b>	<b>0.4</b>
<b>Production (Total)</b>	<b>38.3</b>	<b>31.4</b>	<b>34.4</b>	<b>36.2</b>	<b>36.5</b>	<b>36.8</b>	<b>36.4</b>
Pneumatic Controllers	19.3	17.5	18.6	19.4	19.6	20.5	20.9
Offshore Platforms	5.3	4.6	4.7	4.7	4.7	4.7	4.7
Equipment Leaks <sup>b</sup>	2.2	2.2	2.6	2.7	2.7	2.6	2.5
Gas Engines	2.1	1.7	2.2	2.3	2.3	2.2	2.2
Chemical Injection Pumps	1.2	1.7	2.1	2.2	2.2	2.1	2.0
Tanks	5.4	1.5	1.3	1.6	1.7	2.6	1.5
Other Sources	2.6	2.1	2.9	3.3	3.3	2.2	2.5
<b>Crude Oil Transportation</b>	<b>0.2</b>	<b>0.1</b>	<b>0.2</b>	<b>0.2</b>	<b>0.2</b>	<b>0.2</b>	<b>0.2</b>
<b>Refining</b>	<b>0.6</b>	<b>0.7</b>	<b>0.7</b>	<b>0.7</b>	<b>0.7</b>	<b>0.7</b>	<b>0.7</b>
<b>Total</b>	<b>42.1</b>	<b>36.7</b>	<b>41.6</b>	<b>42.1</b>	<b>39.5</b>	<b>38.2</b>	<b>37.7</b>

<sup>a</sup> Exploration includes well drilling, testing, and completions.

<sup>b</sup> Includes leak emissions from wellheads, separators, heaters/treaters, and headers.

Note: Totals may not sum due to independent rounding.

**Table 3-38: CH<sub>4</sub> Emissions from Petroleum Systems (kt CH<sub>4</sub>)**

Activity	1990	2005	2013	2014	2015	2016	2017
<b>Exploration<sup>a</sup></b>	<b>121</b>	<b>181</b>	<b>254</b>	<b>201</b>	<b>84</b>	<b>19</b>	<b>14</b>
<b>Production (Total)</b>	<b>1,531</b>	<b>1,255</b>	<b>1,377</b>	<b>1,446</b>	<b>1,458</b>	<b>1,473</b>	<b>1,456</b>
Pneumatic Controllers	774	701	743	777	786	818	837
Offshore Platforms	211	185	188	188	188	188	188
Equipment Leaks	88	86	104	109	107	104	102
Gas Engines	86	70	88	93	93	89	89
Chemical Injection Pumps	49	68	84	87	86	83	82
Tanks	218	60	53	63	68	102	61
Other Sources	105	86	118	131	131	89	98
<b>Crude Oil Transportation</b>	<b>7</b>	<b>5</b>	<b>7</b>	<b>8</b>	<b>8</b>	<b>8</b>	<b>8</b>
<b>Refining</b>	<b>24</b>	<b>28</b>	<b>27</b>	<b>26</b>	<b>28</b>	<b>28</b>	<b>28</b>
<b>Total</b>	<b>1,682</b>	<b>1,469</b>	<b>1,665</b>	<b>1,682</b>	<b>1,579</b>	<b>1,528</b>	<b>1,506</b>

<sup>a</sup> Exploration includes well drilling, testing, and completions.

Note: Totals may not sum due to independent rounding.

**Table 3-39: CO<sub>2</sub> Emissions from Petroleum Systems (MMT CO<sub>2</sub>)**

Activity	1990	2005	2013	2014	2015	2016	2017
Exploration	0.3	0.3	2.5	3.0	2.2	1.2	1.7
Production	5.3	7.5	19.1	23.2	25.4	17.0	18.0
Transportation	+	+	+	+	+	+	+
Crude Refining	3.3	3.7	3.6	3.4	4.1	4.0	3.7
<b>Total</b>	<b>9.0</b>	<b>11.6</b>	<b>25.1</b>	<b>29.6</b>	<b>31.7</b>	<b>22.2</b>	<b>23.3</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub>.

Note: Totals may not sum due to independent rounding.

**Table 3-40: CO<sub>2</sub> Emissions from Petroleum Systems (kt CO<sub>2</sub>)**

Activity	1990	2005	2013	2014	2015	2016	2017
Exploration	321	331	2,461	2,976	2,167	1,200	1,657
Production	5,344	7,493	19,059	23,201	25,438	17,008	17,951
Transportation	0.9	0.7	1.0	1.2	1.2	1.1	1.1
Crude Refining	3,284	3,728	3,609	3,419	4,067	3,991	3,728
<b>Total</b>	<b>8,950</b>	<b>11,552</b>	<b>25,130</b>	<b>29,597</b>	<b>31,672</b>	<b>22,200</b>	<b>23,336</b>

Note: Totals may not sum due to independent rounding.

**Table 3-41: N<sub>2</sub>O Emissions from Petroleum Systems (metric tons CO<sub>2</sub> Eq.)**

Activity	1990	2005	2013	2014	2015	2016	2017
Exploration	172	176	1,278	1,543	1,125	618	754
Production	4,414	5,332	12,980	15,817	17,429	12,749	12,640
Transportation	NE	NE	NE	NE	NE	NE	NE
Crude Refining	9,143	10,377	10,187	9,659	11,656	11,575	10,836
<b>Total</b>	<b>13,728</b>	<b>15,885</b>	<b>24,445</b>	<b>27,020</b>	<b>30,210</b>	<b>24,942</b>	<b>24,231</b>

NE (Not Estimated)

Note: Totals may not sum due to independent rounding.

**Table 3-42: N<sub>2</sub>O Emissions from Petroleum Systems (metric tons N<sub>2</sub>O)**

Activity	1990	2005	2013	2014	2015	2016	2017
Exploration	0.6	0.6	4.3	5.2	3.8	2.1	2.5
Production	14.8	17.9	43.6	53.1	58.5	42.8	42.4
Transportation	NE	NE	NE	NE	NE	NE	NE
Crude Refining	30.7	34.8	34.2	32.4	39.1	38.8	36.4
<b>Total</b>	<b>46.1</b>	<b>53.3</b>	<b>82.0</b>	<b>90.7</b>	<b>101.4</b>	<b>83.7</b>	<b>81.3</b>

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NE (Not Estimated)

Note: Totals may not sum due to independent rounding.

## Methodology

See Annex 3.5 for the full time series of emissions data, activity data, and emission factors, and additional information on methods and data sources.

Petroleum systems includes emission estimates for activities occurring in petroleum systems from the oil wellhead through crude oil refining, including activities for crude oil exploration, production field operations, crude oil transportation activities, and refining operations. Generally, emissions are estimated for each activity by multiplying emission factors (e.g., emission rate per equipment or per activity) by corresponding activity data (e.g., equipment count or frequency of activity).

EPA received stakeholder feedback on updates in the Inventory through EPA's stakeholder process on oil and gas in the Inventory. Stakeholder feedback is noted below in Uncertainty and Time-Series Consistency, Recalculations Discussion, and Planned Improvements.

*Emission Factors.* References for emission factors include *Methane Emissions from the Natural Gas Industry by the Gas Research Institute and EPA* (EPA/GRI 1996), *Estimates of Methane Emissions from the U.S. Oil Industry* (EPA 1999), *DrillingInfo* (2018), *Compilation of Air Pollutant Emission Factors, AP-42* (EPA 1997), *Global Emissions of Methane from Petroleum Sources* (API 1992), consensus of industry peer review panels, Bureau of Ocean Energy Management (BOEM) reports, and analysis of GHGRP data (EPA 2018).

Emission factors for hydraulically fractured (HF) oil well completions and workovers (in four control categories) were developed using GHGRP data; year-specific data were used to calculate emission factors from 2016-forward and the year 2016 emission factors were applied to all prior years in the time series. The emission factors for all years for pneumatic controllers and chemical injection pumps were developed using GHGRP data for reporting year 2014. The emission factors for tanks, well testing, and associated gas venting and flaring were developed using year-specific GHGRP data for years 2015 forward; earlier years in the time series use emission factors calculated from year 2015 GHGRP data. For miscellaneous production flaring, year-specific emission factors were developed for years 2015 forward from GHGRP data, an emission factor of 0 was assumed for 1990 through 1992, and linear interpolation was applied to develop emission factors for 1993 through 2014. For offshore oil production, two emission factors were calculated using data collected for all federal offshore platforms by BOEM; one for oil platforms in shallow water, and one for oil platforms in deep water. For most sources, emission factors were held constant for the period 1990 through 2016, and trends in emissions reflect changes in activity levels. Emission factors from EPA 1999 are used for all other production and transportation activities.

For associated gas venting and flaring and miscellaneous production flaring, emission factors were developed on a production basis (i.e., emissions per unit oil produced). Additionally, for these two sources, basin-specific activity and emission factors were developed for each basin that in any year from 2011 forward contributed at least 10 percent of total source emissions (on a CO<sub>2</sub> Eq. basis) in the GHGRP. For associated gas venting and flaring, basin-specific factors were developed for four basins: Williston, Permian, Gulf Coast, and Anadarko; for miscellaneous production flaring, basin-specific factors were developed for three basins: Williston, Permian, and Gulf Coast. Data from all other basins were combined, and activity and emission factors developed for the other basins as a single group for each emission source.

For the exploration and production segments, in general, CO<sub>2</sub> emissions for each source were estimated with GHGRP data or by multiplying CO<sub>2</sub> factors by the corresponding CH<sub>4</sub> data, as the CO<sub>2</sub> content of gas relates to the CH<sub>4</sub> content of gas. Sources with CO<sub>2</sub> emissions calculated from GHGRP data were HF completions and workovers, associated gas venting and flaring, tanks, well testing, pneumatic controllers, chemical injection pumps, and miscellaneous production flaring. For these sources, CO<sub>2</sub> was calculated using the same methods as used for CH<sub>4</sub>. Emission factors for offshore oil production (shallow and deep water) were derived using data from BOEM. For other sources, the production field operations emission factors for CO<sub>2</sub> are generally estimated by multiplying the CH<sub>4</sub> emission factors by a conversion factor, which is the ratio of CO<sub>2</sub> content and CH<sub>4</sub> content in produced associated gas.

For the exploration and production segments, N<sub>2</sub>O emissions were estimated for flaring sources using GHGRP data. Sources with N<sub>2</sub>O emissions in the exploration segment were well testing and HF completions with flaring. Sources with N<sub>2</sub>O emissions in the production segment were associated gas flaring, tank flaring, miscellaneous production flaring, and HF workovers with flaring.

For crude oil transportation, emission factors for CH<sub>4</sub> were largely developed using data from EPA (1997), API (1992), and EPA (1999). Emission factors for CO<sub>2</sub> were estimated by multiplying the CH<sub>4</sub> emission factors by a conversion factor, which is the ratio of CO<sub>2</sub> content and CH<sub>4</sub> content in whole crude post-separator.

For petroleum refining activities, year-specific emissions from 2010 forward were directly obtained from EPA's GHGRP. All U.S. refineries have been required to report CH<sub>4</sub>, CO<sub>2</sub>, and N<sub>2</sub>O emissions for all major activities starting with emissions that occurred in 2010. However, GHGRP does have provisions that refineries are not required to report to the GHGRP if their emissions fall below certain thresholds (see Planned Improvements for additional discussion). The reported total of CH<sub>4</sub>, CO<sub>2</sub>, and N<sub>2</sub>O emissions for each activity was used for the emissions in each year from 2010 forward. To estimate emissions for 1990 to 2009, the 2010 to 2013 emissions data from GHGRP along with the refinery feed data for 2010 to 2013 were used to derive emission factors (i.e., sum of activity emissions/sum of refinery feed), which were then applied to the annual refinery feed in years 1990 to 2009.

A complete list of references for emission factors and activity data by emission source is provided in Annex 3.5.

*Activity Data.* References for activity data include DrillingInfo data (DrillingInfo 2018), Energy Information Administration (EIA) reports, *Methane Emissions from the Natural Gas Industry by the Gas Research Institute and EPA* (EPA/GRI 1996), *Estimates of Methane Emissions from the U.S. Oil Industry* (EPA 1999), consensus of industry peer review panels, BOEM reports, the Oil & Gas Journal, the Interstate Oil and Gas Compact Commission, the United States Army Corps of Engineers, and analysis of GHGRP data (EPA 2018).

For many sources, complete activity data were not available for all years of the time series. In such cases, one of three approaches was employed to estimate values, consistent with IPCC good practice. Where appropriate, the activity data were calculated from related statistics using ratios developed based on EPA/GRI 1996 and/or GHGRP data. In some cases, activity data are developed by interpolating between recent data points (such as from GHGRP) and earlier data points, such as from EPA/GRI 1996. Lastly, the previous year's data were used for domestic barges and tankers as current year were not yet available. For offshore production, the number of platforms in shallow water and the number of platforms in deep water are used as activity data and are taken from BOEM datasets; these activity data have not been recently updated and 2010 activity are applied for all recent years.

A complete list of references for emission factors and activity data by emission source is provided in Annex 3.5.

## Uncertainty and Time-Series Consistency

EPA has conducted a quantitative uncertainty analysis using the IPCC Approach 2 methodology (Monte Carlo Simulation technique) to characterize uncertainty for petroleum systems. For more information, please see the memorandum *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2016: Natural Gas and Petroleum Systems Uncertainty Estimates (2018 Uncertainty Memo)*.<sup>78</sup>

EPA used Microsoft Excel's @RISK add-in tool to estimate the 95 percent confidence bound around methane emissions from petroleum systems for the current Inventory, then applied the calculated bounds to both CH<sub>4</sub> and CO<sub>2</sub> emissions estimates. For the analysis, EPA focused on the four highest methane-emitting sources for the year 2017, which together emitted 79 percent of methane from petroleum systems in 2017, and extrapolated the estimated uncertainty for the remaining sources. The @RISK add-in provides for the specification of probability density functions (PDFs) for key variables within a computational structure that mirrors the calculation of the inventory estimate. The IPCC guidance notes that in using this method, "some uncertainties that are not addressed by statistical means may exist, including those arising from omissions or double counting, or other conceptual errors, or from incomplete understanding of the processes that may lead to inaccuracies in estimates developed from models." As a result, the understanding of the uncertainty of emission estimates for this category evolves and improves as the underlying methodologies and datasets improve. The uncertainty bounds reported below only reflect those

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<sup>78</sup> See <<https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems>>

uncertainties that EPA has been able to quantify and do not incorporate considerations such as modeling uncertainty, data representativeness, measurement errors, misreporting or misclassification.

The results presented below provide the 95 percent confidence bound within which actual emissions from this source category are likely to fall for the year 2017, using the recommended IPCC methodology. The results of the Approach 2 uncertainty analysis are summarized in Table 3-43. Petroleum systems CH<sub>4</sub> emissions in 2017 were estimated to be between 25.0 and 51.9 MMT CO<sub>2</sub> Eq., while CO<sub>2</sub> emissions were estimated to be between 15.5 and 32.2 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. Uncertainty bounds for other years of the time series have not been calculated, but uncertainty is expected to vary over the time series. For example, years where many emission sources are calculated with interpolated data would likely have higher uncertainty than years with predominantly year-specific data.

**Table 3-43: Approach 2 Quantitative Uncertainty Estimates for CH<sub>4</sub> and CO<sub>2</sub> Emissions from Petroleum Systems (MMT CO<sub>2</sub> Eq. and Percent)**

Source	Gas	2017 Emission Estimate (MMT CO <sub>2</sub> Eq.) <sup>b</sup>	Uncertainty Range Relative to Emission Estimate <sup>a</sup> (%)			
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Petroleum Systems	CH <sub>4</sub>	37.7	25.0	51.9	-34%	+38%
Petroleum Systems <sup>c</sup>	CO <sub>2</sub>	23.3	15.5	32.2	-34%	+38%

<sup>a</sup> Range of emission estimates estimated by applying the 95 percent confidence intervals obtained from the Monte Carlo Simulation analysis conducted for the year 2017 CH<sub>4</sub> emissions.

<sup>b</sup> All reported values are rounded after calculation. As a result, lower and upper bounds may not be duplicable from other rounded values as shown in table.

<sup>c</sup> An uncertainty analysis for the petroleum systems CO<sub>2</sub> emissions was not performed. The relative uncertainty estimated (expressed as a percent) from the CH<sub>4</sub> uncertainty analysis was applied to the point estimate of petroleum systems CO<sub>2</sub> emissions.

GHGRP data, available starting in 2010 for refineries and in 2011 for other sources, have improved estimates of emissions from petroleum systems. Many of the previously available datasets were collected in the 1990s. To develop a consistent time series for sources with new data, EPA reviewed available information on factors that may have resulted in changes over the time series (e.g., regulations, voluntary actions) and requested stakeholder feedback on trends as well. For most sources, EPA developed annual data for 1993 through 2009 or 2014 by interpolating activity data or emission factors or both between 1992 and 2010 or 2015 data points. Information on time-series consistency for sources updated in this year's Inventory can be found in the Recalculations Discussion below, with additional detail provided in supporting memos (relevant memos are cited in the Recalculations Discussion). For information on other sources, please see the Methodology Discussion above and Annex 3.5.

## QA/QC and Verification Discussion

The petroleum systems emission estimates in the Inventory are continually being reviewed and assessed to determine whether emission factors and activity factors accurately reflect current industry practices. A QA/QC analysis was performed for data gathering and input, documentation, and calculation. QA/QC checks are consistently conducted to minimize human error in the model calculations. EPA performs a thorough review of information associated with new studies, GHGRP data, regulations, public webcasts, and the Natural Gas STAR Program to assess whether the assumptions in the Inventory are consistent with current industry practices. EPA has a multi-step data verification process for GHGRP data, including automatic checks during data-entry, statistical analyses on completed reports, and staff review of the reported data. Based on the results of the verification process, EPA follows up with facilities to resolve mistakes that may have occurred.<sup>79</sup>

As in previous years, EPA conducted early engagement and communication with stakeholders on updates prior to public review. EPA held a stakeholder workshop on greenhouse gas data for oil and gas in October of 2018, and webinars in June of 2018 and February of 2019. EPA released memos detailing updates under consideration and

<sup>79</sup> See <[https://www.epa.gov/sites/production/files/2015-07/documents/ghgrp\\_verification\\_factsheet.pdf](https://www.epa.gov/sites/production/files/2015-07/documents/ghgrp_verification_factsheet.pdf)>

requesting stakeholder feedback. Stakeholder feedback received through these processes is discussed in the Recalculations Discussion and Planned Improvements sections below.

In recent years, several studies have measured emissions at the source level and at the national or regional level and calculated emission estimates that may differ from the Inventory. There are a variety of potential uses of data from new studies, including replacing a previous estimate or factor, verifying or QA of an existing estimate or factor, and identifying areas for updates. In general, there are two major types of studies related to oil and gas greenhouse gas data: studies that focus on measurement or quantification of emissions from specific activities, processes, and equipment, and studies that use tools such as inverse modeling to estimate the level of overall emissions needed to account for measured atmospheric concentrations of greenhouse gases at various scales. The first type of study can lead to direct improvements to or verification of Inventory estimates. In the past few years, EPA has reviewed and in many cases, incorporated data from these data sources. The second type of study can provide general indications on potential over- and under-estimates. A key challenge in using these types of studies to assess Inventory results is having a relevant basis for comparison (i.e., the independent study should assess data from the Inventory and not another data set, such as EDGAR). In an effort to improve the ability to compare the national-level Inventory with measurement results that may be at other scales, a team at Harvard University along with EPA and other coauthors developed a gridded inventory of U.S. anthropogenic methane emissions with 0.1 degree x 0.1 degree spatial resolution, monthly temporal resolution, and detailed scale-dependent error characterization.<sup>80</sup> The gridded methane inventory is designed to be consistent with the U.S. EPA's *Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1990-2014* estimates for the year 2012, which presents national totals.<sup>81</sup>

## Recalculations Discussion

EPA received information and data related to the emission estimates through GHGRP reporting, the annual Inventory formal public notice periods, stakeholder feedback on updates under consideration, and new studies. In June, October, and November 2018, EPA released draft memoranda that discussed changes under consideration and requested stakeholder feedback on those changes. The EPA then created updated versions of the memoranda to document the methodology implemented into the current Inventory.<sup>82</sup> The EPA memorandum *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2017: Other Updates Considered for 2019 and Future GHGIs* (Apr. 2019 *Other Updates* memo) is cited in the Recalculations Discussion below.

EPA thoroughly evaluated relevant information available and made updates to exploration and production segment methodologies for the Inventory, specifically: using GHGRP data to calculate emissions and activity factors for oil well completions and workovers with hydraulic fracturing; using DrillingInfo data (DrillingInfo 2018) to calculate well drilling activity; and revising the basis for calculating the number of active wells represented in GHGRP reporting. In addition, certain sources did not undergo methodological updates, but CH<sub>4</sub> and/or CO<sub>2</sub> emissions changed by greater than 0.05 MMT CO<sub>2</sub> Eq., comparing the previous estimate for 2016 to the current (recalculated) estimate for 2016 (the emissions changes were mostly due to GHGRP data submission revisions); these sources are discussed below and include production tanks, associated gas venting and flaring, miscellaneous production flaring, pneumatic controllers, chemical injection pumps, heaters, and refineries.

Finally, emissions estimates were included for N<sub>2</sub>O from flaring activities in the exploration, production, and refineries segments, and for CO<sub>2</sub> from the crude oil transportation segment.

The combined impact of revisions to 2016 petroleum systems CH<sub>4</sub> emissions, compared to the previous Inventory, is a decrease from 38.6 to 38.2 MMT CO<sub>2</sub> Eq. (0.4 MMT CO<sub>2</sub> Eq., or 1 percent). The recalculations resulted in an average increase in CH<sub>4</sub> emission estimates across the 1990 through 2016 time series, compared to the previous Inventory, of 3.3 MMT CO<sub>2</sub> Eq., or 10 percent, with the largest increases in the estimates for 2005 to 2013 due to the revised data on hydraulically fractured oil well completions.

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<sup>80</sup> See <<https://www.epa.gov/ghgemissions/gridded-2012-methane-emissions>>

<sup>81</sup> See <<https://www.epa.gov/ghgemissions/us-greenhouse-gas-inventory-report-1990-2014>>

<sup>82</sup> Stakeholder materials including draft and final EPA memoranda for the current (i.e., 1990 to 2017) Inventory are available at <<https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems>>.

The combined impact of revisions to 2016 petroleum systems CO<sub>2</sub> emissions, compared to the previous Inventory, is a decrease from 22.8 to 22.2 MMT CO<sub>2</sub> (0.6 MMT CO<sub>2</sub>, or 2 percent). The recalculations resulted in an average increase in emission estimates across the 1990 through 2016 time series, compared to the previous Inventory, of 0.6 MMT CO<sub>2</sub> Eq., or 4 percent.

In Table 3-44 and Table 3-45 below are categories in Petroleum Systems with updated methodologies or with recalculations resulting in a change of greater than 0.05 MMT CO<sub>2</sub> Eq., comparing the previous estimate for 2016 to the current (recalculated) estimate for 2016. For more information, please see the Recalculations Discussion below.

**Table 3-44: Recalculations of CO<sub>2</sub> in Petroleum Systems (MMT CO<sub>2</sub>)**

	<i>Previous Estimate Year 2016, 2018 Inventory</i>	<i>Current Estimate Year 2016, 2019 Inventory</i>	<i>Current Estimate Year 2017, 2019 Inventory</i>
<b>Exploration</b>	+	<b>1.2</b>	<b>1.7</b>
HF Oil Well Completions	+	1.2	1.6
<b>Production</b>	<b>19.0</b>	<b>17.0</b>	<b>18.0</b>
Tanks	7.4	5.9	4.4
Associated Gas Venting & Flaring	9.1	8.6	10.5
Miscellaneous Flaring	2.5	2.2	2.6
HF Oil Well Workovers	+	0.2	0.3
<b>Transportation</b>	<i>NE</i>	+	+
<b>Refining</b>	<b>3.7</b>	<b>4.0</b>	<b>3.7</b>
<b>Petroleum Systems Total</b>	<b>22.8</b>	<b>22.2</b>	<b>23.3</b>

NE (Not Estimated)

+ Does not exceed 0.05 MMT CO<sub>2</sub>.

**Table 3-45: Recalculations of CH<sub>4</sub> in Petroleum Systems (MMT CO<sub>2</sub> Eq.)**

	<i>Previous Estimate Year 2016, 2018 Inventory</i>	<i>Current Estimate Year 2016, 2019 Inventory</i>	<i>Current Estimate Year 2017, 2019 Inventory</i>
<b>Exploration</b>	<b>2.1</b>	<b>0.5</b>	<b>0.4</b>
HF Oil Well Completions	2.0	0.4	0.3
<b>Production</b>	<b>35.4</b>	<b>36.8</b>	<b>36.4</b>
Pneumatic Controllers	18.5	20.5	20.9
Tanks	3.2	2.6	1.5
Heaters	0.8	0.7	0.7
Chemical Injection Pumps	2.0	2.1	2.0
HF Oil Well Workovers	+	0.1	0.1
<b>Transportation</b>	<b>0.2</b>	<b>0.2</b>	<b>0.2</b>
<b>Refining</b>	<b>0.9</b>	<b>0.7</b>	<b>0.7</b>
<b>Petroleum Systems Total</b>	<b>38.6</b>	<b>38.2</b>	<b>37.7</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub>.

## Exploration

### *HF Oil Well Completions (Methodological Update)*

EPA revised the HF oil well completions methodology by establishing four control categories (non-REC with venting, non-REC with flaring, REC with venting, and REC with flaring) and developing new activity and emission factors for these categories. The new methodology is detailed in the Apr. 2019 *Other Updates* memo. The previous factors (for controlled and uncontrolled event categories) relied on data analysis from the 2015 NSPS OOOOa rulemaking proposal. As described above in the Methodology discussion, EPA has newly calculated year-specific activity factors (fraction of events in each category) and emission factors for years 2016 forward using GHGRP data. To estimate emissions over the time series, EPA applied the year 2016 emission factors for all prior years and developed activity factors by following the existing methodology for HF gas well events combined with oil well-

specific assumptions regarding when controls became prevalent. For HF oil well event activity factors, the following assumptions are applied: (1) for years 1990 to 2007, all completions and workovers are non-REC, and 10 percent of events flare; (2) for the first year in which GHGRP data are available, 2016, control fractions across the four categories are developed directly from reported GHGRP data; and (3) for intermediate years, 2008 to 2015, control fractions are developed through linear interpolation. This approach produces activity factors across the time series that are generally consistent with the previous assumption that oil well RECs are introduced beginning in year 2008, during which 7 percent of completions and workovers are REC, and 10 percent of both REC and non-REC events flare. EPA did not change the methodology of calculating total activity for this source, which relies on analyzing DrillingInfo data (DrillingInfo 2018) to obtain the total HF oil well completion event count in each year of the time series. Stakeholder feedback supported an approach of using GHGRP data to update activity and emissions factors on an annual basis from 2016 forward.

**Table 3-46: HF Oil Well Completions National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
HF Completions: Non-REC with Venting	110,326	171,881	227,632	171,542	60,488	7,043	2,168
HF Completions: Non-REC with Flaring	360	560	2,502	2,788	1,804	1,018	1,791
HF Completions: REC with Venting	0	0	4,800	6,081	4,383	2,714	2,223
HF Completions: REC with Flaring	0	0	7,707	9,764	7,037	4,358	6,424
<b>Total Emissions</b>	<b>110,685</b>	<b>172,441</b>	<b>242,642</b>	<b>190,175</b>	<b>73,712</b>	<b>15,132</b>	<b>12,606</b>
<i>Previous Estimate</i>	20,796	31,070	109,422	120,925	78,525	78,525	NA

NA (Not Applicable)

**Table 3-47: HF Oil Well Completions National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
HF Completions: Non-REC with Venting	2.5	4.0	5.3	4.0	1.4	0.2	0.2
HF Completions: Non-REC with Flaring	79	123	547	610	395	223	410
HF Completions: REC with Venting	0.0	0.0	0.3	0.3	0.3	0.2	0.1
HF Completions: REC with Flaring	0.0	0.0	1,661	2,104	1,517	939	1,209
<b>Total Emissions</b>	<b>81.2</b>	<b>126.5</b>	<b>2,214</b>	<b>2,719</b>	<b>1,913</b>	<b>1,162</b>	<b>1,619</b>
<i>Previous Estimate</i>	1.2	1.7	6.1	6.7	4.4	4.4	NA

NA (Not Applicable)

### *Well Drilling (Methodological Update)*

EPA updated the methodology for estimating the number of oil wells drilled across the time series to use DrillingInfo data (DrillingInfo 2018). The new methodology is detailed in the Apr. 2019 *Other Updates* memo. In previous Inventories, the U.S. Department of Energy's Energy Information Administration (DOE/EIA) *Monthly Energy Review* well drilling activity data set was used to develop well drilling activity inputs, but this publication does not provide data after year 2010. EPA therefore developed a methodology of analyzing DrillingInfo data to estimate counts of oil wells drilled in each time series year, 1990 through 2017. These activity data for select years are shown in Table 3-48 below.

**Table 3-48: Count of Oil Wells Drilled**

Source	1990	2005	2013	2014	2015	2016	2017
Oil Wells Drilled	19,919	18,216	35,671	36,910	17,359	10,242	10,242
<i>Previous Estimate</i>	17,234	12,053	17,774 <sup>a</sup>	17,774 <sup>a</sup>	17,774 <sup>a</sup>	17,774 <sup>a</sup>	NA

<sup>a</sup> – Year-specific data not available; the year 2010 estimate was assigned as a surrogate value.



NA (Not Applicable)

## Production

### *HF Oil Well Workovers (Methodological Update)*

EPA revised the HF oil well workovers methodology to use the same general approach as described above for HF oil well completions. EPA revised the oil well workovers methodology by separating HF and non-HF events, then establishing four control categories for HF events (non-REC with venting, non-REC with flaring, REC with venting, and REC with flaring) and developing new activity and emission factors for these categories. The new methodology is detailed in the Apr. 2019 *Other Updates* memo. The previous methodology did not use separate emissions or activity assumptions for HF versus non-HF workover events. As described above in the Methodology discussion, EPA has newly calculated year-specific activity factors (fraction of events in each category) and emission factors for years 2016 forward using GHGRP data. To estimate emissions over the time series, EPA applied the year 2016 emission factors for all prior years and developed activity factors by following the existing methodology for HF gas well events combined with oil well-specific assumptions regarding when controls became prevalent. For HF oil well event activity factors, the following assumptions are applied: (1) for years 1990 to 2007, all completions and workovers are non-REC, and 10 percent of events flare; (2) for the first year in which GHGRP data are available, 2016, control fractions across the four categories are developed directly from reported GHGRP data; and (3) for intermediate years, 2008–2015, control fractions are developed through linear interpolation. This approach produces activity factors across the time series that are generally consistent with the previous assumption that oil well RECs are introduced beginning in year 2008, during which 7 percent of completions and workovers are REC, and 10 percent of both REC and non-REC events flare. EPA also updated the methodology of calculating total activity for this source; EPA applies the existing assumption used for HF gas wells, that 1 percent of HF wells are worked over in a given year. Stakeholder feedback supported an approach of using GHGRP data to update activity and emissions factors on an annual basis from 2016 forward.

**Table 3-49: HF Oil Well Workovers National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
HF Workovers: Non-REC with Venting	31,119	35,018	22,290	17,601	10,808	3,318	0
HF Workovers: Non-REC with Flaring	101	114	142	148	142	130	114
HF Workovers: REC with Venting	0	0	745	966	1,146	1,275	678
HF Workovers: REC with Flaring	0	0	485	629	746	830	1,229
<b>Total Emissions</b>	<b>31,220</b>	<b>35,132</b>	<b>23,662</b>	<b>19,344</b>	<b>12,842</b>	<b>5,552</b>	<b>2,022</b>
<i>Previous Estimate<sup>a</sup></i>	77	65	79	82	82	78	NA

NA (Not Applicable)

<sup>a</sup> Estimate includes emissions for HF and non-HF workovers.

**Table 3-50: HF Oil Well Workovers National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
HF Workovers: Non-REC with Venting	0.7	0.8	0.5	0.4	0.2	0.1	0.0
HF Workovers: Non-REC with Flaring	22.2	25.0	31.1	32.3	31.1	28.4	26.2
HF Workovers: REC with Venting	0.0	0.0	0.0	0.1	0.1	0.1	0.0
HF Workovers: REC with Flaring	0.0	0.0	104.5	135.5	160.7	178.8	231.3
<b>Total Emissions</b>	<b>22.9</b>	<b>25.8</b>	<b>136.1</b>	<b>168.3</b>	<b>192.1</b>	<b>207.4</b>	<b>257.5</b>
<i>Previous Estimate<sup>a</sup></i>	0.0	0.0	0.0	0.0	0.0	0.0	NA

NA (Not Applicable)

<sup>a</sup> Estimate includes emissions for HF and non-HF workovers.

### *Tanks (Recalculation with Updated Data)*

Production tank CH<sub>4</sub> and CO<sub>2</sub> emission estimates decreased in the current Inventory, compared to the previous Inventory. This change was due to GHGRP submission revisions and updated production data (see the Oil Production discussion below). For CO<sub>2</sub> emissions, in general, a smaller fraction of the GHGRP tank throughput went through tanks with flares and certain GHGRP-based emission factors were lower. For CH<sub>4</sub>, while a larger fraction of the GHGRP tank throughput went through tanks without controls, the calculated GHGRP-based emission factors were lower.

**Table 3-51: Production Storage Tank National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Large Tanks w/ Flares	0	2,510	5,649	6,704	7,230	5,105	5,687
Large Tanks w/ VRU	0	1,133	2,550	3,026	3,263	19,180	8,963
Large Tanks w/o Control	209,643	52,011	38,001	45,093	48,631	66,448	40,056
Small Tanks w/ Flares	0	15	34	41	44	22	44
Small Tanks w/o Flares	4,246	2,041	2,992	3,551	3,830	3,358	2,248
Malfunctioning Dump Valves	3,998	2,345	3,770	4,473	4,824	8,079	4,339
<b>Total Emissions</b>	<b>217,887</b>	<b>60,055</b>	<b>52,997</b>	<b>62,887</b>	<b>67,821</b>	<b>102,191</b>	<b>61,336</b>
<i>Previous Estimate</i>	257,923	84,409	65,467	76,752	82,496	127,025	NA

NA (Not Applicable)

**Table 3-52: Production Storage Tank National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Large Tanks w/ Flares	0	2,619	5,896	6,997	7,546	5,843	4,380
Large Tanks w/ VRU	0	5	11	13	14	4.6	4
Large Tanks w/o Control	23	6	4	4.9	5	7	5
Small Tanks w/ Flares	0	2	5	6	7	17	15
Small Tanks w/o Flares	6	3	4	5	5	5	3
Malfunctioning Dump Valves	17	10	16	19	20	18	15
<b>Total Emissions</b>	<b>46</b>	<b>2,645</b>	<b>5,937</b>	<b>7,045</b>	<b>7,598</b>	<b>5,894</b>	<b>4,422</b>
<i>Previous Estimate</i>	53	3,444	6,922	8,115	8,722	7,351	NA

NA (Not Applicable)

### *Pneumatic Controllers (Recalculation with Updated Data)*

Pneumatic controller CH<sub>4</sub> emission estimates increased in the current Inventory, compared to the previous Inventory, due to GHGRP submission revisions and the use of GHGRP well counts from the facility overview table (see the Well Counts discussion below). The well count change shifted certain controllers from being assigned to natural gas systems to petroleum systems. Pneumatic controller CH<sub>4</sub> emission estimates increased by an average of 5 percent across the 1990 to 2016 time series.

**Table 3-53: Pneumatic Controller National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Pneumatic Controllers: High Bleed	724,225	418,481	100,587	87,778	77,849	82,071	52,265
Pneumatic Controllers: Low Bleed	49,429	43,906	29,291	28,589	25,341	17,415	19,162
Pneumatic Controllers: Int Bleed	0	238,603	613,112	660,145	682,514	718,683	765,378
<b>Total Emissions</b>	<b>773,655</b>	<b>700,990</b>	<b>742,990</b>	<b>776,512</b>	<b>785,704</b>	<b>818,169</b>	<b>836,804</b>
<i>Previous Estimate</i>	765,975	663,461	687,210	715,768	720,996	739,125	NA

NA (Not Applicable)

### *Associated Gas Venting and Flaring (Recalculation with Updated Data)*

Associated gas venting and flaring CO<sub>2</sub> emission estimates decreased for 2016 and increased for 1990 through 2015 in the current Inventory, compared to the previous Inventory. Compared to the previous inventory, on average, calculated CO<sub>2</sub> emissions increased across the 1990 to 2015 time series by 20 percent, and decreased by 6 percent for 2016. This change was due to GHGRP submission revisions and updated production data (see the Oil Production discussion below). The emission calculations are performed at a basin-level, and the changes impacted each basin uniquely. However, the changes in CO<sub>2</sub> emissions were mainly driven by the Permian Basin data.

**Table 3-54: Associated Gas Venting and Flaring National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Associated Gas Venting	21	11	11	12	13	6	19
Associated Gas Flaring	5,172	3,925	10,384	12,711	13,955	8,587	10,506
<b>Total Emissions</b>	<b>5,193</b>	<b>3,937</b>	<b>10,395</b>	<b>12,723</b>	<b>13,968</b>	<b>8,593</b>	<b>10,525</b>
<i>Previous Estimate</i>	4,028	3,314	9,193	11,248	12,234	9,108	NA

NA (Not Applicable)

### *Miscellaneous Production Flaring (Recalculation with Updated Data)*

Miscellaneous production flaring CO<sub>2</sub> emission estimates decreased for most years of the current Inventory, except for an increase for 2015, compared to the previous Inventory. There were several underlying factors that impacted the changes each year; GHGRP submission revisions, use of GHGRP well counts from the facility overview table (see the Well Counts discussion below), a correction to the linear interpolation calculation for emission factors in years 1993 through 2014, and updated production data (see the Oil Production discussion below). In addition, the emission calculations are performed at a basin-level, and the changes impacted each basin uniquely.

**Table 3-55: Miscellaneous Production Flaring National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Misc. Flaring	0	800	2,487	3,157	3,571	2,201	2,631
<i>Previous Estimate</i>	0	929	2,541	3,181	3,418	2,455	NA

NA (Not Applicable)

### *Chemical Injection Pumps (Recalculation with Updated Data)*

Chemical injection pump CH<sub>4</sub> emission estimates increased by an average of 1.4 percent over the time series and for certain recent years increased by approximately 3 percent for the current Inventory, compared to the previous Inventory. The emission estimates increases are due to updated well counts (see the Well Counts discussion below); emission factors and activity factors were not updated.

**Table 3-56: Chemical Injection Pump National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Chemical Injection Pump	49,465	67,785	83,972	87,212	86,114	83,215	81,660
<i>Previous Estimate</i>	49,131	66,585	82,084	84,934	85,016	80,974	NA

NA (Not Applicable)

### *Heaters (Recalculation with Updated Data)*

Combustion CH<sub>4</sub> emission estimates from heaters decreased by an average of approximately 22 percent for each year of the time series in the current Inventory, compared to the previous Inventory. The decrease is due to a decrease in total oil production in each year, the applicable activity data for heaters, which was updated for the current Inventory (see the Oil Production discussion below).

**Table 3-57: Heater National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Heater	23,935	14,038	22,570	26,782	28,883	26,504	28,051
<i>Previous Estimate</i>	26,944	18,991	27,350	32,065	34,465	32,446	NA

NA (Not Applicable)

### *Well Counts (Recalculation with Updated Data)*

For total national well counts, EPA has used a more recent version of the DrillingInfo data set (DrillingInfo 2018) to update well counts data in the Inventory. EPA also updated the DrillingInfo data processing methodology to more accurately count wells in states with lease-level reporting (e.g., Kansas), which resulted in slight increased counts across the time series. While this was not a significant recalculation (increases are 2 to 3 percent across the time series), this is a key input to the Inventory, so results are highlighted here.

**Table 3-58: Producing Oil Well Count Data**

Oil Well Count	1990	2005	2013	2014	2015	2016	2017
Number of Oil Wells	564,090	480,482	582,769	605,259	597,635	577,515	566,726
<i>Previous Estimate</i>	553,899	469,632	569,670	589,450	590,017	561,964	NA

NA (Not Applicable)

In October 2018, EIA released an updated time series of national oil and gas well counts (covering 2000 through 2017). EIA estimates 991,000 total producing wells for year 2017. EPA's total well count for this year is 978,176. EPA's well counts in recent time series years are generally 2 percent lower than EIA's. EIA's well counts include side tracks, completions, and recompletions, and therefore are expected to be higher than EPA's which include only producing wells. EPA and EIA use a different threshold for distinguishing between oil versus gas (EIA uses 6 mcf/bbl, while EPA uses 100 mcf/bbl), which results in EIA having a lower fraction of oil wells and a higher fraction of gas wells than EPA.

For the count of wells included in GHGRP reporting (used to develop wellhead-based emissions and activity factors), EPA previously referenced the wellhead counts contained within the reporting table for onshore production equipment leak emissions. Due to updated reporting requirements for year 2017 forward, well counts provided as part of the facility overview information (i.e., wells producing at the end of the calendar year plus wells removed from production in a given year) provide more complete estimates. Therefore, EPA used well counts from the facility overview table for source-specific methodologies that rely on GHGRP reported well counts in the current Inventory. Comparing the GHGRP well counts from the facility overview table to the equipment leaks table: a larger population of the wells were reported as "oil" production type in the facility overview information table, compared to the equipment leaks table, which generally led to increased activity and emissions for petroleum systems. For example, as discussed in the sections above, production segment emissions from pneumatic controllers and miscellaneous production flaring increased.

### *Oil Production*

EPA reviewed the national oil production data that were previously used in the Inventory and determined a more appropriate dataset were available. In previous Inventories, production from the EIA's *Monthly Energy Review* were used; specifically, Table 3.1 Petroleum Overview, "Total Crude Oil Field Production". However, this dataset includes both onshore and offshore production and did not distinguish between the two. EIA provides more detailed production data in an online database, including specifically reporting federal offshore production.<sup>83</sup> The EIA online database production data were used for the current Inventory and federal offshore production data were excluded. This meant the production values decreased across the time series, but are more specific to onshore production. The emission sources that rely on oil production as an activity driver and that were impacted the most by this change are production tanks, associated gas venting and flaring, miscellaneous production flaring, and heaters (all of which are

<sup>83</sup> Available at <[https://www.eia.gov/dnav/pet/pet\\_crd\\_crpdn\\_adc\\_mbbbl\\_a.htm](https://www.eia.gov/dnav/pet/pet_crd_crpdn_adc_mbbbl_a.htm)>

discussed above). In addition, oil production data are activity drivers for estimating fugitive emissions from production compressors and the sales area (loadings), and emissions due to pressure relief valve releases.

**Table 3-59: Oil Production Data (Million Barrels)**

Source	1990	2005	2013	2014	2015	2016	2017
Oil Production	2,385	1,399	2,249	2,668	2,878	2,641	2,795
<i>Previous Estimate</i>	2,685	1,892	2,725	3,195	3,434	3,233	NA

NA (Not Applicable)

### Floating Roof Tanks

EPA removed the line item estimate for production segment floating roof tanks that was included in previous Inventories. The number of floating roof tanks and their emissions were minimal in the context of the petroleum production segment, and available data are limited; data on the number of floating roof tanks are only available for 1995, and the 1995 count is applied to all other years. EPA sought stakeholder input on whether and how to include floating roof tank emission estimates in the production segment and did not receive objections to the removal of this source. The emission estimate for this source in the previous Inventory was 159 metric tons CH<sub>4</sub> in each year, or 0.01 percent of CH<sub>4</sub> emissions in year 2016.

### Crude Oil Transportation

EPA newly calculated CO<sub>2</sub> emissions from crude oil transportation in the current Inventory. Prior Inventories did not calculate CO<sub>2</sub> emissions from crude oil transportation. CO<sub>2</sub> emission factors were calculated by multiplying the CH<sub>4</sub> emission factors for each source by a conversion factor, which is the ratio of CO<sub>2</sub> content and CH<sub>4</sub> content in whole crude post-separator. Total CO<sub>2</sub> emissions from crude oil transportation are included in Table 3-60 below, and emissions for each source can be found in Annex 3.5.

**Table 3-60: Crude Oil Transportation National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Crude Oil Transportation	0.9	0.7	1.0	1.2	1.2	1.1	1.1

Recalculations due to updated activity data for the quantity of petroleum transported by barge or tanker in the crude oil transportation segment did not result in a change in CH<sub>4</sub> emission estimates for 1990 to 2015. Updated activity data for 2016 resulted in a decrease in calculated CH<sub>4</sub> emissions of approximately 3 percent.

### Refining

There are minimal changes in calculated CH<sub>4</sub> and CO<sub>2</sub> emissions for 1990 to 2015 for this segment (e.g., average change is less than 0.1 percent each year). However, recalculations for 2016 resulted in CO<sub>2</sub> emission estimates increasing by 8 percent and CH<sub>4</sub> emissions decreasing by 24 percent. The 2016 emissions changes are due to GHGRP submission revisions.

### N<sub>2</sub>O Emissions

EPA newly calculated N<sub>2</sub>O emissions in the current Inventory, as discussed in the Apr. 2019 *Other Updates* memo. Prior Inventories did not calculate N<sub>2</sub>O emissions from petroleum systems. For each flaring emission source calculation methodology which uses GHGRP data, the existing source-specific methodology was applied to calculate N<sub>2</sub>O emission factors. This update was applied for flaring sources in the exploration, production, and refining segments.

**Table 3-61: N<sub>2</sub>O National Emissions (Metric Tons N<sub>2</sub>O)**

Source	1990	2005	2013	2014	2015	2016	2017
Exploration	0.6	0.6	4.3	5.2	3.8	2.1	2.5
HF Completions with Flaring	0.1	0.2	3.8	4.7	3.3	2.0	2.5

Non-Completion Well Testing with Flaring	0.4	0.4	0.4	0.5	0.5	0.1	0.1
<b>Production</b>	<b>14.8</b>	<b>17.9</b>	<b>43.6</b>	<b>53.1</b>	<b>58.5</b>	<b>42.8</b>	<b>42.4</b>
Associated Gas Flaring	14.8	11.0	26.3	32.1	35.5	25.9	28.2
Storage Tanks with Flaring	NO	5.7	12.7	15.1	16.3	12.6	9.0
Misc. Production Flaring	NO	1.2	4.3	5.6	6.3	4.0	4.9
HF Workovers with Flaring	+	+	0.2	0.3	0.3	0.4	0.4
<b>Crude Oil Transportation</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>	<b>NE</b>
<b>Refining</b>	<b>30.7</b>	<b>34.8</b>	<b>34.2</b>	<b>32.4</b>	<b>39.1</b>	<b>38.8</b>	<b>36.4</b>
Refinery Flares	30.7	34.8	34.2	32.4	39.1	38.8	36.4
<b>Total</b>	<b>46.1</b>	<b>53.3</b>	<b>82.0</b>	<b>90.7</b>	<b>101.4</b>	<b>83.7</b>	<b>81.3</b>

NE (Not Estimated)

NO (Not Occurring)

+ less than 0.05

## Planned Improvements

### Offshore Platforms

EPA is considering updates to the offshore platform emissions calculation methodology, as discussed in the memorandum *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2016: Additional Revisions Under Consideration*.<sup>84</sup> The current emission factors were based on data from the 2011 DOI/Bureau of Ocean Energy Management's (BOEM) dataset, and 2014 BOEM data are available. A different source for platform counts is also being considered.

### Well-Related Activity Data

EPA will continue to assess available data, including data from the GHGRP and stakeholder feedback on considerations, to improve activity estimates for sources that rely on well-related activity data. For example, EPA will review GHGRP data regarding reported well workover rates and seek information on other data sets that might inform estimates of non-hydraulically fractured oil well completions and workovers.

### Upcoming Data, and Additional Data that Could Inform the Inventory

EPA will assess new data received by the Methane Challenge Program on an ongoing basis, which may be used to confirm or improve existing estimates and assumptions.

EPA continues to track studies that contain data that may be used to update the Inventory, such as an upcoming field study by American Petroleum Institute (API) on pneumatic controllers and separate studies by research groups that will examine offshore platform emissions. EPA will also continue to assess studies that include and compare both top-down and bottom-up estimates, and which could lead to improved understanding of unassigned high emitters (e.g., identification of emission sources and information on frequency of high emitters) as recommended in stakeholder comments.

EPA also continues to seek new data that could be used to assess or update the estimates in the Inventory. For example, in recent years, stakeholder comments have highlighted areas where additional data that could inform the Inventory are currently limited or unavailable:

- Tank malfunction and control efficiency data.
- Activity data and emissions data for production facilities that do not report to GHGRP.
- Associated gas venting and flaring data on practices from 1990 through 2010.
- Refineries emissions data.

<sup>84</sup> See <<https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems>>

- Anomalous leak events.

EPA will continue to seek available data on these and other sources as part of the process to update the Inventory.

### Box 3-7: Carbon Dioxide Transport, Injection, and Geological Storage

Carbon dioxide is produced, captured, transported, and used for Enhanced Oil Recovery (EOR) as well as commercial and non-EOR industrial applications. This CO<sub>2</sub> is produced from both naturally-occurring CO<sub>2</sub> reservoirs and from industrial sources such as natural gas processing plants and ammonia plants. In the Inventory, emissions from naturally-produced CO<sub>2</sub> are estimated based on the specific application.

In the Inventory, CO<sub>2</sub> that is used in non-EOR industrial and commercial applications (e.g., food processing, chemical production) is assumed to be emitted to the atmosphere during its industrial use. These emissions are discussed in the Carbon Dioxide Consumption section. The naturally-occurring CO<sub>2</sub> used in EOR operations is assumed to be fully sequestered. Additionally, all anthropogenic CO<sub>2</sub> emitted from natural gas processing and ammonia plants is assumed to be emitted to the atmosphere, regardless of whether the CO<sub>2</sub> is captured or not. These emissions are currently included in the Natural Gas Systems and the Ammonia Production sections of the Inventory report, respectively.

IPCC includes methodological guidance to estimate emissions from the capture, transport, injection, and geological storage of CO<sub>2</sub>. The methodology is based on the principle that the carbon capture and storage system should be handled in a complete and consistent manner across the entire Energy sector. The approach accounts for CO<sub>2</sub> captured at natural and industrial sites as well as emissions from capture, transport, and use. For storage specifically, a Tier 3 methodology is outlined for estimating and reporting emissions based on site-specific evaluations. However, IPCC (IPCC 2006) notes that if a national regulatory process exists, emissions information available through that process may support development of CO<sub>2</sub> emission estimates for geologic storage.

In the United States, facilities that produce CO<sub>2</sub> for various end-use applications (including capture facilities such as acid gas removal plants and ammonia plants), importers of CO<sub>2</sub>, exporters of CO<sub>2</sub>, facilities that conduct geologic sequestration of CO<sub>2</sub>, and facilities that inject CO<sub>2</sub> underground, are required to report greenhouse gas data annually to EPA through its GHGRP. Facilities conducting geologic sequestration of CO<sub>2</sub> are required to develop and implement an EPA-approved site-specific monitoring, reporting and verification plan, and to report the amount of CO<sub>2</sub> sequestered using a mass balance approach.

GHGRP data relevant for this inventory estimate consists of national-level annual quantities of CO<sub>2</sub> captured and extracted for EOR applications for 2010 to 2017. However, for 2015 through 2017, data from EPA's GHGRP (Subpart PP) were held constant from 2014 levels, due to data confidentiality reasons. EPA will continue to evaluate the availability of additional GHGRP data and other opportunities for improving the emission estimates. Several facilities are reporting under subpart RR (Geologic Sequestration of Carbon Dioxide). In 2016, one facility reported 3.1 MMT of CO<sub>2</sub> sequestered in subsurface geological formations and 9,818 metric tons of CO<sub>2</sub> emitted from equipment leaks. In 2017, three facilities reported 6.0 MMT of CO<sub>2</sub> sequestered in subsurface geological formations, and 9,577 metric tons of CO<sub>2</sub> emitted from equipment leaks.

These estimates indicate that the amount of CO<sub>2</sub> captured and extracted from natural and industrial sites for EOR applications in 2017 is 59.3 MMT CO<sub>2</sub> Eq. (59,318 kt) (see Table 3-62 and Table 3-63). The quantity of CO<sub>2</sub> captured and extracted is noted here for information purposes only; CO<sub>2</sub> captured and extracted from industrial and commercial processes is assumed to be emitted and included in emissions totals from those processes.

**Table 3-62: Quantity of CO<sub>2</sub> Captured and Extracted for EOR Operations (MMT CO<sub>2</sub>)**

Stage	1990	2005	2013	2014	2015	2016	2017
Capture Facilities	4.8	6.5	12.2	13.1	13.1	13.1	13.1
Extraction Facilities	20.8	28.3	47.0	46.2	46.2	46.2	46.2
<b>Total</b>	<b>25.6</b>	<b>34.7</b>	<b>59.2</b>	<b>59.3</b>	<b>59.3</b>	<b>59.3</b>	<b>59.3</b>

Note: Totals may not sum due to independent rounding.

**Table 3-63: Quantity of CO<sub>2</sub> Captured and Extracted for EOR Operations (kt)**

Stage	1990	2005	2013	2014	2015	2016	2017
Capture Facilities	4,832	6,475	12,205	13,093	13,093	13,093	13,093
Extraction Facilities	20,811	28,267	46,984	46,225	46,225	46,225	46,225
<b>Total</b>	<b>25,643</b>	<b>34,742</b>	<b>59,189</b>	<b>59,318</b>	<b>59,318</b>	<b>59,318</b>	<b>59,318</b>

Note: Totals may not sum due to independent rounding.

## 3.7 Natural Gas Systems (CRF Source Category 1B2b)

The U.S. natural gas system encompasses hundreds of thousands of wells, hundreds of processing facilities, and over a million miles of transmission and distribution pipelines. Overall, natural gas systems emitted 165.6 MMT CO<sub>2</sub> Eq. (6,624 kt) of CH<sub>4</sub> in 2017, a 14 percent decrease compared to 1990 emissions, and less than 1 percent decrease compared to 2016 emissions (see Table 3-64, Table 3-65, and Table 3-66), 26.3 MMT CO<sub>2</sub> Eq. (26,327 kt) of non-combustion CO<sub>2</sub> in 2017, a 12 percent decrease compared to 1990 emissions, and a 3 percent increase compared to 2016 levels, and 0.005 MMT CO<sub>2</sub> Eq. (0.02 kt) of N<sub>2</sub>O, a 438 percent increase compared to 1990 emissions.

The 1990 to 2017 trend in CH<sub>4</sub> is not consistent across segments. Overall, the 1990 to 2017 decrease in CH<sub>4</sub> emissions is due primarily to the decrease in emissions from the distribution (73 percent decrease), transmission and storage (43 percent decrease), processing (45 percent decrease), and exploration (69 percent decrease) segments. Over the same time period, the production segment saw increased methane emissions of 62 percent (with onshore production emissions increasing 29 percent, offshore production emissions increasing 7 percent, and gathering and boosting (G&B) emissions increasing 109 percent). The 1990 to 2017 decrease in CO<sub>2</sub> is due primarily to decreases in acid gas removal emissions in the processing segment, where acid gas removal emissions per plant have decreased over time.

Methane and non-combustion CO<sub>2</sub> emissions from natural gas systems include those resulting from normal operations, routine maintenance, and system upsets. Emissions from normal operations include: natural gas engine and turbine uncombusted exhaust, flaring, and leak emissions from system components. Routine maintenance emissions originate from pipelines, equipment, and wells during repair and maintenance activities. Pressure surge relief systems and accidents can lead to system upset emissions. Below is a characterization of the five major stages of the natural gas system. Each of the stages is described and the different factors affecting CH<sub>4</sub> and non-combustion CO<sub>2</sub> emissions are discussed.

Emissions of N<sub>2</sub>O from flaring activities are included in the Inventory, with most of the emissions occurring in the processing and production segments.

Each year, some estimates in the Inventory are recalculated with improved methods and/or data. These improvements are implemented consistently across the previous Inventory's time series (i.e., 1990 to 2016) to ensure that the trend is accurate. Recalculations in natural gas systems in this year's Inventory include:

- Updated methodology for G&B pipeline emissions.
- Updated methodology for transmission pipeline blowdown emissions.
- Updated methodology for LNG estimates (emissions for both storage stations and import/export terminals) within the transmission and storage segment.
- Added N<sub>2</sub>O emissions that were not previously reported in the Inventory.
- Updated the data source for well drilling activity.
- Recalculations due to GHGRP submission revisions.

The Recalculations Discussion section below provides more details on the updated methods.



*Exploration.* Exploration includes well drilling, testing, and completions. Emissions from exploration account for 1 percent of CH<sub>4</sub> emissions and 2 percent of CO<sub>2</sub> emissions from natural gas systems in 2017. Well completions account for most of the CH<sub>4</sub> emissions in 2017, with well testing and drilling also contributing emissions. Flaring emissions account for most of the non-combustion CO<sub>2</sub> emissions. Methane emissions from exploration decreased by 69 percent from 1990 to 2017, with the largest decreases coming from hydraulically fractured gas well completions without reduced emissions completions (RECs) or flaring. Methane emissions increased 75 percent from 2016 to 2017 due to increases in emissions from completions, mostly from hydraulically fractured well completions with RECs without flaring. Methane emissions were highest from 2006 to 2008. Carbon dioxide emissions from exploration increased by 18 percent from 1990 to 2017, and by 149 percent from 2016 to 2017 due to increases in flaring. Carbon dioxide emissions were highest from 2006 to 2008. Nitrous oxide emissions decreased 37 percent from 1990 to 2017, and increased 156 percent from 2016 to 2017.

*Production (including gathering and boosting).* In the production stage, wells are used to withdraw raw gas from underground formations. Emissions arise from the wells themselves, and well-site gas treatment equipment such as dehydrators and separators. Gathering and boosting emission sources are included within the production sector. The gathering and boosting sources include gathering and boosting stations (with multiple emission sources on site) and gathering pipelines. The gathering and boosting stations receive natural gas from production sites and transfer it, via gathering pipelines, to transmission pipelines or processing facilities (custody transfer points are typically used to segregate sources between each segment). Emissions from production (including gathering and boosting) account for 65 percent of CH<sub>4</sub> emissions and 11 percent of non-combustion CO<sub>2</sub> emissions from natural gas systems in 2017. Emissions from compressors, pneumatic controllers, and offshore platforms account for most of the CH<sub>4</sub> emissions in 2017. Flaring emissions account for most of the non-combustion CO<sub>2</sub> emissions with the highest emissions coming from miscellaneous production flaring, flaring to control tank emissions, and offshore flaring. National total dry gas production in the U.S. increased by 53 percent from 1990 to 2017, and by 3 percent from 2016 to 2017. Methane emissions from production increased by 62 percent from 1990 to 2017, due primarily to increases in emissions from pneumatic controllers (due to an increase in the number of controllers, particularly in the number of intermittent bleed controllers) and increases in emissions from gathering and boosting stations. Methane emissions increased 1 percent from 2016 to 2017 due to increases in emissions from gathering and boosting stations and hydraulically fractured well workovers with RECs and venting. Carbon dioxide emissions from production increased by 175 percent from 1990 to 2017 due to increases in flaring, and decreased 11 percent from 2016 to 2017 due primarily to a decrease in emissions from large tanks with flares. Nitrous oxide emissions increased 480 percent from 1990 to 2017 and decreased 8 percent from 2016 to 2017.

*Processing.* In this stage, natural gas liquids and various other constituents from the raw gas are removed, resulting in “pipeline quality” gas, which is injected into the transmission system. Fugitive CH<sub>4</sub> emissions from compressors, including compressor seals, are the primary emission source from this stage. Most of the non-combustion CO<sub>2</sub> emissions come from acid gas removal (AGR) units, which are designed to remove CO<sub>2</sub> from natural gas. Processing plants account for 7 percent of CH<sub>4</sub> emissions and 85 percent of non-combustion CO<sub>2</sub> emissions from natural gas systems. Methane emissions from processing decreased by 45 percent from 1990 to 2017 as emissions from compressors (leaks and venting) and equipment leaks decreased, and increased 3 percent from 2016 to 2017 due to increased emissions from centrifugal and reciprocating compressors. Carbon dioxide emissions from processing decreased by 21 percent from 1990 to 2017, due to a decrease in acid gas removal emissions, and increased 3 percent from 2016 to 2017 due to increased emissions from flaring. Nitrous oxide emissions increased from 1990 to 2017, and have decreased 20 percent from 2016 to 2017.

*Transmission and Storage.* Natural gas transmission involves high pressure, large diameter pipelines that transport gas long distances from field production and processing areas to distribution systems or large volume customers such as power plants or chemical plants. Compressor station facilities are used to move the gas throughout the U.S. transmission system. Leak CH<sub>4</sub> emissions from these compressor stations and venting from pneumatic controllers account for most of the emissions from this stage. Uncombusted engine exhaust and pipeline venting are also sources of CH<sub>4</sub> emissions from transmission. Natural gas is also injected and stored in underground formations, or liquefied and stored in above ground tanks, during periods of low demand (e.g., summer), and withdrawn, processed, and distributed during periods of high demand (e.g., winter). Compressors and dehydrators are the primary contributors to emissions from storage. Emissions from LNG are also included under transmission and storage. Methane emissions from the transmission and storage sector account for approximately 20 percent of emissions from natural gas systems, while CO<sub>2</sub> emissions from transmission and storage account for 2 percent of the non-combustion CO<sub>2</sub> emissions from natural gas systems. CH<sub>4</sub> emissions from this source decreased by 43 percent

from 1990 to 2017 due to reduced compressor station emissions (including emissions from compressors and leaks), and decreased 6 percent from 2016 to 2017 due to reduced pipeline venting and the plugging of the Aliso Canyon leak. CO<sub>2</sub> emissions from transmission and storage have increased by 147 percent from 1990 to 2017, and by 45 percent from 2016 to 2017, due to increased emissions from LNG export terminals. The quantity of LNG exported from the U.S. increased by a factor of 12 from 1990 to 2017, and by 279 percent from 2016 to 2017. LNG emissions are about 2 percent of CH<sub>4</sub> and 74 percent of CO<sub>2</sub> emissions from transmission and storage in year 2017. Nitrous oxide emissions from transmission and storage increased by 79 percent from 1990 to 2017 and increased 22 percent from 2016 to 2017.

*Distribution.* Distribution pipelines take the high-pressure gas from the transmission system at “city gate” stations, reduce the pressure and distribute the gas through primarily underground mains and service lines to individual end users. There were 1,294,091 miles of distribution mains in 2017, an increase of nearly 350,000 miles since 1990 (PHMSA 2018). Distribution system emissions, which account for 7 percent of CH<sub>4</sub> emissions from natural gas systems and less than 1 percent of non-combustion CO<sub>2</sub> emissions, result mainly from leak emissions from pipelines and stations. An increased use of plastic piping, which has lower emissions than other pipe materials, has reduced both CH<sub>4</sub> and CO<sub>2</sub> emissions from this stage, as have station upgrades at metering and regulating (M&R) stations. Distribution system CH<sub>4</sub> emissions in 2017 were 73 percent lower than 1990 levels (changed from 43.5 MMT CO<sub>2</sub> Eq. to 11.9 MMT CO<sub>2</sub> Eq.) and 1 percent lower than 2016 emissions, while distribution CO<sub>2</sub> emissions in 2017 were also 73 percent lower than 1990 levels and 1 percent lower than 2016 Emissions. CO<sub>2</sub> emission from this segment are less than 0.1 MMT CO<sub>2</sub> Eq. across the time series.

Total CH<sub>4</sub> emissions for the five major stages of natural gas systems are shown in MMT CO<sub>2</sub> Eq. (Table 3-64) and kt (Table 3-65). Table 3-66 provides additional information on how the estimates in Table 3-62 were calculated. With recent updates to the Inventory, most emissions are calculated using a net emission approach. However, certain sources are still calculated with a potential emission approach. Table 3-66 shows the calculated potential CH<sub>4</sub> release (i.e., potential emissions before any controls are applied) from each stage, and the amount of CH<sub>4</sub> that is estimated to have been flared, captured, or otherwise controlled, and therefore not emitted to the atmosphere. Subtracting the value for CH<sub>4</sub> that is controlled, from the value for calculated potential release of CH<sub>4</sub>, results in the total net emissions values. More disaggregated information on potential emissions and emissions is available in Annex 3.6. See Methodology for Estimating CH<sub>4</sub> and CO<sub>2</sub> Emissions from Natural Gas Systems.

**Table 3-64: CH<sub>4</sub> Emissions from Natural Gas Systems (MMT CO<sub>2</sub> Eq.)<sup>a</sup>**

Stage	1990	2005	2013	2014	2015	2016	2017
<b>Exploration<sup>b</sup></b>	<b>4.0</b>	<b>10.9</b>	<b>3.0</b>	<b>1.0</b>	<b>1.0</b>	<b>0.7</b>	<b>1.2</b>
<b>Production</b>	<b>67.0</b>	<b>89.5</b>	<b>108.5</b>	<b>108.5</b>	<b>108.8</b>	<b>107.1</b>	<b>108.4</b>
Onshore Production	35.0	51.5	53.3	49.3	47.2	46.0	45.1
Offshore Production	3.5	4.3	3.8	3.8	3.8	3.8	3.8
Gathering and Boosting <sup>c</sup>	28.5	33.7	51.4	55.4	57.9	57.4	59.5
<b>Processing</b>	<b>21.3</b>	<b>11.6</b>	<b>10.8</b>	<b>11.1</b>	<b>11.1</b>	<b>11.4</b>	<b>11.7</b>
<b>Transmission and Storage</b>	<b>57.2</b>	<b>36.1</b>	<b>31.0</b>	<b>32.4</b>	<b>34.2</b>	<b>34.5</b>	<b>32.4</b>
<b>Distribution</b>	<b>43.5</b>	<b>23.3</b>	<b>12.3</b>	<b>12.2</b>	<b>12.0</b>	<b>12.0</b>	<b>11.9</b>
<b>Total</b>	<b>193.1</b>	<b>171.4</b>	<b>165.6</b>	<b>165.1</b>	<b>167.2</b>	<b>165.7</b>	<b>165.6</b>

<sup>a</sup> These values represent CH<sub>4</sub> emitted to the atmosphere. CH<sub>4</sub> that is captured, flared, or otherwise controlled (and not emitted to the atmosphere) has been calculated and removed from emission totals.

<sup>b</sup> Exploration includes well drilling, testing, and completions.

<sup>c</sup> Gathering and boosting includes gathering and boosting station routine vented and leak sources, gathering pipeline leaks and blowdowns, and gathering and boosting station episodic events.

Note: Totals may not sum due to independent rounding.

**Table 3-65: CH<sub>4</sub> Emissions from Natural Gas Systems (kt)<sup>a</sup>**

Stage	1990	2005	2013	2014	2015	2016	2017
<b>Exploration<sup>b</sup></b>	<b>162</b>	<b>437</b>	<b>119</b>	<b>39</b>	<b>42</b>	<b>28</b>	<b>49</b>
<b>Production</b>	<b>2,679</b>	<b>3,578</b>	<b>4,340</b>	<b>4,338</b>	<b>4,353</b>	<b>4,286</b>	<b>4,337</b>
Onshore Production	1,399	2,058	2,133	1,972	1,888	1,840	1,806
Offshore Production	141	173	151	151	151	151	151
Gathering and Boosting <sup>c</sup>	1,139	1,347	2,056	2,216	2,315	2,295	2,380
<b>Processing</b>	<b>853</b>	<b>464</b>	<b>432</b>	<b>443</b>	<b>443</b>	<b>456</b>	<b>469</b>

<b>Transmission and Storage</b>	<b>2,289</b>	<b>1,444</b>	<b>1,239</b>	<b>1,295</b>	<b>1,368</b>	<b>1,380</b>	<b>1,295</b>
<b>Distribution</b>	<b>1,741</b>	<b>932</b>	<b>494</b>	<b>487</b>	<b>481</b>	<b>480</b>	<b>475</b>
<b>Total</b>	<b>7,723</b>	<b>6,856</b>	<b>6,624</b>	<b>6,603</b>	<b>6,686</b>	<b>6,629</b>	<b>6,624</b>

<sup>a</sup> These values represent CH<sub>4</sub> emitted to the atmosphere. CH<sub>4</sub> that is captured, flared, or otherwise controlled (and not emitted to the atmosphere) has been calculated and removed from emission totals.

<sup>b</sup> Exploration includes well drilling, testing, and completions.

<sup>c</sup> Gathering and boosting includes gathering and boosting station routine vented and leak sources, gathering pipeline leaks and blowdowns, and gathering and boosting station episodic events.

Note: Totals may not sum due to independent rounding.

**Table 3-66: Calculated Potential CH<sub>4</sub> and Captured/Combusted CH<sub>4</sub> from Natural Gas Systems (MMT CO<sub>2</sub> Eq.)**

	<b>1990</b>	<b>2005</b>	<b>2013</b>	<b>2014</b>	<b>2015</b>	<b>2016</b>	<b>2017</b>
<b>Calculated Potential<sup>a</sup></b>	<b>193.1</b>	<b>182.0</b>	<b>179.0</b>	<b>178.4</b>	<b>180.5</b>	<b>179.1</b>	<b>179.0</b>
Exploration	4.0	10.9	3.0	1.0	1.0	0.7	1.2
Production	67.0	94.8	115.2	115.1	115.5	113.8	115.1
Processing	21.3	11.6	10.8	11.1	11.1	11.4	11.7
Transmission and Storage	57.2	41.4	37.7	39.1	40.9	41.2	39.1
Distribution	43.5	23.3	12.3	12.2	12.0	12.0	11.9
<b>Captured/Combusted</b>	<b>0.0</b>	<b>10.6</b>	<b>13.4</b>	<b>13.4</b>	<b>13.4</b>	<b>13.4</b>	<b>13.4</b>
Exploration	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Production	0.0	5.3	6.7	6.7	6.7	6.7	6.7
Processing	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Transmission and Storage	0.0	5.3	6.7	6.7	6.7	6.7	6.7
Distribution	0.0	0.0	0.0	0.0	0.0	0.0	0.0
<b>Net Emissions</b>	<b>193.1</b>	<b>171.4</b>	<b>165.6</b>	<b>165.1</b>	<b>167.2</b>	<b>165.7</b>	<b>165.6</b>
Exploration	4.0	10.9	3.0	1.0	1.0	0.7	1.2
Production	67.0	89.5	108.5	108.5	108.8	107.1	108.4
Processing	21.3	11.6	10.8	11.1	11.1	11.4	11.7
Transmission and Storage	57.2	36.1	31.0	32.4	34.2	34.5	32.4
Distribution	43.5	23.3	12.3	12.2	12.0	12.0	11.9

<sup>a</sup> In this context, “potential” means the total emissions calculated before voluntary reductions and regulatory controls are applied.

Note: Totals may not sum due to independent rounding.

**Table 3-67: Non-combustion CO<sub>2</sub> Emissions from Natural Gas Systems (MMT)**

<b>Stage</b>	<b>1990</b>	<b>2005</b>	<b>2013</b>	<b>2014</b>	<b>2015</b>	<b>2016</b>	<b>2017</b>
Exploration	0.4	1.8	1.3	0.8	0.3	0.2	0.5
Production	1.0	1.8	3.1	3.3	3.4	3.2	2.8
Processing	28.3	18.9	20.5	21.0	21.0	21.7	22.5
Transmission and Storage	0.2	0.2	0.3	0.3	0.3	0.4	0.5
Distribution	0.1	+	+	+	+	+	+
<b>Total</b>	<b>30.0</b>	<b>22.6</b>	<b>25.1</b>	<b>25.5</b>	<b>25.1</b>	<b>25.5</b>	<b>26.3</b>

+ Does not exceed 0.1 MMT CO<sub>2</sub> Eq.

Note: Totals may not sum due to independent rounding.

**Table 3-68: Non-combustion CO<sub>2</sub> Emissions from Natural Gas Systems (kt)**

<b>Stage</b>	<b>1990</b>	<b>2005</b>	<b>2013</b>	<b>2014</b>	<b>2015</b>	<b>2016</b>	<b>2017</b>
Exploration	409	1,756	1,281	843	291	194	483
Production	1,035	1,759	3,076	3,342	3,448	3,188	2,845
Processing	28,338	18,876	20,510	21,047	21,047	21,724	22,452
Transmission and Storage	216	219	267	272	271	368	533
Distribution	51	27	15	14	14	14	14
<b>Total</b>	<b>30,048</b>	<b>22,638</b>	<b>25,148</b>	<b>25,518</b>	<b>25,071</b>	<b>25,488</b>	<b>26,327</b>

Note: Totals may not sum due to independent rounding.

**Table 3-69: N<sub>2</sub>O Emissions from Natural Gas Systems (Metric Tons CO<sub>2</sub> Eq.)**

Stage	1990	2005	2013	2014	2015	2016	2017
Exploration	461	1,401	1,179	855	3,215	113	289
Production	162	900	2,330	1,997	2,773	1,019	937
Processing	NO	3,351	5,625	5,772	5,772	3,802	3,049
Transmission and Storage	257	309	341	344	347	377	461
Distribution	NO	NO	NO	NO	NO	NO	NO
<b>Total</b>	<b>880</b>	<b>5,961</b>	<b>9,476</b>	<b>8,969</b>	<b>12,107</b>	<b>5,311</b>	<b>4,735</b>

NO (Not Occurring)

Note: Totals may not sum due to independent rounding.

**Table 3-70: N<sub>2</sub>O Emissions from Natural Gas Systems (Metric Tons N<sub>2</sub>O)**

Stage	1990	2005	2013	2014	2015	2016	2017
Exploration	1.5	4.7	4.0	2.9	10.8	0.4	1.0
Production	0.5	3.0	7.8	6.7	9.3	3.4	3.1
Processing	NO	11.2	18.9	19.4	19.4	12.8	10.2
Transmission and Storage	0.9	1.0	1.1	1.2	1.2	1.3	1.5
Distribution	NO	NO	NO	NO	NO	NO	NO
<b>Total</b>	<b>3.0</b>	<b>20.0</b>	<b>31.8</b>	<b>30.1</b>	<b>40.6</b>	<b>17.8</b>	<b>15.9</b>

NO (Not Occurring)

Note: Totals may not sum due to independent rounding.

## Methodology

See Annex 3.6 for the full time series of emissions data, activity data, and emission factors, and additional information on methods and data sources—for example, the specific years of reporting data from EPA's Greenhouse Gas Reporting Program (GHGRP) that are used to develop certain factors.

This section provides a general overview of the methodology for natural gas emission estimates in the Inventory, which involves the calculation of CH<sub>4</sub>, CO<sub>2</sub>, and N<sub>2</sub>O emissions for over 100 emissions sources, and then the summation of emissions for each natural gas segment.

The approach for calculating emissions for natural gas systems generally involves the application of emission factors to activity data. For most sources, the approach uses technology-specific emission factors or emission factors that vary over time and take into account changes to technologies and practices, which are used to calculate net emissions directly. For others, the approach uses what are considered “potential methane factors” and reduction data to calculate net emissions.

*Emission Factors.* Key references for emission factors for CH<sub>4</sub> and non-combustion-related CO<sub>2</sub> emissions from the U.S. natural gas industry include a 1996 study published by the Gas Research Institute (GRI) and EPA (GRI/EPA 1996), the EPA's GHGRP (EPA 2018), and others.

The EPA/GRI study developed over 80 CH<sub>4</sub> emission factors to characterize emissions from the various components within the operating stages of the U.S. natural gas system. The EPA/GRI study was based on a combination of process engineering studies, collection of activity data, and measurements at representative gas facilities conducted in the early 1990s. Year-specific natural gas CH<sub>4</sub> compositions are calculated using U.S. Department of Energy's Energy Information Administration (EIA) annual gross production for National Energy Modeling System (NEMS) oil and gas supply module regions in conjunction with data from the Gas Technology Institute (GTI, formerly GRI) Unconventional Natural Gas and Gas Composition Databases (GTI 2001). These year-specific CH<sub>4</sub> compositions are applied to emission factors, which therefore may vary from year to year due to slight changes in the CH<sub>4</sub> composition for each NEMS region.

GHGRP Subpart W data were used to develop CH<sub>4</sub>, CO<sub>2</sub>, and N<sub>2</sub>O emission factors for many sources in the Inventory. In the exploration and production segments, GHGRP data were used to develop emission factors used for

all time series years for well testing, gas well completions and workovers with and without hydraulic fracturing, pneumatic controllers and chemical injection pumps, condensate tanks, liquids unloading, miscellaneous flaring, and gathering and boosting pipelines. In the processing segment, for recent years of the times series, GHGRP data were used to develop emission factors for fugitives, compressors, flares, dehydrators, and blowdowns/venting. In the transmission and storage segment, GHGRP data were used to develop factors for all time series years for LNG stations and terminals and transmission pipeline blowdowns, and for pneumatic controllers for recent years of the times series.

Other data sources used for CH<sub>4</sub> emission factors include Zimmerle et al. (2015) for transmission and storage station fugitives and compressors, and Lamb et al. (2015) for recent years for distribution pipelines and meter/regulator stations.

For CO<sub>2</sub> emissions from sources in the exploration, production and processing segments that use emission factors not directly calculated from GHGRP data, data from the 1996 GRI/EPA study and a 2001 GTI publication were used to adapt the CH<sub>4</sub> emission factors into non-combustion related CO<sub>2</sub> emission factors. For sources in the transmission and storage segment that use emission factors not directly calculated from GHGRP data, and for sources in the distribution segment, data from the 1996 GRI/EPA study and a 1993 GTI publication were used to adapt the CH<sub>4</sub> emission factors into non-combustion related CO<sub>2</sub> emission factors.

Flaring N<sub>2</sub>O emissions were estimated for flaring sources using GHGRP data.

See Annex 3.6 for more detailed information on the methodology and data used to calculate CH<sub>4</sub> and non-combustion CO<sub>2</sub> and N<sub>2</sub>O emissions from natural gas systems.

*Activity Data.* Activity data were taken from various published data sets, as detailed in Annex 3.6. Key activity data sources include data sets developed and maintained by EPA's GHGRP; DrillingInfo, Inc. (DrillingInfo 2018); U.S. Department of the Interior's Bureau of Ocean Energy Management, Regulation and Enforcement (BOEMRE, previously Minerals and Management Service); Federal Energy Regulatory Commission (FERC); EIA; the Natural Gas STAR Program annual emissions savings data; Oil and Gas Journal; PHMSA; the Wyoming Conservation Commission; and the Alabama State Oil and Gas Board.

For a few sources, recent direct activity data are not available. For these sources, either 2016 data were used as a proxy for 2017 data, or a set of industry activity data drivers was developed and used to calculate activity data over the time series. Drivers include statistics on gas production, number of wells, system throughput, miles of various kinds of pipe, and other statistics that characterize the changes in the U.S. natural gas system infrastructure and operations. More information on activity data and drivers is available in Annex 3.6.

A complete list of references for emission factors and activity data by emission source is provided in Annex 3.6.

*Calculating Net Emissions.* For most sources, net emissions are calculated directly by applying emission factors to activity data. Emission factors used in net emission approaches reflect technology-specific information, and take into account regulatory and voluntary reductions. However, for certain sectors, some sources are calculated using potential emission factors, and the step of deducting CH<sub>4</sub> that is not emitted from the total CH<sub>4</sub> potential estimates to develop net CH<sub>4</sub> emissions is applied. To take into account use of such technologies and practices that result in lower emissions but are not reflected in "potential" emission factors, data are collected on both regulatory and voluntary reductions. Regulatory actions addressed using this method include National Emission Standards for Hazardous Air Pollutants (NESHAP) regulations for dehydrator vents. Voluntary reductions included in the Inventory are those reported to Natural Gas STAR for certain sources in the production and transmission and storage segments.

In fall of 2015, a well in a California storage field began leaking methane at an initial average rate of around 50 metric tons (MT) of methane (CH<sub>4</sub>) an hour, and continued leaking until it was permanently sealed in February of 2016.<sup>85</sup> An emission estimate from the leak event was included for 2015 and 2016, using the estimate of the leak

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<sup>85</sup> For more information on the Aliso Canyon event, and the measurements conducted of the leak, please see Ensuring Safe and Reliable Underground Natural Gas Storage, *Final Report of the Interagency Task Force on Natural Gas Storage Safety*, available at <<http://www.energy.gov/sites/prod/files/2016/10/f33/Ensuring%20Safe%20and%20Reliable%20Underground%20Natural%20Gas%20Storage%20-%20Final%20Report.pdf>>.

published by the California Air Resources Board (99,638 MT CH<sub>4</sub> for the duration of the leak). The 2015 and 2016 emission estimates of 78,350 MT CH<sub>4</sub> and 21,288 MT CH<sub>4</sub>, respectively, were added to the 2015 and 2016 estimates of fugitive emissions from storage wells. For more information, please see *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2015: Update for Storage Segment Emissions*.<sup>86</sup>

Through EPA’s stakeholder process on oil and gas in the Inventory, EPA received initial stakeholder feedback on updates under consideration for the Inventory. Stakeholder feedback is noted below in Uncertainty and Time-Series Consistency, Recalculations Discussion, and Planned Improvements.

## Uncertainty and Time-Series Consistency

EPA has conducted a quantitative uncertainty analysis using the IPCC Approach 2 methodology (Monte Carlo Simulation technique) to characterize the uncertainty for natural gas systems. For more information, please see the memorandum *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2016: Natural Gas and Petroleum Systems Uncertainty Estimates (2018 Uncertainty Memo)*.<sup>87</sup> EPA used Microsoft Excel's @RISK add-in tool to estimate the 95 percent confidence bound around CH<sub>4</sub> emissions from natural gas systems for the current Inventory, then applied the calculated bounds to both CH<sub>4</sub> and CO<sub>2</sub> emissions estimates. For the analysis, EPA focused on the 14 highest-emitting sources for the year 2016, which together emitted 76 percent of methane from natural gas systems in 2017, and extrapolated the estimated uncertainty for the remaining sources. The @RISK add-in provides for the specification of probability density functions (PDFs) for key variables within a computational structure that mirrors the calculation of the inventory estimate. The IPCC guidance notes that in using this method, "some uncertainties that are not addressed by statistical means may exist, including those arising from omissions or double counting, or other conceptual errors, or from incomplete understanding of the processes that may lead to inaccuracies in estimates developed from models." The uncertainty bounds reported below only reflect those uncertainties that EPA has been able to quantify and do not incorporate considerations such as modeling uncertainty, data representativeness, measurement errors, misreporting or misclassification. The understanding of the uncertainty of emission estimates for this category evolves and improves as the underlying methodologies and datasets improve.

The results presented below provide the 95 percent confidence bound within which actual emissions from this source category are likely to fall for the year 2017, using the IPCC methodology. The results of the Approach 2 uncertainty analysis are summarized in Table 3-71. Natural gas systems CH<sub>4</sub> emissions in 2017 were estimated to be between 141.8 and 193.3 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. Natural gas systems CO<sub>2</sub> emissions in 2017 were estimated to be between 22.5 and 30.7 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. Uncertainty bounds for other years of the time series have not been calculated, but uncertainty is expected to vary over the time series. For example, years where many emission sources are calculated with interpolated data would likely have higher uncertainty than years with predominantly year-specific data.

**Table 3-71: Approach 2 Quantitative Uncertainty Estimates for CH<sub>4</sub> and Non-combustion CO<sub>2</sub> Emissions from Natural Gas Systems (MMT CO<sub>2</sub> Eq. and Percent)**

Source	Gas	2017 Emission Estimate (MMT CO <sub>2</sub> Eq.) <sup>b</sup>	Uncertainty Range Relative to Emission Estimate <sup>a</sup> (MMT CO <sub>2</sub> Eq.) (%)			
			Lower Bound <sup>b</sup>	Upper Bound <sup>b</sup>	Lower Bound <sup>b</sup>	Upper Bound <sup>b</sup>
Natural Gas Systems	CH <sub>4</sub>	165.6	141.8	193.3	-14%	+17%
Natural Gas Systems <sup>c</sup>	CO <sub>2</sub>	26.3	22.5	30.7	-14%	+17%

<sup>a</sup> Range of emission estimates estimated by applying the 95 percent confidence intervals obtained from the Monte Carlo Simulation analysis conducted for the year 2017 CH<sub>4</sub> emissions.

<sup>b</sup> All reported values are rounded after calculation. As a result, lower and upper bounds may not be duplicable from other rounded values as shown in Table 3-64 and Table 3-65.

<sup>86</sup> <<https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems-ghg-inventory-additional-information-1990-2015-ghg>>.

<sup>87</sup> See <<https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems>>.

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<sup>c</sup> An uncertainty analysis for the CO<sub>2</sub> emissions was not performed. The relative uncertainty estimated (expressed as a percent) from the CH<sub>4</sub> uncertainty analysis was applied to the point estimate of CO<sub>2</sub> emissions.

GHGRP data available (starting in 2011) and other recent data sources have improved estimates of emissions from natural gas systems. To develop a consistent time series, for sources with new data, EPA reviewed available information on factors that may have resulted in changes over the time series (e.g., regulations, voluntary actions) and requested stakeholder feedback on trends as well. For most sources, EPA developed annual data for 1993 through 2010 by interpolating activity data or emission factors or both between 1992 and 2011 data points. Information on time-series consistency for sources updated in this year's Inventory can be found in the Recalculations Discussion below, with additional detail provided in supporting memos (relevant memos are cited in the Recalculations Discussion). For detailed documentation of methodologies, please see Annex 3.5.

## QA/QC and Verification Discussion

The natural gas emission estimates in the Inventory are continually being reviewed and assessed to determine whether emission factors and activity factors accurately reflect current industry practices. A QA/QC analysis was performed for data gathering and input, documentation, and calculation. QA/QC checks are consistently conducted to minimize human error in the model calculations. EPA performs a thorough review of information associated with new studies, GHGRP data, regulations, public webcasts, and the Natural Gas STAR Program to assess whether the assumptions in the Inventory are consistent with current industry practices. The EPA has a multi-step data verification process for GHGRP data, including automatic checks during data-entry, statistical analyses on completed reports, and staff review of the reported data. Based on the results of the verification process, the EPA follows up with facilities to resolve mistakes that may have occurred.<sup>88</sup>

As in previous years, EPA conducted early engagement and communication with stakeholders on updates prior to public review. EPA held a stakeholder workshop on greenhouse gas data for oil and gas in October of 2018, and webinars in June of 2018 and February of 2019. EPA released memos detailing updates under consideration and requesting stakeholder feedback. Stakeholder feedback received through these processes is discussed in the Recalculations Discussion and Planned Improvements sections below.

In recent years, several studies have measured emissions at the source level and at the national or regional level and calculated emission estimates that may differ from the Inventory. There are a variety of potential uses of data from new studies, including replacing a previous estimate or factor, verifying or QA of an existing estimate or factor, and identifying areas for updates. In general, there are two major types of studies related to oil and gas greenhouse gas data: studies that focus on measurement or quantification of emissions from specific activities, processes and equipment, and studies that use tools such as inverse modeling to estimate the level of overall emissions needed to account for measured atmospheric concentrations of greenhouse gases at various scales. The first type of study can lead to direct improvements to or verification of Inventory estimates. In the past few years, EPA has reviewed and in many cases, incorporated data from these data sources. The second type of study can provide general indications on potential over- and under-estimates. A key challenge in using these types of studies to assess Inventory results is having a relevant basis for comparison (i.e., the independent study should assess data from the Inventory and not another data set, such as EDGAR.). In an effort to improve the ability to compare the national-level inventory with measurement results that may be at other scales, a team at Harvard University along with EPA and other coauthors developed a gridded inventory of U.S. anthropogenic methane emissions with 0.1° x 0.1° spatial resolution, monthly temporal resolution, and detailed scale-dependent error characterization.<sup>89</sup> The gridded methane inventory is designed to be consistent with the 2016 *Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1990-2014* estimates for the year 2012, which presents national totals.<sup>90</sup>

## Recalculations Discussion

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<sup>88</sup> See <[https://www.epa.gov/sites/production/files/2015-07/documents/ghgrp\\_verification\\_factsheet.pdf](https://www.epa.gov/sites/production/files/2015-07/documents/ghgrp_verification_factsheet.pdf)>.

<sup>89</sup> See <<https://www.epa.gov/ghgemissions/gridded-2012-methane-emissions>>.

<sup>90</sup> See <<https://www.epa.gov/ghgemissions/us-greenhouse-gas-inventory-report-1990-2014>>.

EPA received information and data related to the emission estimates through GHGRP reporting, the annual Inventory formal public notice periods, stakeholder feedback on updates under consideration, and new studies. In June, October and November 2018, EPA released draft memoranda that discussed changes under consideration, and requested stakeholder feedback on those changes. EPA then created updated versions of the memoranda to document the methodology implemented into the current Inventory.<sup>91</sup> Memoranda cited in the Recalculations Discussion below are: *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2017: Update for Natural Gas Gathering & Boosting Emissions (April 2019 G&B memo)*, *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2017: Update for Liquefied Natural Gas Segment Emissions (April 2019 LNG memo)*, and *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2017: Other Updates Considered for 2019 and Future GHGs (April 2019 Other Updates memo)*.

EPA thoroughly evaluated relevant information available and made several updates to the Inventory, including: using GHGRP data to calculate emissions from gathering pipelines, transmission pipeline blowdowns, and LNG storage stations and terminals; calculating new N<sub>2</sub>O emission factors for flaring sources throughout all segments directly from GHGRP data; and updating the data source for well drilling activity. In addition, certain sources did not undergo methodological updates, but CH<sub>4</sub> and/or CO<sub>2</sub> emissions changed by greater than 0.05 MMT CO<sub>2</sub> Eq., comparing the previous estimate for 2016 to the current (recalculated) estimate for 2016 (the emissions changes were mostly due to GHGRP data submission revisions); these sources are discussed below and include hydraulically fractured (HF) gas well completions, production segment miscellaneous flaring, production segment pneumatic controllers, liquids unloading, production segment storage tanks, G&B stations, acid gas removal (AGR) vents and flares at gas processing plants, and gas engines in the production and processing segments. Lastly, for HF gas well workovers, year 2017 emissions estimates are noticeably higher than previous years; the factors driving this increase are described below.

The combined impact of revisions to 2016 natural gas sector CH<sub>4</sub> emissions, compared to the previous Inventory, is an increase from 163.5 to 165.7 MMT CO<sub>2</sub> Eq. (2.2 MMT CO<sub>2</sub> Eq., or 1 percent). The recalculations resulted in an average increase in CH<sub>4</sub> emission estimates across the 1990 through 2016 time series, compared to the previous Inventory, of 0.6 MMT CO<sub>2</sub> Eq., or 0.4 percent.

The combined impact of revisions to 2016 natural gas sector CO<sub>2</sub> emissions, compared to the previous Inventory, is minimal, with emissions of approximately 25.5 MMT CO<sub>2</sub> in both Inventories. The recalculations resulted in an average increase in emission estimates across the 1990 through 2016 time series, compared to the previous Inventory, of 0.2 MMT CO<sub>2</sub> Eq., or 0.7 percent.

In Table 3-72 and Table 3-73 below are categories in Natural Gas Systems with recalculations resulting in a change of greater than 0.05 MMT CO<sub>2</sub> Eq., comparing the previous estimate for 2016 to the current (recalculated) estimate for 2016. For more information, please see the Recalculations Discussion below.

**Table 3-72: Recalculations of CO<sub>2</sub> in Natural Gas Systems (MMT CO<sub>2</sub>)**

<b>Stage and Emission Source</b>	<b>Previous Estimate Year 2016, 2018 Inventory</b>	<b>Current Estimate Year 2016, 2019 Inventory</b>	<b>Current Estimate Year 2017, 2019 Inventory</b>
<b>Exploration</b>	<b>0.1</b>	<b>0.2</b>	<b>0.5</b>
HF Completions	0.1	0.2	0.5
<b>Production</b>	<b>3.2</b>	<b>3.2</b>	<b>2.8</b>
Gathering Pipelines	+	+	+
Miscellaneous Flaring Tanks	1.1	1.2	1.1
HF Workovers	+	0.1	0.4
<b>Processing</b>	<b>22.0</b>	<b>21.7</b>	<b>22.5</b>
AGR Vents	16.6	16.5	16.7
Flares	5.4	5.2	5.7
<b>Transmission and Storage</b>	<b>0.1</b>	<b>0.4</b>	<b>0.5</b>
LNG Storage	+	+	+

<sup>91</sup> Stakeholder materials including draft and final EPA memoranda for the current (i.e., 1990 to 2017) Inventory are available at <<https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems>>.



LNG Import/Export Terminals	+	0.2	0.4
Pipeline Blowdowns	+	+	+
<b>Distribution</b>	+	+	+
<b>Total</b>		<b>25.5</b>	<b>25.5</b>
			<b>26.3</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub>.

**Table 3-73: Recalculations of CH<sub>4</sub> in Natural Gas Systems (MMT CO<sub>2</sub> Eq.)**

Stage and Emission Source	Previous Estimate	Current Estimate	Current Estimate
	Year 2016, 2018 Inventory	Year 2016, 2019 Inventory	Year 2017, 2019 Inventory
<b>Exploration</b>	<b>0.7</b>	<b>0.7</b>	<b>1.2</b>
<b>Production</b>	<b>106.8</b>	<b>107.1</b>	<b>108.4</b>
G&B Stations	53.7	53.6	55.5
Gathering Pipelines	4.0	3.8	4.0
Pneumatic Controllers	26.3	26.6	26.4
Liquids Unloading	3.3	3.3	2.9
HF Workovers	0.4	0.4	0.8
Gas Engines	2.7	3.0	2.8
<b>Processing</b>	<b>11.2</b>	<b>11.4</b>	<b>11.7</b>
Gas Engines	6.1	6.3	6.4
<b>Transmission and Storage</b>	<b>32.8</b>	<b>34.5</b>	<b>32.4</b>
LNG Storage	1.8	0.2	0.3
LNG Import/Export Terminals	0.3	0.4	0.4
Pipeline Blowdowns	4.6	6.3	4.6
<b>Distribution</b>	<b>12.0</b>	<b>12.0</b>	<b>11.9</b>
<b>Total</b>	<b>163.5</b>	<b>165.7</b>	<b>165.6</b>

## Exploration

### *Well Drilling (Methodological Update)*

EPA updated the methodology for estimating the number of gas wells drilled across the time series to use DrillingInfo data (DrillingInfo 2018). The new methodology is detailed in the *Apr. 2019 Other Updates* memo. In previous Inventories, the U.S. Department of Energy's Energy Information Administration (DOE/EIA) *Monthly Energy Review* well drilling activity data set was used to develop well drilling activity inputs, but this publication does not provide data after year 2010. EPA therefore developed a methodology of analyzing DrillingInfo data to estimate counts of gas wells drilled in each time series year, 1990 through 2017. These activity data for select years are shown in Table 3-74 below.

**Table 3-74: Count of Gas Wells Drilled**

Activity	1990	2005	2013	2014	2015	2016	2017
Gas Wells Drilled	17,805	27,568	5,681	5,871	3,585	2,264	2,264
<i>Previous Estimate</i>	<i>15,096</i>	<i>31,969</i>	<i>18,837<sup>a</sup></i>	<i>18,837<sup>a</sup></i>	<i>18,837<sup>a</sup></i>	<i>18,837<sup>a</sup></i>	<i>NA</i>

a – Year-specific data not available; the year 2010 estimate was assigned as a surrogate value.

NA (Not Applicable)

### *HF Gas Well Completions (Recalculation with Updated Data)*

HF gas well completion CO<sub>2</sub> emission estimates increased 47 percent in the current Inventory for year 2016, compared to the previous Inventory, due to GHGRP submission revisions. Specifically, the GHGRP submission revisions reported higher CO<sub>2</sub> emissions for HF reduced emission completions with flaring, which led to a larger CO<sub>2</sub> emission factor. Compared to the previous Inventory, for 1990 to 2015, the CO<sub>2</sub> emission estimates increased by an average of only 0.3 percent.

**Table 3-75: HF Gas Well Completions National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
HF Completions - Non-REC with Venting	10	27	11	2	+	+	+
HF Completions - Non-REC with Flaring	390	1,316	324	327	58	12	37
HF Completions - REC with Venting	NO	3	2	+	1	+	1
HF Completions - REC with Flaring	NO	394	929	502	218	164	438
<b>Total Emissions</b>	<b>400</b>	<b>1,741</b>	<b>1,265</b>	<b>832</b>	<b>277</b>	<b>177</b>	<b>475</b>
<i>Previous Estimate</i>	397	1,748	1,148	844	277	120	NA

NO (Not Occurring)

NA (Not Applicable)

+ Does not exceed 0.5 kt CO<sub>2</sub>.

## Production

### *Gathering Pipelines (Methodological Update)*

EPA developed new activity data and net emission factors for gathering pipeline sources (leaks and blowdowns) using GHGRP data, as detailed in the *Apr. 2019 G&B* memo. Accordingly, the updated methodology no longer incorporates data on the Gas STAR reductions from pipeline leaks. Using GHGRP data to estimate gathering pipeline emissions was supported by stakeholder feedback. Compared to the previous Inventory, gathering pipeline CH<sub>4</sub> emission estimates decreased for recent years due to the newly calculated emission factors from GHGRP and increased for early years due to updated well count activity data that drives pipeline mileage estimates. Compared with the previous Inventory, on average, CH<sub>4</sub> emission estimates decreased by 6 percent across the 1990 to 2016 time series. Compared to the previous Inventory, gathering pipeline CO<sub>2</sub> emission estimates decreased by 10 percent across the 1990 to 2016 time series, or by an average of 1.7 MMT CO<sub>2</sub>. See the *Apr. 2019 G&B* memo for additional discussion.

**Table 3-76: Gathering Pipelines National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
G&B Pipeline Leaks	78,425	117,182	138,669	137,477	136,513	136,776	141,577
G&B Pipeline Blowdowns	8,436	12,605	14,917	14,788	14,685	14,713	19,777
<b>Total Emissions</b>	<b>86,861</b>	<b>129,787</b>	<b>153,586</b>	<b>152,266</b>	<b>151,198</b>	<b>151,489</b>	<b>161,354</b>
<i>Previous Estimate</i>	85,413	136,627	164,443	164,727	162,796	160,311	NA

NA (Not Applicable)

### *Gathering and Boosting Stations (Recalculation with Updated Data)*

G&B station CH<sub>4</sub> emission estimates decreased by 0.3 percent in the current Inventory for year 2016, compared to the previous Inventory. This change was not the result of a methodological update, but due to updated data for marketed onshore gas production, which drives the station count activity data. EPA presented approaches to use GHGRP data to estimate G&B station emissions, but stakeholder feedback supported maintaining the current Inventory methodology (see the *April 2019 G&B memo*). Additional G&B station considerations for future Inventories, particularly for estimating CO<sub>2</sub> and N<sub>2</sub>O emissions, are discussed in the Planned Improvements section below.

**Table 3-77: Gathering Stations National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Gathering and Boosting Stations	956,870	1,107,208	1,730,573	1,877,554	1,968,205	1,949,925	2,018,566
G&B Station Episodic Events	94,905	109,816	171,643	186,221	195,212	193,399	200,207
<b>Total Emissions</b>	<b>1,051,775</b>	<b>1,217,024</b>	<b>1,902,216</b>	<b>2,063,775</b>	<b>2,163,417</b>	<b>2,143,324</b>	<b>2,218,773</b>

<i>Previous Estimate</i>	1,051,775	1,217,024	1,902,216	2,063,775	2,163,417	2,149,065	NA
NA (Not Applicable)							

### Miscellaneous Production Flaring (Recalculation with Updated Data)

Miscellaneous production flaring CO<sub>2</sub> emission estimates decreased in the current Inventory for 1990 to 2015 and increased in the current Inventory for 2016, compared to the previous Inventory. The CO<sub>2</sub> emissions changes are due to GHGRP submission revisions and use of GHGRP well counts from the facility overview table (see the Well Counts discussion below). In addition, the emission calculations are performed at a basin-level, and the changes impacted each basin uniquely.

**Table 3-78: Miscellaneous Production Flaring National Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Miscellaneous Flaring-Gulf Coast Basin	NO	155	250	296	331	243	193
Miscellaneous Flaring-Williston Basin	NO	+	+	+	+	NO	10
Miscellaneous Flaring-Permian Basin	NO	256	434	535	644	506	579
Miscellaneous Flaring-Other Basins	NO	118	293	319	343	438	308
<b>Total Emissions</b>	<b>NO</b>	<b>530</b>	<b>978</b>	<b>1,150</b>	<b>1,317</b>	<b>1,186</b>	<b>1,090</b>
<i>Previous Estimate</i>	<i>NO</i>	<i>572</i>	<i>1,057</i>	<i>1,241</i>	<i>1,415</i>	<i>1,129</i>	<i>NA</i>

NO (Not Occurring)

NA (Not Applicable)

+ Does not exceed 0.5 kt CO<sub>2</sub>.

### Gas Engines (Recalculation with Updated Data)

Natural gas engine CH<sub>4</sub> emission estimates increased in the current Inventory by an average of approximately 4 percent across the time series, compared to the previous Inventory. This change was due to the updated DrillingInfo gas wells counts (see the Well Counts discussion below).

**Table 3-79: Production Segment Gas Engines National Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Gas Engines	116,539	123,210	131,262	128,812	125,437	118,462	113,758
<i>Previous Estimate</i>	<i>116,508</i>	<i>117,852</i>	<i>121,827</i>	<i>118,818</i>	<i>114,774</i>	<i>106,423</i>	<i>NA</i>

NA (Not Applicable)

### Pneumatic Controllers (Recalculation with Updated Data)

Pneumatic controller CH<sub>4</sub> emission estimates increased in the current Inventory by an average of approximately 2 percent across the time series, compared to the previous Inventory. This change was caused by several factors: GHGRP submission revisions, the use of GHGRP well counts from the facility overview table (see the Well Counts discussion below), a correction to the linear interpolation calculation for activity factors in years 1993 through 2010, and updated DrillingInfo gas well counts (see the Well Counts discussion below).

**Table 3-80: Production Segment Pneumatic Controller National Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Low Bleed	NO	23,541	27,554	32,330	30,455	32,646	33,944
High Bleed	297,952	450,013	177,784	129,712	101,930	107,162	107,398
Intermittent Bleed	194,302	531,907	970,065	927,297	943,216	924,261	915,961
<b>Total Emissions</b>	<b>492,254</b>	<b>1,005,461</b>	<b>1,175,402</b>	<b>1,089,339</b>	<b>1,075,601</b>	<b>1,064,069</b>	<b>1,057,303</b>
<i>Previous Estimate</i>	<i>506,905</i>	<i>981,773</i>	<i>1,134,147</i>	<i>1,072,375</i>	<i>1,055,935</i>	<i>1,053,207</i>	<i>NA</i>

NO (Not Occurring)

NA (Not Applicable)

### Liquids Unloading (Recalculation with Updated Data)

Liquids unloading CH<sub>4</sub> emission estimates increased for 2015 and decreased for 2016 in the current Inventory, compared to the previous Inventory. Compared to the previous Inventory, on average across the time series, liquids unloading CH<sub>4</sub> emission estimates increased by 2 percent. These changes were due to GHGRP submission revisions and the use of GHGRP well counts from the facility overview table (see the Well Counts discussion below). In particular, the percent of gas wells requiring liquids unloading increased for the GHGRP reporting year 2015 data (which is applied to all prior years of the time series) and decreased for reporting year 2016.

**Table 3-81: Liquids Unloading National Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Unloading with Plunger Lifts	NO	126,009	124,036	80,880	63,089	61,397	46,843
Unloading without Plunger Lifts	372,325	247,433	110,095	129,904	97,616	69,381	70,536
<b>Total Emissions</b>	<b>372,325</b>	<b>373,442</b>	<b>234,132</b>	<b>210,784</b>	<b>160,706</b>	<b>130,778</b>	<b>117,379</b>
<i>Previous Estimated Emissions</i>	<i>379,837</i>	<i>365,310</i>	<i>220,990</i>	<i>202,745</i>	<i>153,975</i>	<i>132,871</i>	<i>NA</i>

NO (Not Occurring)  
NA (Not Applicable)

### Tanks (Recalculation with Updated Data)

Production tank CO<sub>2</sub> emission estimates increased by an average of approximately 30 percent across 1990 to 2015 in the current Inventory and decreased by about 8 percent in the current Inventory for 2016, compared to the previous Inventory. The change in production tank CO<sub>2</sub> emission estimates is mainly driven by GHGRP submission revisions. For example, GHGRP reporting year 2015 CO<sub>2</sub> emission estimates increased, which led to an increase in the calculated emission factors, and year 2015 emission factors are applied to all prior years of the time series.

**Table 3-82: Production Segment Storage Tanks National Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Large Tanks w/Flares	287	363	984	1,030	1,041	1,080	558
Large Tanks w/VRU	NO	1	3	3	3	2	+
Large Tanks w/o Control	167	90	147	154	155	1	+
Small Tanks w/Flares	NO	8	30	31	31	33	22
Small Tanks w/o Flares	6	4	9	10	10	12	5
Malfunctioning Separator Dump Valves	+	+	+	+	+	+	1
<b>Total Emissions</b>	<b>460</b>	<b>466</b>	<b>1,173</b>	<b>1,227</b>	<b>1,240</b>	<b>1,129</b>	<b>585</b>
<i>Previous Estimate</i>	<i>294</i>	<i>378</i>	<i>1,030</i>	<i>1,078</i>	<i>1,089</i>	<i>1,224</i>	<i>NA</i>

NO (Not Occurring)  
NA (Not Applicable)  
+ Does not exceed 0.5 kt CO<sub>2</sub>.

### HF Gas Well Workovers (Year 2017 Emissions)

Recalculations of HF gas well workover emissions did not result in large changes across the 1990 to 2016 time series when comparing the current Inventory to the previous Inventory. The large increase for HF gas well workover emissions in 2017 seen in the tables below results from the use of GHGRP data reported for 2017, a year with a sharp increase in reduced emission workovers with flaring.

**Table 3-83: HF Gas Well Workovers National Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
HF Workovers - Non-REC with Venting	25,823	66,053	69,935	25,517	2,518	7,878	11,795
HF Workovers - Non-REC with Flaring	366	1,034	350	476	225	76	527
HF Workovers - REC with Venting	NO	625	2,711	589	8,035	6,301	17,193

HF Workovers - REC with Flaring	NO	5	281	26	2,383	1,297	4,197
<b>Total Emissions</b>	<b>26,188</b>	<b>67,717</b>	<b>73,276</b>	<b>26,608</b>	<b>13,161</b>	<b>15,551</b>	<b>33,711</b>
<i>Previous Estimate</i>	25,244	66,781	72,557	26,957	13,228	16,986	NA

NO (Not Occurring)  
NA (Not Applicable)

**Table 3-84: HF Gas Well Workovers National Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
HF Workovers - Non-REC with Venting	2	4	8	2	+	+	2
HF Workovers - Non-REC with Flaring	66	187	70	156	17	12	41
HF Workovers - REC with Venting	NO	+	+	+	+	+	+
HF Workovers - REC with Flaring	NO	1	55	5	59	47	313
<b>Total Emissions</b>	<b>68</b>	<b>193</b>	<b>133</b>	<b>163</b>	<b>77</b>	<b>59</b>	<b>356</b>
<i>Previous Estimate</i>	65	190	125	156	77	44	NA

NO (Not Occurring)  
NA (Not Applicable)  
+ Does not exceed 0.5 kt CO<sub>2</sub>.

### Well Counts (Recalculation with Updated Data)

For total national well counts, EPA has used a more recent version of the DrillingInfo data set (DrillingInfo 2018) to update well counts data in the Inventory. EPA also updated the DrillingInfo data processing methodology to more accurately count wells in states with lease-level reporting (e.g., Kansas), which resulted in slight increased counts across the time series. While this was not a significant recalculation (increases are 2 to 3 percent across the time series), this is a key input to the Inventory, so results are highlighted here.

**Table 3-85: Producing Gas Well Count Data**

Activity	1990	2005	2013	2014	2015	2016	2017
Number of Gas Wells	193,718	346,862	428,947	424,308	420,418	419,005	411,450
<i>Previous Estimate</i>	197,626	348,470	427,828	431,446	425,651	416,881	NA

NA (Not Applicable)

In October 2018, EIA released an updated time series of national oil and gas well counts (covering 2000 through 2017). EIA estimates 991,000 total producing wells for year 2017. EPA's total well count for this year is 978,176. EPA's well counts in recent time series years are generally 2 percent lower than EIA's. EIA's well counts include side tracks, completions, and recompletions, and therefore are expected to be higher than EPA's which include only producing wells. EPA and EIA use a different threshold for distinguishing between oil versus gas (EIA uses 6 mcf/bbl, while EPA uses 100 mcf/bbl), which results in EIA having a lower fraction of oil wells and a higher fraction of gas wells than EPA.

For the count of wells included in GHGRP reporting (used to develop wellhead-based emissions and activity factors), EPA previously referenced the wellhead counts contained within the reporting table for onshore production equipment leak emissions. Due to updated reporting requirements for year 2017 forward, well counts provided as part of the facility overview information (i.e., wells producing at the end of the calendar year plus wells removed from production in a given year) provide more complete estimates. Therefore, EPA used well counts from the facility overview table for source-specific methodologies that rely on GHGRP reported well counts in the current Inventory. Comparing the GHGRP well counts from the facility overview table to the equipment leaks table: a larger population of the wells were reported as "oil" production type in the facility overview information table, compared to the equipment leaks table, which generally led to decreased activity and emissions for natural gas systems; for example, as discussed in the sections above, production segment emissions from pneumatic controllers and miscellaneous production flaring decreased across most of the time series.

## Processing

### *Acid Gas Removal (Recalculation with Updated Data)*

Acid gas removal unit (AGR) CO<sub>2</sub> emission estimates were essentially unchanged across the 1990 to 2015 time series, comparing the current Inventory to the previous Inventory, with an average increase of less than 0.01 percent. There was a decrease in CO<sub>2</sub> emission estimates for 2016, comparing the current Inventory to the previous Inventory. This decrease in CO<sub>2</sub> emission estimates for 2016 is due to GHGRP submission revisions, where a lower emission factor was calculated from the GHGRP data.

**Table 3-86: AGR National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Acid Gas Removal	28,282	15,320	14,565	14,946	14,946	16,481	16,728
<i>Previous Estimate</i>	28,282	15,320	14,565	14,946	14,946	16,565	NA

NA (Not Applicable)

### *Flares (Recalculation with Updated Data)*

Processing segment flare CO<sub>2</sub> emission estimates increased by only 0.03 percent across the 1990 to 2015 time series in the current Inventory and decreased by approximately 4 percent for 2016 in the current Inventory, compared to the previous Inventory. This decrease in CO<sub>2</sub> emission estimates for 2016 is due to GHGRP submission revisions, where a lower emission factor was calculated from the GHGRP data.

**Table 3-87: Processing Segment Flares National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Flares	NO	3,517	5,904	6,058	6,058	5,203	5,683
<i>Previous Estimate</i>	NO	3,516	5,902	6,056	6,056	5,404	NA

NO (Not Occurring)  
NA (Not Applicable)

### *Gas Engines (Recalculation with Updated Data)*

Gas engine CH<sub>4</sub> emission estimates increased by approximately 0.1 percent across the 1990 to 2015 time series in the current Inventory and increased by approximately 3 percent for 2016 in the current Inventory, compared to the previous Inventory. This increase in CH<sub>4</sub> emissions for 2016 is due to GHGRP submission revisions, where a higher activity factor (MMhphr/plant) was calculated from the GHGRP data.

**Table 3-88: Processing Segment Gas Engines National Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Gas Engines	137,102	169,388	228,152	234,119	234,119	250,368	255,822
<i>Previous Estimate</i>	137,102	169,101	227,671	233,626	233,626	242,451	NA

NA (Not Applicable)

## Transmission and Storage

### *Transmission Pipeline Blowdowns (Methodological Update)*

EPA developed new CH<sub>4</sub> and CO<sub>2</sub> emission factors for transmission pipeline blowdowns using GHGRP data, as detailed in the *Apr. 2019 Other Updates* memo. In response to stakeholder comments on the Public Review draft, EPA applied year-specific emission factors calculated from GHGRP data for years 2016 forward and retained historical emission factors for earlier years. As a result, compared to the previous Inventory, calculated CH<sub>4</sub> emissions from this source increased by 37 percent for year 2016 and remained constant over the rest of the time series, while CO<sub>2</sub> emission estimates increased by 33 percent for year 2016. As the updated methodology uses net emission factors, the Gas STAR reduction data from pipeline blowdowns have been removed from the calculations

(see discussion below). See the *Apr. 2019 Other Updates* memo for additional discussion and the Planned Improvements section below for considerations for future Inventories.

**Table 3-89: Transmission Pipeline Blowdowns National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Pipeline Blowdowns	177,951	183,159	184,628	183,984	183,583	250,175	184,455
<i>Previous Estimate</i>	<i>177,951</i>	<i>183,159</i>	<i>184,596</i>	<i>183,973</i>	<i>183,538</i>	<i>183,081</i>	<i>NA</i>

NA (Not Applicable)

**Table 3-90: Transmission Pipeline Blowdowns National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
Pipeline Blowdowns	5	5	5	5	5	7	5
<i>Previous Estimate</i>	<i>5</i>	<i>5</i>	<i>5</i>	<i>5</i>	<i>5</i>	<i>5</i>	<i>NA</i>

NA (Not Applicable)

### LNG Storage (Methodological Update)

For LNG storage facilities, EPA updated the Inventory methodology to use available GHGRP data paired with updated activity estimates, as detailed in the *Apr. 2019 LNG* memo. EPA developed facility-level average CH<sub>4</sub> and CO<sub>2</sub> emission factors that represent emissions from station fugitives, compressor vented and fugitive sources, and flaring using combined GHGRP data from years 2015 through 2017 and applied these emission factors across the time series. To estimate LNG storage station CH<sub>4</sub> and CO<sub>2</sub> blowdown emissions, EPA maintained the current Inventory emission factors. For activity data (storage station counts), EPA used the existing estimates for years 1990 through 2009 (although the total count of complete storage stations plus satellite stations were used, not a fraction of the satellite stations like the previous Inventory methodology) and reviewed current PHMSA data in conjunction with GHGRP data to obtain a count of active storage stations for years 2010 forward. For compressor exhaust CH<sub>4</sub> emissions, EPA updated activity factors and maintained the current Inventory emission factors. EPA developed average activity factors (i.e., MMhphr/station for each compressor driver type) using combined GHGRP data from years 2015 through 2017 and applied these activity factors across the time series. Compared to the previous Inventory, these updates resulted in an average decrease of 86 percent in CH<sub>4</sub> emission estimates across the time series and CO<sub>2</sub> emission estimates increased by an average factor of 17 across the time series.

**Table 3-91: LNG Storage Station National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
LNG storage stations	1,138	1,396	1,411	1,411	1,425	1,382	1,396
LNG storage station blowdowns	6,571	8,060	8,144	8,144	8,228	7,976	8,060
LNG storage engines	476	584	590	590	596	578	584
LNG storage turbines	26	32	33	33	33	32	32
<b>Total Emissions</b>	<b>8,212</b>	<b>10,072</b>	<b>10,177</b>	<b>10,177</b>	<b>10,282</b>	<b>9,967</b>	<b>10,072</b>
<i>Previous Estimate</i>	<i>63,258</i>	<i>73,124</i>	<i>73,124</i>	<i>73,124</i>	<i>73,124</i>	<i>73,124</i>	<i>NA</i>

NA (Not Applicable)

**Table 3-92: LNG Storage Station National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
LNG storage stations	36	45	45	45	45	44	45
LNG storage station blowdowns	+	+	+	+	+	+	+
<b>Total Emissions</b>	<b>37</b>	<b>45</b>	<b>45</b>	<b>45</b>	<b>46</b>	<b>44</b>	<b>45</b>
<i>Previous Estimate</i>	<i>2</i>	<i>2</i>	<i>2</i>	<i>2</i>	<i>2</i>	<i>2</i>	<i>NA</i>

NA (Not Applicable)

+ Does not exceed 0.5 kt.

### LNG Import/Export Terminals (Methodological Update)

For LNG terminals, EPA updated the Inventory methodology to use available GHGRP data paired with updated activity estimates, as detailed in the *Apr. 2019 LNG* memo. This methodological update also resulted in the creation of a new category for export terminals in the Inventory; previously, emissions were only estimated for import terminals. EPA used GHGRP data to develop facility-level CH<sub>4</sub> and CO<sub>2</sub> emission factors that represent emissions from station fugitives, blowdowns, compressor vented and fugitive sources, and flaring. EPA developed these facility-level emission factors for two categories of facilities: import-only terminals (import terminals) and terminals with export capability (export terminals). For import terminals, EPA calculated average CH<sub>4</sub> and CO<sub>2</sub> emission factors using combined GHGRP data from years 2015 through 2017 and applied these emission factors across the time series. For export terminals, EPA used year-specific GHGRP CH<sub>4</sub> and CO<sub>2</sub> data for 2015 through 2017 to develop emission factors and applied the year 2015 emission factors to prior time series years. For import terminals activity data, EPA used the existing Inventory import terminal counts for years 1990 through 2003 and reviewed current DOE data in conjunction with GHGRP data to obtain a count of existing import terminals for years 2004 forward. For export terminals activity data, EPA reviewed current DOE data in conjunction with GHGRP data to obtain a count of existing terminals with export capability across the time series. For compressor exhaust CH<sub>4</sub> emissions, EPA updated activity factors and maintained the current Inventory emission factors. For import terminals compressor exhaust, EPA developed average activity factors (i.e., MMhphr/station for each compressor driver type) using combined GHGRP data from years 2015 through 2017 and applied these factors across the time series. For export terminals compressor exhaust, EPA used year-specific GHGRP activity data for 2015 through 2017 to develop activity factors (i.e., MMhphr/station for each compressor driver type) and applied the year 2015 activity factors to prior time series years. These LNG terminal updates resulted in an average increase of 8 percent in CH<sub>4</sub> emissions across the time series and CO<sub>2</sub> emission estimates increased by an average factor of approximately 286 across the time series, when comparing the emissions from import and export terminals in the current Inventory to emissions from import terminals in the previous Inventory.

**Table 3-93: LNG Import/Export Terminal National CH<sub>4</sub> Emissions (Metric Tons CH<sub>4</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
LNG Import Terminals Misc. Sources <sup>a</sup>	114	284	625	625	625	568	568
LNG Import Terminal Blowdowns	2,635	6,587	14,491	14,491	14,491	13,174	13,174
LNG Import Terminal Engines	226	566	1,245	1,245	1,245	1,132	1,132
LNG Import Terminal Turbines	+	+	+	+	+	+	+
LNG Export Terminals Misc. Sources <sup>a</sup>	801	801	801	801	801	350	1,014
LNG Export Terminal Blowdowns	+	+	+	+	+	52	NO
LNG Export Terminal Engines	NO	NO	NO	NO	NO	85	NO
LNG Export Terminal Turbines	11	11	11	11	11	1	1
<b>Total Emissions</b>	<b>3,787</b>	<b>8,249</b>	<b>17,174</b>	<b>17,174</b>	<b>17,174</b>	<b>15,363</b>	<b>15,889</b>
<i>Previous Estimate<sup>b</sup></i>	<i>3,341</i>	<i>15,445</i>	<i>10,902</i>	<i>10,190</i>	<i>10,801</i>	<i>10,741</i>	<i>NA</i>

<sup>a</sup> Equipment leaks, compressor vented and leak emissions, and flares.

<sup>b</sup> Includes emissions from LNG import terminals only.

NO (Not Occurring)

NA (Not Applicable)

+ Does not exceed 0.5 MT CH<sub>4</sub>.

**Table 3-94: LNG Import/Export Terminal National CO<sub>2</sub> Emissions (kt CO<sub>2</sub>)**

Source	1990	2005	2013	2014	2015	2016	2017
LNG Import Terminals Misc. Sources <sup>a</sup>	15	37	80	80	80	73	73
LNG Import Terminal Blowdowns	+	+	1	1	1	1	1
LNG Export Terminals Misc. Sources <sup>a</sup>	+	+	+	+	+	98	278



LNG Export Terminal								
Blowdowns	NO		NO		NO	NO	NO	+
<b>Total Emissions</b>	<b>15</b>		<b>37</b>		<b>81</b>	<b>81</b>	<b>81</b>	<b>172</b>
<i>Previous Estimate</i>	+		+		+	+	+	+
								NO
								<b>352</b>
								NA

<sup>a</sup> Equipment leaks, compressor vented and leak emissions, and flares.

<sup>a</sup> Includes emissions from LNG import terminals only.

NO (Not Occurring)

NA (Not Applicable)

+ Does not exceed 0.5 kt CO<sub>2</sub>.

## Distribution

There were no methodological updates to the distribution segment, but there were recalculations due to updated data (e.g., GHGRP M&R station counts) that resulted in an average increase in calculated emissions over the time series from this segment of 0.01 MMT CO<sub>2</sub> Eq. CH<sub>4</sub> (or 0.1 percent) and less than 0.01 MMT CO<sub>2</sub> (or 0.1 percent).

## Gas STAR Data Revisions

EPA updated its calculation of production and transmission segment Gas STAR reductions to take into account new methods using net emission factors for certain sources. As in previous inventories, the “other” reductions scaling factor for production is calculated as one minus the sum of emissions from sources with net approaches, divided by the sum of all production segment emissions. The calculation was updated this year to remove reductions associated with gathering pipeline blowdowns, as net emission factors are now used to calculate emission for that source. In addition, the line item for gathering pipeline leak Gas STAR reductions was removed. Similarly, reductions associated with transmission pipelines blowdowns were removed from the transmission segment. As a result of the update, Gas STAR reductions averaged 8.0 MMT CO<sub>2</sub> Eq. over the time series, which is an average decrease across the time series of 19 percent (or 1.7 MMT CO<sub>2</sub> Eq.).

## N<sub>2</sub>O Emissions

EPA newly calculated N<sub>2</sub>O emissions in the current Inventory, as discussed in the *Apr. 2018 Other Updates* memo. Prior Inventories did not calculate N<sub>2</sub>O emissions from natural gas systems. For each flaring emission source calculation methodology which uses GHGRP data, the existing source-specific methodology was applied to calculate N<sub>2</sub>O emission factors. This update was applied for sources in the exploration, production, processing, and transmission and storage segments.

**Table 3-95: N<sub>2</sub>O National Emissions (Metric Tons N<sub>2</sub>O)**

Activity	1990	2005	2013	2014	2015	2016	2017
<b>Exploration</b>	<b>1.5</b>	<b>4.7</b>	<b>4.0</b>	<b>2.9</b>	<b>10.8</b>	<b>0.4</b>	<b>1.0</b>
Non-completion well testing - flared	0.8	1.4	1.7	1.7	1.7	NO	+
HF Completions with Flaring	0.8	3.3	2.2	1.1	8.1	0.4	0.9
Non-HF Completions with Flaring	+	+	+	+	0.9	+	+
<b>Production</b>	<b>0.5</b>	<b>3.0</b>	<b>7.8</b>	<b>6.7</b>	<b>9.3</b>	<b>3.4</b>	<b>3.1</b>
HF Workovers with Flaring	0.1	0.4	0.2	0.2	2.2	0.1	0.7
Non-HF Workovers with Flaring	NO	+	1.7	+	+	NO	NO
Misc. Onshore Production Flaring	NO	2.1	4.2	4.8	5.3	2.2	1.8
Tanks with Flares	0.4	0.6	1.6	1.7	1.7	1.1	0.6
<b>Processing</b>	<b>NO</b>	<b>11.2</b>	<b>18.9</b>	<b>19.4</b>	<b>19.4</b>	<b>12.8</b>	<b>10.2</b>
Flares	NO	11.2	18.9	19.4	19.4	12.8	10.2
<b>Transmission and Storage</b>	<b>0.9</b>	<b>1.0</b>	<b>1.1</b>	<b>1.2</b>	<b>1.2</b>	<b>1.3</b>	<b>1.5</b>
Transmission Flaring	0.1	0.1	0.1	0.1	0.1	+	0.1
Storage Flaring	+	+	+	+	+	0.1	+
LNG Storage Flaring	0.7	0.8	0.8	0.8	0.8	0.8	0.8

LNG Import Terminals Flaring	+	0.1	0.2	0.2	0.2	0.1	0.1
LNG Export Terminals Flaring	NO	NO	NO	NO	NO	0.2	0.5
<b>Distribution</b>	<b>NO</b>	<b>NO</b>	<b>NO</b>	<b>NO</b>	<b>NO</b>	<b>NO</b>	<b>NO</b>
<b>Total Emissions</b>	<b>3.0</b>	<b>20.0</b>	<b>31.8</b>	<b>30.1</b>	<b>40.6</b>	<b>17.8</b>	<b>15.9</b>

NO (Not Occurring)

+ Does not exceed 0.05 MT N<sub>2</sub>O.

## Planned Improvements

EPA seeks stakeholder feedback on the improvements noted below for future Inventories.

### Gathering and Boosting Stations

In the *Oct. 2018 G&B* memo, EPA presented approaches that rely on GHGRP data to estimate G&B station emissions. Stakeholder feedback received in response to the *Oct. 2018 G&B* memo supported maintaining the current Inventory approach. As such, EPA maintained the current Inventory approach to estimate G&B station emissions, and did not use a methodology that relies on GHGRP data. EPA will continue to review GHGRP data and other research that becomes available to estimate G&B station emissions. EPA also requests specific feedback on options to estimate G&B station flaring (CO<sub>2</sub> and N<sub>2</sub>O) emissions and AGR (CO<sub>2</sub>) emissions for future Inventories. The current Inventory approach does not account for the CO<sub>2</sub> emissions from flaring and AGR units. EPA plans to review available data from upcoming studies and additional years of data reported to GHGRP to improve estimated emissions from these sources. The GHGRP emissions from flaring in gathering and boosting total 3,894 kt CO<sub>2</sub> and 0.01 kt N<sub>2</sub>O reported for year 2017 (2,143 kt CO<sub>2</sub> from miscellaneous flaring, 686 kt CO<sub>2</sub> from flaring from dehydrators, 579 kt CO<sub>2</sub> from flaring from tanks, and 486 kt CO<sub>2</sub> from AGR units).

### Transmission Pipeline Blowdowns

For the final 2019 Inventory estimate, in response to stakeholder feedback, EPA calculated year-specific emission factors for transmission pipeline blowdowns using data from the first two years of GHGRP reporting, 2016 and 2017, and applied historical emission factors to all previous time series years. EPA is considering other approaches for future Inventories, as additional years of GHGRP data become available. EPA requests feedback on whether an updated methodology should be applied for earlier time series years (e.g., retain current emission factors for 1990 to 1992, then use linear interpolation to calculate emission factors for years 1993 through 2015; or develop an average factor from 2016 through 2018 GHGRP data to apply for 1990 through 2015).

### Well-Related Activity Data

As described in the Recalculations Discussion, EPA has updated the emission factors for several well-related emission sources, including testing, completions, and workovers. EPA will continue to assess available data, including data from the GHGRP and stakeholder feedback on considerations, to improve activity estimates for sources that rely on well-related activity data. For example, EPA will seek information on other data sets that might inform estimates of non-hydraulically fractured gas well completions and workovers.

### Offshore Platforms

EPA is considering updates to the offshore platform emissions calculation methodology, as discussed in the *2018 Other Updates Memo*. The current emission factors were based on data from the 2011 DOI/Bureau of Ocean Energy Management's (BOEM) dataset, and 2014 BOEM data are available. A different source for platform counts is also being considered.

### Upcoming Data, and Additional Data that Could Inform the Inventory

EPA will assess new data received by the Methane Challenge Program on an ongoing basis, which may be used to confirm or improve existing estimates and assumptions.

EPA continues to track studies that contain data that may be used to update the Inventory. Key studies in progress include: DOE-funded work on vintage and new plastic pipelines (distribution segment), industrial meters (distribution segment), and sources within the gathering and storage segments<sup>92</sup>; an API field study on pneumatic controllers; a Pipeline Research Council International (PRCI) project in which researchers are gathering and analyzing subpart W data on transmission compressor stations and underground storage facilities; and other studies by research groups that will examine gathering and boosting emissions and offshore platform emissions. EPA will also continue to assess studies that include and compare both top-down and bottom-up estimates, and which could lead to improved understanding of unassigned high emitters (e.g., identification of emission sources and information on frequency of high emitters) as recommended in stakeholder comments.

EPA also continues to seek new data that could be used to assess or update the estimates in the Inventory. For example, stakeholder comments have highlighted areas where additional data that could inform the Inventory are currently limited or unavailable:

- Tank malfunction and control efficiency data.
- Consider updating engine emission factors, including using subpart W data to the extent possible, and considering whether and how to represent differences between rich- and lean-burn engines.
- Activity data and emissions data for production facilities that do not report to GHGRP.
- Natural gas leaks at point of use estimates.
- Anomalous leak events, such as a 2018 well blowout in Ohio.

EPA will continue to seek available data on these and other sources as part of the process to update the Inventory.

## 3.8 Abandoned Oil and Gas Wells (CRF Source Categories 1B2a and 1B2b)

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The term "abandoned wells" encompasses various types of wells:

- Wells with no recent production, and not plugged. Common terms (such as those used in state databases) might include: inactive, temporarily abandoned, shut-in, dormant, and idle.
- Wells with no recent production and no responsible operator. Common terms might include: orphaned, deserted, long-term idle, and abandoned.
- Wells that have been plugged to prevent migration of gas or fluids.

The U.S. population of abandoned wells is around 3.2 million (with around 2.6 million abandoned oil wells and 0.6 million abandoned gas wells). Abandoned wells emit both CH<sub>4</sub> and CO<sub>2</sub>. Wells that are plugged have much lower average emissions than wells that are unplugged (less than 1 kg CH<sub>4</sub> per well per year, versus over 100 kg CH<sub>4</sub> per well per year). Around a third of the abandoned well population in the United States is plugged. This fraction has increased over the time series (from around 19 percent in 1990) as more wells fall under regulations and programs requiring or promoting plugging of abandoned wells.

*Abandoned oil wells.* Abandoned oil wells emitted 224 kt CH<sub>4</sub> and 5 kt CO<sub>2</sub> in 2017. Emissions of both gases decreased by 1 percent from 1990, while the total population of abandoned oil wells increased 26 percent. Emissions of both gases decreased by 4 percent between 2016 and 2017 as a result of well plugging activities.

*Abandoned gas wells.* Abandoned gas wells emitted 53 kt CH<sub>4</sub> and 2 kt CO<sub>2</sub> in 2017. Emissions of both gases increased by 47 percent from 1990, as the total population of abandoned gas wells increased 75 percent. Emissions of both gases decreased by 4 percent between 2016 and 2017 as a result of well plugging activities.

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<sup>92</sup> See <<https://www.energy.gov/under-secretary-science-and-energy/articles/doe-announces-13-million-quantify-and-mitigate-methane>>.

**Table 3-96: CH<sub>4</sub> Emissions from Abandoned Oil and Gas Wells (MMT CO<sub>2</sub> Eq.)**

Activity	1990	2005	2013	2014	2015	2016	2017
Abandoned Oil Wells	5.7	5.9	5.8	5.8	5.8	5.8	5.6
Abandoned Gas Wells	0.9	1.1	1.2	1.3	1.3	1.4	1.3
<b>Total</b>	<b>6.6</b>	<b>6.9</b>	<b>7.0</b>	<b>7.1</b>	<b>7.1</b>	<b>7.2</b>	<b>6.9</b>

Note: Totals may not sum due to independent rounding.

**Table 3-97: CH<sub>4</sub> Emissions from Abandoned Oil and Gas Wells (kt)**

Activity	1990	2005	2013	2014	2015	2016	2017
Abandoned Oil Wells	226	235	232	232	232	234	224
Abandoned Gas Wells	36	42	50	52	53	55	53
<b>Total</b>	<b>262</b>	<b>277</b>	<b>282</b>	<b>283</b>	<b>285</b>	<b>289</b>	<b>277</b>

Note: Totals may not sum due to independent rounding.

**Table 3-98: CO<sub>2</sub> Emissions from Abandoned Oil and Gas Wells (MMT CO<sub>2</sub>)**

Activity	1990	2005	2013	2014	2015	2016	2017
Abandoned Oil Wells	+	+	+	+	+	+	+
Abandoned Gas Wells	+	+	+	+	+	+	+
<b>Total</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>	<b>+</b>

+ Does not exceed 0.05 MMT CO<sub>2</sub>.

**Table 3-99: CO<sub>2</sub> Emissions from Abandoned Oil and Gas Wells (kt)**

Activity	1990	2005	2013	2014	2015	2016	2017
Abandoned Oil Wells	5	5	5	5	5	5	5
Abandoned Gas Wells	2	2	2	2	2	2	2
<b>Total</b>	<b>6</b>	<b>7</b>	<b>7</b>	<b>7</b>	<b>7</b>	<b>7</b>	<b>7</b>

Note: Totals may not sum due to independent rounding.

## Methodology

EPA developed abandoned well CH<sub>4</sub> emission factors using data from Kang et al. (2016) and Townsend-Small et al. (2016). Plugged and unplugged abandoned well CH<sub>4</sub> emission factors were developed at the national-level (emission data from Townsend-Small et al.) and for the Appalachia region (using emission data from measurements in Pennsylvania and Ohio conducted by Kang et al. and Townsend-Small et al., respectively). The Appalachia region emissions factors were applied to abandoned wells in states in the Appalachian basin region, and the national-level emission factors were applied to all other abandoned wells.

EPA developed abandoned well CO<sub>2</sub> emission factors using the CH<sub>4</sub> emission factors and an assumed ratio of CO<sub>2</sub>-to-CH<sub>4</sub> gas content, similar to the approach used to calculate CO<sub>2</sub> emissions for many sources in Petroleum Systems and Natural Gas Systems. For abandoned oil wells, EPA used the Petroleum Systems default production segment associated gas ratio of 0.020 MT CO<sub>2</sub>/MT CH<sub>4</sub>, which was derived through API TankCalc modeling runs. For abandoned gas wells, EPA used the Natural Gas Systems default production segment CH<sub>4</sub> and CO<sub>2</sub> gas content values (GRI/EPA 1996, GTI 2001) to develop a ratio of 0.044 MT CO<sub>2</sub>/MT CH<sub>4</sub>.

The total population of abandoned wells over the time series was estimated using historical data and DrillingInfo data. For the most recent year of the Inventory time series (year 2017), the prior year total counts are used as surrogate data, as the DrillingInfo query approach for the most recent year would likely overestimate abandoned well counts, because many wells might be spud and not reporting production—not because they are dry/abandoned, but due to the time required for completion. The abandoned well population was then split into plugged and unplugged wells by assuming that all abandoned wells were unplugged in 1950, using year-specific Drilling info

data to calculate the fraction of abandoned wells plugged in 2016 and 2017 (31 percent and 34 percent, respectively), and applying linear interpolation between the 1950 value and 2016 value to calculate plugged fraction for intermediate years. See the memorandum *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2016: Abandoned Wells in Natural Gas and Petroleum Systems (2018 Abandoned Wells Memo)* for details.<sup>93</sup>

### Abandoned Oil Wells

**Table 3-100: Abandoned Oil Wells Activity Data, CH<sub>4</sub> and CO<sub>2</sub> Emissions (Metric Tons)**

Source	1990	2005	2013	2014	2015	2016	2017
Plugged abandoned oil wells	386,145	616,421	741,135	758,150	778,912	800,330	871,069
Unplugged abandoned oil wells	1,682,514	1,785,249	1,779,764	1,780,330	1,788,961	1,798,176	1,727,437
<b>Total Abandoned Oil Wells</b>	<b>2,068,659</b>	<b>2,401,670</b>	<b>2,520,900</b>	<b>2,538,480</b>	<b>2,567,873</b>	<b>2,598,506</b>	<b>2,598,506</b>
Abandoned oil wells in Appalachia	26%	24%	23%	23%	23%	23%	23%
Abandoned oil wells outside of Appalachia	74%	76%	77%	77%	77%	77%	77%
CH <sub>4</sub> from plugged abandoned oil wells (MT)	317	476	552	563	575	591	643
CH <sub>4</sub> from unplugged abandoned oil wells (MT)	225,944	234,654	231,228	230,964	231,744	232,937	223,774
<b>Total CH<sub>4</sub> from Abandoned oil wells (MT)</b>	<b>226,261</b>	<b>235,129</b>	<b>231,781</b>	<b>231,526</b>	<b>232,319</b>	<b>233,529</b>	<b>224,417</b>
<b>Total CO<sub>2</sub> from Abandoned oil wells (MT)</b>	<b>4,591</b>	<b>4,771</b>	<b>4,703</b>	<b>4,698</b>	<b>4,714</b>	<b>4,739</b>	<b>4,554</b>

### Abandoned Gas Wells

**Table 3-101: Abandoned Gas Wells Activity Data, CH<sub>4</sub> and CO<sub>2</sub> Emissions (Metric Tons)**

Source	1990	2005	2013	2014	2015	2016	2017
Plugged abandoned gas wells	59,627	103,856	145,970	154,171	161,814	172,296	187,525
Unplugged abandoned gas wells	259,807	300,784	350,532	362,033	371,645	387,115	371,886
<b>Total Abandoned Gas Wells</b>	<b>319,434</b>	<b>404,640</b>	<b>496,501</b>	<b>516,203</b>	<b>533,459</b>	<b>559,411</b>	<b>559,411</b>
Abandoned gas wells in Appalachia	28%	29%	30%	30%	30%	30%	30%
Abandoned gas wells outside of Appalachia	72%	71%	70%	70%	70%	70%	70%
CH <sub>4</sub> from plugged abandoned gas wells (MT)	53	96	139	147	154	164	179
CH <sub>4</sub> from unplugged abandoned gas wells (MT)	35,899	42,258	49,681	51,367	52,788	54,985	52,822
<b>Total CH<sub>4</sub> from abandoned gas wells (MT)</b>	<b>35,952</b>	<b>42,354</b>	<b>49,819</b>	<b>51,513</b>	<b>52,942</b>	<b>55,150</b>	<b>53,001</b>
<b>Total CO<sub>2</sub> from abandoned gas wells (MT)</b>	<b>1,576</b>	<b>1,856</b>	<b>2,183</b>	<b>2,258</b>	<b>2,320</b>	<b>2,417</b>	<b>2,323</b>

## Uncertainty and Time-Series Consistency

To characterize uncertainty surrounding estimates of abandoned well emissions, EPA conducted a quantitative uncertainty analysis using the IPCC Approach 2 methodology (Monte Carlo simulation technique). See the *2018 Abandoned Wells Memo* for details of the uncertainty analysis methods. EPA used Microsoft Excel's @RISK add-in

<sup>93</sup> See <<https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems>>

tool to estimate the 95 percent confidence bound around total methane emissions from abandoned oil and gas wells in year 2017, then applied the calculated bounds to both CH<sub>4</sub> and CO<sub>2</sub> emissions estimates for each population. The @RISK add-in provides for the specification of probability density functions (PDFs) for key variables within a computational structure that mirrors the calculation of the inventory estimate. EPA used measurement data from the Kang et al. (2016) and Townsend-Small et al. (2016) studies to characterize the CH<sub>4</sub> emission factor PDFs. For activity data inputs (e.g., total count of abandoned wells, split between plugged and unplugged), EPA assigned default uncertainty bounds of +/- 10 percent based on expert judgment.

The IPCC guidance notes that in using this method, "some uncertainties that are not addressed by statistical means may exist, including those arising from omissions or double counting, or other conceptual errors, or from incomplete understanding of the processes that may lead to inaccuracies in estimates developed from models." As a result, the understanding of the uncertainty of emission estimates for this category evolves and improves as the underlying methodologies and datasets improve. The uncertainty bounds reported below only reflect those uncertainties that EPA has been able to quantify and do not incorporate considerations such as modeling uncertainty, data representativeness, measurement errors, misreporting or misclassification.

The results presented below in Table 3-102 provide the 95 percent confidence bound within which actual emissions from abandoned oil and gas wells are likely to fall for the year 2017, using the recommended IPCC methodology. Abandoned oil well CH<sub>4</sub> emissions in 2017 were estimated to be between 1.0 and 17.9 MMT CO<sub>2</sub> Eq., while abandoned gas well CH<sub>4</sub> emissions were estimated to be between 0.2 and 4.2 MMT CO<sub>2</sub> Eq. at a 95 percent confidence level. Uncertainty bounds for other years of the time series have not been calculated, but uncertainty is expected to vary over the time series.

**Table 3-102: Approach 2 Quantitative Uncertainty Estimates for CH<sub>4</sub> and CO<sub>2</sub> Emissions from Petroleum and Natural Gas Systems (MMT CO<sub>2</sub> Eq. and Percent)**

Source	Gas	2017 Emission Estimate (MMT CO <sub>2</sub> Eq.) <sup>b</sup>	Uncertainty Range Relative to Emission Estimate <sup>a</sup>			
			(MMT CO <sub>2</sub> Eq.)		(%)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Abandoned Oil Wells	CH <sub>4</sub>	5.6	1.0	17.9	-83%	+219%
Abandoned Gas Wells	CH <sub>4</sub>	1.3	0.2	4.2	-83%	+219%
Abandoned Oil Wells	CO <sub>2</sub>	0.005	0.001	0.015	-83%	+219%
Abandoned Gas Wells	CO <sub>2</sub>	0.002	0.0004	0.007	-83%	+219%

<sup>a</sup> Range of emission estimates estimated by applying the 95 percent confidence intervals obtained from the Monte Carlo Simulation analysis conducted for total abandoned oil and gas well CH<sub>4</sub> emissions in year 2017.

<sup>b</sup> All reported values are rounded after calculation. As a result, lower and upper bounds may not be duplicable from other rounded values as shown in table.

To calculate a time series of emissions for abandoned wells, EPA developed annual activity data for 1990 through 2017 by summing an estimate of total abandoned wells not included in recent databases, to an annual estimate of abandoned wells in the DrillingInfo data set (with year 2016 estimates used as surrogates for year 2017 data). As discussed above, the abandoned well population was split into plugged and unplugged wells by assuming that all abandoned wells were unplugged in 1950, using year-specific Drilling info data to calculate the fraction of abandoned wells plugged in 2016 and 2017 (31 percent and 34 percent, respectively), and applying linear interpolation between the 1950 value and 2016 value to calculate plugged fraction for intermediate years. The same emission factors were applied to the corresponding categories for each year of the time series.

## QA/QC and Verification Discussion

The emission estimates in the Inventory are continually being reviewed and assessed to determine whether emission factors and activity factors accurately reflect current industry practices. A QA/QC analysis was performed for data gathering and input, documentation, and calculation. QA/QC checks are consistently conducted to minimize human error in the model calculations. EPA performs a thorough review of information associated with new studies to assess whether the assumptions in the Inventory are consistent with industry practices and whether new data is available that could be considered for updates to the estimates. As in previous years, EPA conducted early

engagement and communication with stakeholders on updates prior to public review. EPA held a stakeholder workshop on greenhouse gas data for oil and gas in October of 2018, and webinars in June of 2018 and February of 2019.

## Recalculations Discussion

The counts of national abandoned wells were recalculated across the time series to use the latest DrillingInfo data, which resulted in minor changes to the total abandoned well population and the allocation between petroleum and natural gas systems. The minor changes resulted from changes to the year-specific data for 1990 to 2016 as processed from DrillingInfo, which led EPA to recalculate the estimate of historical wells not included in the DrillingInfo data set (which decreased from 1,149,618 to 1,108,648 historical wells not included in DrillingInfo). Compared with the previous Inventory, counts of abandoned oil and gas wells are on average 1.0 percent and 0.2 percent, respectively, higher over 1990 to 2016. The impact was largest in recent years, with abandoned oil and gas well counts recalculated to be 1.5 percent and 1.6 percent, respectively, higher for 2016 comparing the previous Inventory values to the current Inventory values; this change is also due to the use of year-specific data for year 2016 (as the previous Inventory used year 2015 estimates as surrogate for year 2016 per the established methodology described above).

## Planned Improvements

The abandoned wells source was added to the previous (i.e., 1990 through 2016) Inventory in 2018. Through EPA's stakeholder process on oil and gas in the development of the 2018 Inventory, EPA received stakeholder feedback on the abandoned wells update to the Inventory. EPA will continue to assess new data and stakeholder feedback on considerations (such as the disaggregation of the well population into Appalachia and other regions, and emission factor data from regions not included in the measurement studies on which current emission factors are based) to improve the abandoned well count estimates and emission factors.

## 3.9 Energy Sources of Precursor Greenhouse Gas Emissions

In addition to the main greenhouse gases addressed above, energy-related activities are also sources of precursor gases. The reporting requirements of the UNFCCC<sup>94</sup> request that information be provided on precursor greenhouse gases, which include carbon monoxide (CO), nitrogen oxides (NO<sub>x</sub>), non-CH<sub>4</sub> volatile organic compounds (NMVOCs), and sulfur dioxide (SO<sub>2</sub>). These gases are not direct greenhouse gases, but indirectly affect terrestrial radiation absorption by influencing the formation and destruction of tropospheric and stratospheric ozone, or, in the case of SO<sub>2</sub>, by affecting the absorptive characteristics of the atmosphere. Additionally, some of these gases may react with other chemical compounds in the atmosphere to form compounds that are greenhouse gases. Total emissions of NO<sub>x</sub>, CO, and NMVOCs from energy-related activities from 1990 to 2017 are reported in Table 3-103. Sulfur dioxide emissions are presented in Section 2.3 of the Trends chapter and Annex 6.3.

**Table 3-103: NO<sub>x</sub>, CO, and NMVOC Emissions from Energy-Related Activities (kt)**

Gas/Activity	1990	2005	2013	2014	2015	2016	2017
<b>NO<sub>x</sub></b>	<b>21,106</b>	<b>16,602</b>	<b>10,740</b>	<b>10,204</b>	<b>9,529</b>	<b>9,042</b>	<b>8,559</b>
Mobile Fossil Fuel Combustion	10,862	10,295	6,523	6,138	5,740	5,413	5,051
Stationary Fossil Fuel Combustion	10,023	5,858	3,487	3,319	3,042	2,882	2,761
Oil and Gas Activities	139	321	641	650	650	650	650
Waste Combustion	82	128	89	97	97	97	97
<i>International Bunker Fuels<sup>a</sup></i>	<i>1,956</i>	<i>1,704</i>	<i>1,139</i>	<i>1,139</i>	<i>1,226</i>	<i>1,322</i>	<i>1,323</i>

<sup>94</sup> See <<http://unfccc.int/resource/docs/2013/cop19/eng/10a03.pdf>>.

<b>CO</b>	<b>125,640</b>	<b>64,985</b>	<b>41,519</b>	<b>40,234</b>	<b>39,258</b>	<b>36,885</b>	<b>35,211</b>
Mobile Fossil Fuel Combustion	119,360	58,615	35,525	34,135	33,159	30,786	29,112
Stationary Fossil Fuel Combustion	5,000	4,648	3,847	3,686	3,686	3,686	3,686
Waste Combustion	978	1,403	1,518	1,776	1,776	1,776	1,776
Oil and Gas Activities	302	318	628	637	637	637	637
<i>International Bunker Fuels<sup>a</sup></i>	<i>103</i>	<i>133</i>	<i>129</i>	<i>135</i>	<i>141</i>	<i>146</i>	<i>152</i>
<b>NMVOCs</b>	<b>12,620</b>	<b>7,191</b>	<b>7,419</b>	<b>7,247</b>	<b>7,082</b>	<b>6,835</b>	<b>6,629</b>
Mobile Fossil Fuel Combustion	10,932	5,724	4,023	3,754	3,589	3,342	3,137
Oil and Gas Activities	554	510	2,741	2,853	2,853	2,853	2,853
Stationary Fossil Fuel Combustion	912	716	532	497	497	497	497
Waste Combustion	222	241	122	143	143	143	143
<i>International Bunker Fuels<sup>a</sup></i>	<i>57</i>	<i>54</i>	<i>41</i>	<i>42</i>	<i>47</i>	<i>50</i>	<i>51</i>

<sup>a</sup> These values are presented for informational purposes only and are not included in totals.

Note: Totals may not sum due to independent rounding.

## Methodology

Emission estimates for 1990 through 2017 were obtained from data published on the National Emission Inventory (NEI) Air Pollutant Emission Trends web site (EPA 2018), and disaggregated based on EPA (2003). Emissions were calculated either for individual categories or for many categories combined, using basic activity data (e.g., the amount of raw material processed) as an indicator of emissions. National activity data were collected for individual applications from various agencies.

Activity data were used in conjunction with emission factors, which together relate the quantity of emissions to the activity. Emission factors are generally available from the EPA's *Compilation of Air Pollutant Emission Factors, AP-42* (EPA 1997). The EPA currently derives the overall emission control efficiency of a source category from a variety of information sources, including published reports, the 1985 National Acid Precipitation and Assessment Program emissions inventory, and other EPA databases.

## Uncertainty and Time-Series Consistency

Uncertainties in these estimates are partly due to the accuracy of the emission factors used and accurate estimates of activity data. A quantitative uncertainty analysis was not performed.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2017. Details on the emission trends through time are described in more detail in the Methodology section, above.

## 3.10 International Bunker Fuels (CRF Source Category 1: Memo Items)

Emissions resulting from the combustion of fuels used for international transport activities, termed international bunker fuels under the UNFCCC, are not included in national emission totals, but are reported separately based upon location of fuel sales. The decision to report emissions from international bunker fuels separately, instead of allocating them to a particular country, was made by the Intergovernmental Negotiating Committee in establishing the Framework Convention on Climate Change.<sup>95</sup> These decisions are reflected in the IPCC methodological guidance, including IPCC (2006), in which countries are requested to report emissions from ships or aircraft that

<sup>95</sup> See report of the Intergovernmental Negotiating Committee for a Framework Convention on Climate Change on the work of its ninth session, held at Geneva from 7 to 18 February 1994 (A/AC.237/55, annex I, para. 1c).



depart from their ports with fuel purchased within national boundaries and are engaged in international transport separately from national totals (IPCC 2006).<sup>96</sup>

Two transport modes are addressed under the IPCC definition of international bunker fuels: aviation and marine.<sup>97</sup> Greenhouse gases emitted from the combustion of international bunker fuels, like other fossil fuels, include CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O for marine transport modes, and CO<sub>2</sub> and N<sub>2</sub>O for aviation transport modes. Emissions from ground transport activities—by road vehicles and trains—even when crossing international borders are allocated to the country where the fuel was loaded into the vehicle and, therefore, are not counted as bunker fuel emissions.

The 2006 IPCC Guidelines distinguish between different modes of air traffic. Civil aviation comprises aircraft used for the commercial transport of passengers and freight, military aviation comprises aircraft under the control of national armed forces, and general aviation applies to recreational and small corporate aircraft. The 2006 IPCC Guidelines further define international bunker fuel use from civil aviation as the fuel combusted for civil (e.g., commercial) aviation purposes by aircraft arriving or departing on international flight segments. However, as mentioned above, and in keeping with the 2006 IPCC Guidelines, only the fuel purchased in the United States and used by aircraft taking-off (i.e., departing) from the United States are reported here. The standard fuel used for civil aviation is kerosene-type jet fuel, while the typical fuel used for general aviation is aviation gasoline.<sup>98</sup>

Emissions of CO<sub>2</sub> from aircraft are essentially a function of fuel use. Nitrous oxide emissions also depend upon engine characteristics, flight conditions, and flight phase (i.e., take-off, climb, cruise, decent, and landing). Recent data suggest that little or no CH<sub>4</sub> is emitted by modern engines (Anderson et al. 2011), and as a result, CH<sub>4</sub> emissions from this category are considered zero. In jet engines, N<sub>2</sub>O is primarily produced by the oxidation of atmospheric nitrogen, and the majority of emissions occur during the cruise phase. International marine bunkers comprise emissions from fuels burned by ocean-going ships of all flags that are engaged in international transport. Ocean-going ships are generally classified as cargo and passenger carrying, military (i.e., U.S. Navy), fishing, and miscellaneous support ships (e.g., tugboats). For the purpose of estimating greenhouse gas emissions, international bunker fuels are solely related to cargo and passenger carrying vessels, which is the largest of the four categories, and military vessels. Two main types of fuels are used on sea-going vessels: distillate diesel fuel and residual fuel oil. Carbon dioxide is the primary greenhouse gas emitted from marine shipping.

Overall, aggregate greenhouse gas emissions in 2017 from the combustion of international bunker fuels from both aviation and marine activities were 121.2 MMT CO<sub>2</sub> Eq., or 16.0 percent above emissions in 1990 (see Table 3-104 and Table 3-105). Emissions from international flights and international shipping voyages departing from the United States have increased by 104.3 percent and decreased by 35.3 percent, respectively, since 1990. The majority of these emissions were in the form of CO<sub>2</sub>; however, small amounts of CH<sub>4</sub> (from marine transport modes) and N<sub>2</sub>O were also emitted.

**Table 3-104: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from International Bunker Fuels (MMT CO<sub>2</sub> Eq.)**

Gas/Mode	1990	2005	2013	2014	2015	2016	2017
<b>CO<sub>2</sub></b>	<b>103.5</b>	<b>113.1</b>	<b>99.8</b>	<b>103.4</b>	<b>110.9</b>	<b>116.6</b>	<b>120.1</b>
Aviation	38.0	60.1	65.7	69.6	71.9	74.1	77.7
<i>Commercial</i>	30.0	55.6	62.8	66.3	68.6	70.8	74.5
<i>Military</i>	8.1	4.5	2.9	3.3	3.3	3.3	3.2
Marine	65.4	53.0	34.1	33.8	38.9	42.5	42.4
<b>CH<sub>4</sub></b>	<b>0.2</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>	<b>0.1</b>
Aviation <sup>a</sup>	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Marine	0.2	0.1	0.1	0.1	0.1	0.1	0.1
<b>N<sub>2</sub>O</b>	<b>0.9</b>	<b>1.0</b>	<b>0.9</b>	<b>0.9</b>	<b>0.9</b>	<b>1.0</b>	<b>1.0</b>
Aviation	0.4	0.6	0.6	0.7	0.7	0.7	0.7

<sup>96</sup> Note that the definition of international bunker fuels used by the UNFCCC differs from that used by the International Civil Aviation Organization.

<sup>97</sup> Most emission related international aviation and marine regulations are under the rubric of the International Civil Aviation Organization (ICAO) or the International Maritime Organization (IMO), which develop international codes, recommendations, and conventions, such as the International Convention of the Prevention of Pollution from Ships (MARPOL).

<sup>98</sup> Naphtha-type jet fuel was used in the past by the military in turbojet and turboprop aircraft engines.

Marine	0.5	0.4	0.2	0.2	0.3	0.3	0.3
<b>Total</b>	<b>104.5</b>	<b>114.2</b>	<b>100.7</b>	<b>104.4</b>	<b>111.9</b>	<b>117.7</b>	<b>121.2</b>

<sup>a</sup> CH<sub>4</sub> emissions from aviation are estimated to be zero.

Notes: Totals may not sum due to independent rounding. Includes aircraft cruise altitude emissions.

**Table 3-105: CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O Emissions from International Bunker Fuels (kt)**

Gas/Mode	1990	2005	2013	2014	2015	2016	2017
<b>CO<sub>2</sub></b>	<b>103,463</b>	<b>113,139</b>	<b>99,763</b>	<b>103,400</b>	<b>110,887</b>	<b>116,594</b>	<b>120,107</b>
Aviation	38,034	60,125	65,664	69,609	71,942	74,059	77,696
Marine	65,429	53,014	34,099	33,791	38,946	42,535	42,412
<b>CH<sub>4</sub></b>	<b>7</b>	<b>5</b>	<b>3</b>	<b>3</b>	<b>3</b>	<b>4</b>	<b>4</b>
Aviation <sup>a</sup>	0	0	0	0	0	0	0
Marine	7	5	3	3	3	4	4
<b>N<sub>2</sub>O</b>	<b>3</b>	<b>3</b>	<b>3</b>	<b>3</b>	<b>3</b>	<b>3</b>	<b>3</b>
Aviation	1	2	2	2	2	2	3
Marine	2	1	1	1	1	1	1

<sup>a</sup> CH<sub>4</sub> emissions from aviation are estimated to be zero.

Notes: Totals may not sum due to independent rounding. Includes aircraft cruise altitude emissions.

## Methodology

Emissions of CO<sub>2</sub> were estimated by applying C content and fraction oxidized factors to fuel consumption activity data. This approach is analogous to that described under Section 3.1 – CO<sub>2</sub> from Fossil Fuel Combustion. Carbon content and fraction oxidized factors for jet fuel, distillate fuel oil, and residual fuel oil were taken directly from EIA and are presented in Annex 2.1, Annex 2.2, and Annex 3.8 of this Inventory. Density conversions were taken from Chevron (2000), ASTM (1989), and USAF (1998). Heat content for distillate fuel oil and residual fuel oil were taken from EIA (2019) and USAF (1998), and heat content for jet fuel was taken from EIA (2019). A complete description of the methodology and a listing of the various factors employed can be found in Annex 2.1. See Annex 3.8 for a specific discussion on the methodology used for estimating emissions from international bunker fuel use by the U.S. military.

Emission estimates for CH<sub>4</sub> and N<sub>2</sub>O were calculated by multiplying emission factors by measures of fuel consumption by fuel type and mode. Emission factors used in the calculations of CH<sub>4</sub> and N<sub>2</sub>O emissions were obtained from the *2006 IPCC Guidelines* (IPCC 2006). For aircraft emissions, the following value, in units of grams of pollutant per kilogram of fuel consumed (g/kg), was employed: 0.1 for N<sub>2</sub>O (IPCC 2006). For marine vessels consuming either distillate diesel or residual fuel oil the following values (g/MJ), were employed: 0.32 for CH<sub>4</sub> and 0.08 for N<sub>2</sub>O. Activity data for aviation included solely jet fuel consumption statistics, while the marine mode included both distillate diesel and residual fuel oil.

Activity data on domestic and international aircraft fuel consumption were developed by the U.S. Federal Aviation Administration (FAA) using radar-informed data from the FAA Enhanced Traffic Management System (ETMS) for 1990 and 2000 through 2017 as modeled with the Aviation Environmental Design Tool (AEDT). This bottom-up approach is built from modeling dynamic aircraft performance for each flight occurring within an individual calendar year. The analysis incorporates data on the aircraft type, date, flight identifier, departure time, arrival time, departure airport, arrival airport, ground delay at each airport, and real-world flight trajectories. To generate results for a given flight within AEDT, the radar-informed aircraft data is correlated with engine and aircraft performance data to calculate fuel burn and exhaust emissions. Information on exhaust emissions for in-production aircraft engines comes from the International Civil Aviation Organization (ICAO) Aircraft Engine Emissions Databank (EDB). This bottom-up approach is in accordance with the Tier 3B method from the *2006 IPCC Guidelines* (IPCC 2006).

International aviation CO<sub>2</sub> estimates for 1990 and 2000 through 2017 were obtained from FAA's AEDT model (FAA 2019). The radar-informed method that was used to estimate CO<sub>2</sub> emissions for commercial aircraft for 1990 and 2000 through 2017 was not possible for 1991 through 1999 because the radar dataset was not available for years

prior to 2000. FAA developed Official Airline Guide (OAG) schedule-informed inventories modeled with AEDT and great circle trajectories for 1990, 2000, and 2010. Because fuel consumption and CO<sub>2</sub> emission estimates for years 1991 through 1999 are unavailable, consumption estimates for these years were calculated using fuel consumption estimates from the Bureau of Transportation Statistics (DOT 1991 through 2013), adjusted based on 2000 through 2005 data. See Annex 3.3 for more information on the methodology for estimating emissions from commercial aircraft jet fuel consumption.

Data on U.S. Department of Defense (DoD) aviation bunker fuels and total jet fuel consumed by the U.S. military was supplied by the Office of the Under Secretary of Defense (Installations and Environment), DoD. Estimates of the percentage of each Service's total operations that were international operations were developed by DoD. Military aviation bunkers included international operations, operations conducted from naval vessels at sea, and operations conducted from U.S. installations principally over international water in direct support of military operations at sea. Military aviation bunker fuel emissions were estimated using military fuel and operations data synthesized from unpublished data from DoD's Defense Logistics Agency Energy (DLA Energy 2018). Together, the data allow the quantity of fuel used in military international operations to be estimated. Densities for each jet fuel type were obtained from a report from the U.S. Air Force (USAF 1998). Final jet fuel consumption estimates are presented in Table 3-106. See Annex 3.8 for additional discussion of military data.

In order to quantify the civilian international component of bunker fuels, activity data on distillate diesel and residual fuel oil consumption by cargo or passenger carrying marine vessels departing from U.S. ports were collected for individual shipping agents on a monthly basis by the U.S. Customs and Border Protection. This information was then reported in unpublished data collected by the Foreign Trade Division of the U.S. Department of Commerce's Bureau of the Census (DOC 1991 through 2018) for 1990 through 2001, 2007 through 2017, and the Department of Homeland Security's *Bunker Report* for 2003 through 2006 (DHS 2008). Fuel consumption data for 2002 was interpolated due to inconsistencies in reported fuel consumption data. Activity data on distillate diesel consumption by military vessels departing from U.S. ports were provided by DLA Energy (2018). The total amount of fuel provided to naval vessels was reduced by 21 percent to account for fuel used while the vessels were not-underway (i.e., in port). Data on the percentage of steaming hours underway versus not-underway were provided by the U.S. Navy. These fuel consumption estimates are presented in Table 3-107.

**Table 3-106: Aviation Jet Fuel Consumption for International Transport (Million Gallons)**

Nationality	1990	2005	2013	2014	2015	2016	2017
U.S. and Foreign Carriers	3,222	5,983	6,748	7,126	7,383	7,610	8,011
U.S. Military	862	462	294	339	341	333	326
<b>Total</b>	<b>4,084</b>	<b>6,445</b>	<b>7,042</b>	<b>7,465</b>	<b>7,725</b>	<b>7,943</b>	<b>8,338</b>

Note: Totals may not sum due to independent rounding.

**Table 3-107: Marine Fuel Consumption for International Transport (Million Gallons)**

Fuel Type	1990	2005	2013	2014	2015	2016	2017
Residual Fuel Oil	4,781	3,881	2,537	2,466	2,718	3,011	2,975
Distillate Diesel Fuel & Other	617	444	235	261	492	534	568
U.S. Military Naval Fuels	522	471	308	331	326	314	307
<b>Total</b>	<b>5,920</b>	<b>4,796</b>	<b>3,081</b>	<b>3,058</b>	<b>3,536</b>	<b>3,858</b>	<b>3,850</b>

Note: Totals may not sum due to independent rounding.

## Uncertainty and Time-Series Consistency

Emission estimates related to the consumption of international bunker fuels are subject to the same uncertainties as those from domestic aviation and marine mobile combustion emissions; however, additional uncertainties result from the difficulty in collecting accurate fuel consumption activity data for international transport activities separate

from domestic transport activities.<sup>99</sup> For example, smaller aircraft on shorter routes often carry sufficient fuel to complete several flight segments without refueling in order to minimize time spent at the airport gate or take advantage of lower fuel prices at particular airports. This practice, called tankering, when done on international flights, complicates the use of fuel sales data for estimating bunker fuel emissions. Tankering is less common with the type of large, long-range aircraft that make many international flights from the United States, however. Similar practices occur in the marine shipping industry where fuel costs represent a significant portion of overall operating costs and fuel prices vary from port to port, leading to some tankering from ports with low fuel costs.

Uncertainties exist with regard to the total fuel used by military aircraft and ships, and in the activity data on military operations and training that were used to estimate percentages of total fuel use reported as bunker fuel emissions. Total aircraft and ship fuel use estimates were developed from DoD records, which document fuel sold to the Navy and Air Force from the Defense Logistics Agency. These data may slightly over or under estimate actual total fuel use in aircraft and ships because each Service may have procured fuel from, and/or may have sold to, traded with, and/or given fuel to other ships, aircraft, governments, or other entities. There are uncertainties in aircraft operations and training activity data. Estimates for the quantity of fuel actually used in Navy and Air Force flying activities reported as bunker fuel emissions had to be estimated based on a combination of available data and expert judgment. Estimates of marine bunker fuel emissions were based on Navy vessel steaming hour data, which reports fuel used while underway and fuel used while not underway. This approach does not capture some voyages that would be classified as domestic for a commercial vessel. Conversely, emissions from fuel used while not underway preceding an international voyage are reported as domestic rather than international as would be done for a commercial vessel. There is uncertainty associated with ground fuel estimates for 1997 through 2001. Small fuel quantities may have been used in vehicles or equipment other than that which was assumed for each fuel type.

There are also uncertainties in fuel end-uses by fuel type, emissions factors, fuel densities, diesel fuel sulfur content, aircraft and vessel engine characteristics and fuel efficiencies, and the methodology used to back-calculate the data set to 1990 using the original set from 1995. The data were adjusted for trends in fuel use based on a closely correlating, but not matching, data set. All assumptions used to develop the estimate were based on process knowledge, Department and military Service data, and expert judgments. The magnitude of the potential errors related to the various uncertainties has not been calculated, but is believed to be small. The uncertainties associated with future military bunker fuel emission estimates could be reduced through additional data collection.

Although aggregate fuel consumption data have been used to estimate emissions from aviation, the recommended method for estimating emissions of gases other than CO<sub>2</sub> in the *2006 IPCC Guidelines* (IPCC 2006) is to use data by specific aircraft type, number of individual flights and, ideally, movement data to better differentiate between domestic and international aviation and to facilitate estimating the effects of changes in technologies. The IPCC also recommends that cruise altitude emissions be estimated separately using fuel consumption data, while landing and take-off (LTO) cycle data be used to estimate near-ground level emissions of gases other than CO<sub>2</sub>.<sup>100</sup>

There is also concern regarding the reliability of the existing DOC (1991 through 2018) data on marine vessel fuel consumption reported at U.S. customs stations due to the significant degree of inter-annual variation.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2017. Details on the emission trends through time are described in more detail in the Methodology section, above.

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<sup>99</sup> See uncertainty discussions under Carbon Dioxide Emissions from Fossil Fuel Combustion.

<sup>100</sup> U.S. aviation emission estimates for CO, NO<sub>x</sub>, and NMVOCs are reported by EPA's National Emission Inventory (NEI) Air Pollutant Emission Trends website, and reported under the Mobile Combustion section. It should be noted that these estimates are based solely upon LTO cycles and consequently only capture near ground-level emissions, which are more relevant for air quality evaluations. These estimates also include both domestic and international flights. Therefore, estimates reported under the Mobile Combustion section overestimate IPCC-defined domestic CO, NO<sub>x</sub>, and NMVOC emissions by including landing and take-off (LTO) cycles by aircraft on international flights, but underestimate because they do not include emissions from aircraft on domestic flight segments at cruising altitudes. The estimates in Mobile Combustion are also likely to include emissions from ocean-going vessels departing from U.S. ports on international voyages.

## QA/QC and Verification

In order to ensure the quality of the emission estimates from international bunker fuels, General (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and emission factor sources and methodology used for estimating CO<sub>2</sub>, CH<sub>4</sub>, and N<sub>2</sub>O emissions from international bunker fuels in the United States. Emission totals for the different sectors and fuels were compared and trends were investigated. No corrective actions were necessary.

## Planned Improvements

The feasibility of including data from a broader range of domestic and international sources for bunker fuels is being considered.

## 3.11 Wood Biomass and Biofuels Consumption (CRF Source Category 1A)

The combustion of biomass fuels such as wood, charcoal, and wood waste and biomass-based fuels such as ethanol, biogas, and biodiesel generates CO<sub>2</sub> in addition to CH<sub>4</sub> and N<sub>2</sub>O already covered in this chapter. In line with the reporting requirements for inventories submitted under the UNFCCC, CO<sub>2</sub> emissions from biomass combustion have been estimated separately from fossil fuel CO<sub>2</sub> emissions and are not directly included in the energy sector contributions to U.S. totals. In accordance with IPCC methodological guidelines, any such emissions are calculated by accounting for net carbon (C) fluxes from changes in biogenic C reservoirs in wooded or crop lands. For a more complete description of this methodological approach, see the Land Use, Land-Use Change, and Forestry chapter (Chapter 6), which accounts for the contribution of any resulting CO<sub>2</sub> emissions to U.S. totals within the Land Use, Land-Use Change, and Forestry sector's approach.

Therefore, CO<sub>2</sub> emissions from wood biomass and biofuel consumption are not included specifically in summing energy sector totals. However, they are presented here for informational purposes and to provide detail on wood biomass and biofuels consumption.

In 2017, total CO<sub>2</sub> emissions from the burning of woody biomass in the industrial, residential, commercial, and electric power sectors were approximately 221.4 MMT CO<sub>2</sub> Eq. (221,432 kt) (see Table 3-108 and Table 3-109). As the largest consumer of woody biomass, the industrial sector was responsible for 65.3 percent of the CO<sub>2</sub> emissions from this source. The residential sector was the second largest emitter, constituting 20.2 percent of the total, while the commercial and electric power sectors accounted for the remainder.

**Table 3-108: CO<sub>2</sub> Emissions from Wood Consumption by End-Use Sector (MMT CO<sub>2</sub> Eq.)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
Industrial	135.3	136.3	139.8	140.3	138.5	138.3	144.5
Residential	59.8	44.3	58.9	59.7	52.9	46.2	44.6
Commercial	6.8	7.2	7.2	7.9	8.2	8.6	8.6
Electric Power	13.3	19.1	21.4	25.9	25.1	23.1	23.6
<b>Total</b>	<b>215.2</b>	<b>206.9</b>	<b>227.3</b>	<b>233.8</b>	<b>224.7</b>	<b>216.3</b>	<b>221.4</b>

Note: Totals may not sum due to independent rounding.

**Table 3-109: CO<sub>2</sub> Emissions from Wood Consumption by End-Use Sector (kt)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
Industrial	135,348	136,269	139,769	140,331	138,537	138,339	144,502
Residential	59,808	44,340	58,947	59,657	52,872	46,180	44,649
Commercial	6,779	7,218	7,235	7,867	8,176	8,635	8,634

Electric Power	13,252	19,074	21,389	25,908	25,146	23,140	23,647
<b>Total</b>	<b>215,186</b>	<b>206,901</b>	<b>227,340</b>	<b>233,762</b>	<b>224,730</b>	<b>216,293</b>	<b>221,432</b>

Note: Totals may not sum due to independent rounding.

The transportation sector is responsible for most of the fuel ethanol consumption in the United States. Ethanol used for fuel is currently produced primarily from corn grown in the Midwest, but it can be produced from a variety of biomass feedstocks. Most ethanol for transportation use is blended with gasoline to create a 90 percent gasoline, 10 percent by volume ethanol blend known as E-10 or gasohol.

In 2017, the United States transportation sector consumed an estimated 1,135.2 trillion Btu of ethanol (95 percent of total), and as a result, produced approximately 77.7 MMT CO<sub>2</sub> Eq. (77,712 kt) (see Table 3-110 and Table 3-111) of CO<sub>2</sub> emissions. Smaller quantities of ethanol were also used in the industrial and commercial sectors. Ethanol fuel production and consumption has grown significantly since 1990 due to the favorable economics of blending ethanol into gasoline and federal policies that have encouraged use of renewable fuels.

**Table 3-110: CO<sub>2</sub> Emissions from Ethanol Consumption (MMT CO<sub>2</sub> Eq.)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
Transportation <sup>a</sup>	4.1	21.6	70.5	74.0	74.2	76.9	77.7
Industrial	0.1	1.2	3.7	1.6	1.9	1.8	1.8
Commercial	0.1	0.1	0.6	0.4	2.8	2.6	2.6
<b>Total</b>	<b>4.2</b>	<b>22.9</b>	<b>74.7</b>	<b>76.1</b>	<b>78.9</b>	<b>81.2</b>	<b>82.1</b>

<sup>a</sup> See Annex 3.2, Table A-98 for additional information on transportation consumption of these fuels.

Note: Totals may not sum due to independent rounding.

**Table 3-111: CO<sub>2</sub> Emissions from Ethanol Consumption (kt)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
Transportation <sup>a</sup>	4,059	21,633	70,522	74,006	74,187	76,903	77,712
Industrial	105	1,161	3,665	1,647	1,931	1,789	1,801
Commercial	63	149	557	422	2,816	2,558	2,575
<b>Total</b>	<b>4,227</b>	<b>22,943</b>	<b>74,743</b>	<b>76,075</b>	<b>78,934</b>	<b>81,250</b>	<b>82,088</b>

<sup>a</sup> See Annex 3.2, Table A-98 for additional information on transportation consumption of these fuels.

Note: Totals may not sum due to independent rounding.

The transportation sector is assumed to be responsible for all of the biodiesel consumption in the United States (EIA 2019a). Biodiesel is currently produced primarily from soybean oil, but it can be produced from a variety of biomass feedstocks including waste oils, fats and greases. Biodiesel for transportation use appears in low-level blends (less than 5 percent) with diesel fuel, high-level blends (between 6 and 20 percent) with diesel fuel, and 100 percent biodiesel (EIA 2019b).

In 2017, the United States consumed an estimated 253.3 trillion Btu of biodiesel, and as a result, produced approximately 18.7 MMT CO<sub>2</sub> Eq. (18,705 kt) (see Table 3-112 and Table 3-113) of CO<sub>2</sub> emissions. Biodiesel production and consumption has grown significantly since 2001 due to the favorable economics of blending biodiesel into diesel and federal policies that have encouraged use of renewable fuels (EIA 2019b). There was no measured biodiesel consumption prior to 2001 EIA (2019a).

**Table 3-112: CO<sub>2</sub> Emissions from Biodiesel Consumption (MMT CO<sub>2</sub> Eq.)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
Transportation <sup>a</sup>	NO	0.9	13.5	13.3	14.1	19.6	18.7
<b>Total</b>	<b>NO</b>	<b>0.9</b>	<b>13.5</b>	<b>13.3</b>	<b>14.1</b>	<b>19.6</b>	<b>18.7</b>

NO (Not Occurring)

<sup>a</sup> See Annex 3.2, Table A-98 for additional information on transportation consumption of these fuels.

Note: Totals may not sum due to independent rounding.

**Table 3-113: CO<sub>2</sub> Emissions from Biodiesel Consumption (kt)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
Transportation <sup>a</sup>	NO	856	13,462	13,349	14,077	19,648	18,705
<b>Total</b>	<b>NO</b>	<b>856</b>	<b>13,462</b>	<b>13,349</b>	<b>14,077</b>	<b>19,648</b>	<b>18,705</b>

NO (Not Occurring)

<sup>a</sup> See Annex 3.2, Table A-98 for additional information on transportation consumption of these fuels.

Note: Totals may not sum due to independent rounding.

## Methodology

Woody biomass emissions were estimated by applying two gross heat contents from EIA (Lindstrom 2006) to U.S. consumption data (EIA 2019a) (see Table 3-114), provided in energy units for the industrial, residential, commercial, and electric power sectors. One heat content (16.95 MMBtu/MT wood and wood waste) was applied to the industrial sector's consumption, while the other heat content (15.43 MMBtu/MT wood and wood waste) was applied to the consumption data for the other sectors. An EIA emission factor of 0.434 MT C/MT wood (Lindstrom 2006) was then applied to the resulting quantities of woody biomass to obtain CO<sub>2</sub> emission estimates. The woody biomass is assumed to contain black liquor and other wood wastes, have a moisture content of 12 percent, and undergo complete combustion to be converted into CO<sub>2</sub>.

The amount of ethanol allocated across the transportation, industrial, and commercial sectors was based on the sector allocations of ethanol-blended motor gasoline. The sector allocations of ethanol-blended motor gasoline were determined using a bottom-up analysis conducted by EPA, as described in the Methodology section of 3.1 Fossil Fuel Combustion (CRF Source Category 1A). Total U.S. ethanol consumption from EIA (2019a) was allocated to individual sectors using the same sector allocations as ethanol-blended motor gasoline. The emissions from ethanol consumption were calculated by applying an emission factor of 18.7 MMT C/QBtu (EPA 2010) to adjusted ethanol consumption estimates (see Table 3-115). The emissions from biodiesel consumption were calculated by applying an emission factor of 20.1 MMT C/QBtu (EPA 2010) to U.S. biodiesel consumption estimates that were provided in energy units (EIA 2019a) (see Table 3-116).<sup>101</sup>

**Table 3-114: Woody Biomass Consumption by Sector (Trillion Btu)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
Industrial	1,441.9	1,451.7	1,489.0	1,495.0	1,475.9	1,473.8	1,539.4
Residential	580.0	430.0	571.7	578.5	512.7	447.8	433.0
Commercial	65.7	70.0	70.2	76.3	79.3	83.7	83.7
Electric Power	128.5	185.0	207.4	251.3	243.9	224.4	229.3
<b>Total</b>	<b>2,216.2</b>	<b>2,136.7</b>	<b>2,338.3</b>	<b>2,401.1</b>	<b>2,311.8</b>	<b>2,229.8</b>	<b>2,285.5</b>

Note: Totals may not sum due to independent rounding.

**Table 3-115: Ethanol Consumption by Sector (Trillion Btu)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
Transportation	59.3	316.0	1,030.2	1,081.1	1,083.7	1,123.4	1,135.2
Industrial	1.5	17.0	53.5	24.1	28.2	26.1	26.3
Commercial	0.9	2.2	8.1	6.2	41.1	37.4	37.6
<b>Total</b>	<b>61.7</b>	<b>335.1</b>	<b>1,091.8</b>	<b>1,111.3</b>	<b>1,153.1</b>	<b>1,186.9</b>	<b>1,199.1</b>

Note: Totals may not sum due to independent rounding.

<sup>101</sup> CO<sub>2</sub> emissions from biodiesel do not include emissions associated with the C in the fuel that is from the methanol used in the process. Emissions from methanol use and combustion are assumed to be accounted for under Non-Energy Use of Fuels. See Annex 2.3 – Methodology for Estimating Carbon Emitted from Non-Energy Uses of Fossil Fuels.

**Table 3-116: Biodiesel Consumption by Sector (Trillion Btu)**

End-Use Sector	1990	2005	2013	2014	2015	2016	2017
Transportation	NO	11.6	182.3	180.8	190.6	266.1	253.3
<b>Total</b>	<b>NO</b>	<b>11.6</b>	<b>182.3</b>	<b>180.8</b>	<b>190.6</b>	<b>266.1</b>	<b>253.3</b>

NO (Not Occurring)

Note: Totals may not sum due to independent rounding.

## Uncertainty and Time-Series Consistency

It is assumed that the combustion efficiency for woody biomass is 100 percent, which is believed to be an overestimate of the efficiency of wood combustion processes in the United States. Decreasing the combustion efficiency would decrease emission estimates for CO<sub>2</sub>. Additionally, the heat content applied to the consumption of woody biomass in the residential, commercial, and electric power sectors is unlikely to be a completely accurate representation of the heat content for all the different types of woody biomass consumed within these sectors. Emission estimates from ethanol and biodiesel production are more certain than estimates from woody biomass consumption due to better activity data collection methods and uniform combustion techniques.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2017. Details on the emission trends through time are described in more detail in the Methodology section, above.

## Recalculations Discussion

EIA updated wood biomass consumption statistics in the residential sector from 2009 to 2016, and the commercial sector from 2014 to 2016 (EIA 2019a). Ethanol consumption was reallocated across the Transportation, Industrial, and Commercial sectors to match motor gasoline's sectoral distribution used to estimate fossil fuel combustion emissions based on a bottom-up analysis of transportation fuel consumption. Revisions to wood biomass consumption resulted in an average annual increase of 1.2 MMT CO<sub>2</sub> Eq. (0.6 percent) in CO<sub>2</sub> emissions from wood consumption for the period 1990 through 2016, relative to the previous Inventory.

## Planned Improvements

Future research will look into the availability of data on woody biomass heat contents and carbon emission factors the see if there are newer, improved data sources available for these factors.

The availability of facility-level combustion emissions through EPA's GHGRP will be examined to help better characterize the industrial sector's energy consumption in the United States, and further classify woody biomass consumption by business establishments according to industrial economic activity type. Most methodologies used in EPA's GHGRP are consistent with IPCC, though for EPA's GHGRP, facilities collect detailed information specific to their operations according to detailed measurement standards, which may differ with the more aggregated data collected for the Inventory to estimate total, national U.S. emissions. In addition, and unlike the reporting requirements for this chapter under the UNFCCC reporting guidelines, some facility-level fuel combustion emissions reported under EPA's GHGRP may also include industrial process emissions.<sup>102</sup> In line with UNFCCC reporting guidelines, fuel combustion emissions are included in this chapter, while process emissions are included in the Industrial Processes and Product Use chapter of this report. In examining data from EPA's GHGRP that would be useful to improve the emission estimates for the CO<sub>2</sub> from biomass combustion category, particular attention will also be made to ensure time series consistency, as the facility-level reporting data from EPA's GHGRP are not available for all inventory years as reported in this Inventory. Additionally, analyses will focus on aligning reported facility-level fuel types and IPCC fuel types per the national energy statistics, ensuring CO<sub>2</sub> emissions from biomass are separated in the facility-level reported data, and maintaining consistency with national energy statistics provided

<sup>102</sup> See <<https://unfccc.int/resource/docs/2013/cop19/eng/10a03.pdf#page=2>>.



by EIA. In implementing improvements and integration of data from EPA's GHGRP, the latest guidance from the IPCC on the use of facility-level data in national inventories will be relied upon.<sup>103</sup>

Currently emission estimates from biomass and biomass-based fuels included in this Inventory are limited to woody biomass, ethanol, and biodiesel. Other forms of biomass-based fuel consumption include biogas and the biogenic components of MSW. EPA will examine EIA data on biogas to see if it can be included in future inventories. EIA (2019a) natural gas data already deducts biogas used in the natural gas supply, so no adjustments are needed to the natural gas fuel consumption data to account for biogas. Sources of estimates for the biogenic fraction of MSW will be examined, including the GHGRP, EIA data, and EPA MSW characterization data.

Carbon dioxide emissions from biomass used in the electric power sector are calculated using woody biomass consumption data from EIA's *Monthly Energy Review* (EIA 2019a), whereas non-CO<sub>2</sub> biomass emissions from the electric power sector are estimated by applying technology and fuel use data from EPA's Clean Air Market Acid Rain Program dataset (EPA 2018) to fuel consumption data from EIA (2019a). There were significant discrepancies identified between the EIA woody biomass consumption data and the consumption data estimated using EPA's Acid Rain Program dataset (see the Methodology section for CH<sub>4</sub> and N<sub>2</sub>O from Stationary Combustion). EPA will continue to investigate this discrepancy in order to apply a consistent approach to both CO<sub>2</sub> and non-CO<sub>2</sub> emission calculations for woody biomass consumption in the electric power sector.

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<sup>103</sup> See <[http://www.ipcc-nggip.iges.or.jp/public/tb/TFI\\_Technical\\_Bulletin\\_1.pdf](http://www.ipcc-nggip.iges.or.jp/public/tb/TFI_Technical_Bulletin_1.pdf)>.